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# Abiotic hydroxylamine nitrification involving manganese- and iron-bearing minerals



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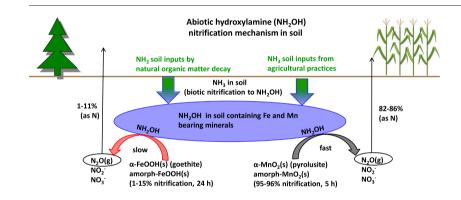
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#### HIGHLIGHTS

#### Abiotic NH<sub>2</sub>OH nitrification by Mn minerals was rapid, but not with Fe minerals.

- The total N mass balance was: input =  $NH_2OH$ ; outputs =  $N_2O(g) + N_2O(aq) + NO_2^- + NO_3^-$ .
- The total N recovery in 4.5 h using pyrolusite and amorph-MnO<sub>2</sub>(s) was 95–96%.
- Total N recovery in 17 d using goethite and amorph-FeOOH(s) was 1.1–14.5%.
- NH<sub>2</sub>OH nitrification by Mn was ≫Fe, despite similar specific [Mn] and [Fe] (mg/m<sup>2</sup>).

#### GRAPHICAL ABSTRACT



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#### ABSTRACT

Hydroxylamine (NH<sub>2</sub>OH) undergoes biotic and abiotic transformation processes in soil, producing nitrous oxide gas  $(N_2O(g))$ . Little is known about the magnitude of the abiotic chemical processes in the global N cycle, and the role of abiotic nitrification is still neglected in most current nitrogen trace gas studies. The abiotic fate of  $NH_2OH$  in soil systems is often focused on transition metals including manganese (Mn) and iron (Fe), and empirical correlations of nitrogen residual species including nitrite (NO<sub>2</sub>), nitrate (NO<sub>3</sub>), and N<sub>2</sub>O(g). In this study, abiotic NH<sub>2</sub>OH nitrification by well-characterized manganese (Mn)- and iron (Fe)-bearing minerals (pyrolusite, amorphous MnO<sub>2</sub>(s), goethite, amorphous FeOOH(s)) was investigated. A nitrogen mass balance analysis involving NH<sub>2</sub>OH, and the abiotic nitrification residuals, N<sub>2</sub>O(g), N<sub>2</sub>O(aq), NO<sub>2</sub>, NO<sub>3</sub>, was used, and specific reactions and mechanisms were investigated. Rapid and complete NH<sub>2</sub>OH nitrification occurred (4–5 h) in the presence of pyrolusite and amorphous MnO<sub>2</sub>(s), achieving a 95–96% mass balance of N byproducts. Conversely, NH<sub>2</sub>OH nitrification was considerably slower by amorphous FeOOH(s) (14.5%) and goethite (1.1%). Direct reactions between the Mn- and Fe-bearing mineral species and NO<sub>2</sub> and NO<sub>3</sub> were not detected. Brunauer-Emmett-Teller surface area and energy dispersive X-ray measurements for elemental composition were used to determine the specific concentrations of Mn and Fe. Despite similar specific concentrations of Mn and Fe in crystalline and amorphous minerals, the rate of NH<sub>2</sub>OH nitrification was much greater in the Mn-bearing minerals. Results underscore the intrinsically faster NH<sub>2</sub>OH nitrification by Mn minerals than Fe minerals.

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#### 1. Introduction

### 1.1. Nitrous oxide byproduct from the abiotic transformation of hydroxylamine

Nitrous oxide ( $N_2O$ ) is a long-lived greenhouse gas present in the atmosphere, is chemically stable, persists in the atmosphere for centuries or longer, and will have a long-term influence on the climate (Anderson et al., 2010).  $N_2O$  is the 4th most important anthropogenic greenhouse gas (Davidson, 2009), and is a major contributor in the destruction of stratospheric ozone (Ravishankara et al., 2009). Upland soil and riparian areas, in conjunction with manmade agricultural activities, account for the majority of the estimated global  $N_2O$  emissions from natural sources (Anderson et al., 2010) with an estimated 50–60% of global  $N_2O$  emissions (US EPA, 2010).

Ammonia (NH<sub>3</sub>) released from the decay of organic matter, the reduction of atmospheric nitrogen  $(N_2)$  to the ammonium ion  $(NH_4^+)$  by certain species of bacteria (i.e., nitrogen fixation), and ammonia from nitrogen-based fertilizer input to croplands are three predominant sources of nitrogen (N) in environmental systems. A large part of the soil ammonium pool will form as hydroxylamine (NH2OH) during nitrification (Arp and Stein, 2003). Once the NH<sub>2</sub>OH is microbiologically produced, it may leak from autotrophic and heterotrophic nitrifiers into the soil matrix (Heil et al., 2015). Abiotic nitrification of NH2OH in soil, including by iron (Fe) and manganese (Mn), forms N<sub>2</sub>O(g) as a major byproduct (Bremner et al., 1980; Bremner, 1997; Heil et al., 2014, 2015, 2016; Liu et al., 2016). The simultaneous occurrence of biotic nitrifier pathways and abiotic nitrification mechanisms in N2O formation introduces challenges in differentiating the role of each. Studies that determine the relative contributions of N2O formation by abiotic NH<sub>2</sub>OH nitrification and nitrifier-denitrification are urgently needed to assess the importance of NH<sub>2</sub>OH oxidation in soil (Snider et al., 2015). Further, little is known about the magnitude of abiotic nitrification processes in the global nitrogen cycle (Heil et al., 2016), and the role of abiotic nitrification is still neglected in most nitrogen trace gas studies (Heil et al., 2015).

Laboratory studies have been successfully used to gain insight into NH<sub>2</sub>OH fate mechanisms and nitrogen residuals in soil systems. The fate of  $NH_2OH$  in soils (n = 19) was investigated where  $NH_2OH$  transformation produced more N<sub>2</sub>O than N<sub>2</sub> (Bremner et al., 1980), N<sub>2</sub>O formation increased over a 5 d period, and was positively correlated with pH, CaCO<sub>3</sub> equivalent, exchangeable Ca<sup>2+</sup>, and oxidized manganese (Mn). The production of N<sub>2</sub>O via chemical decomposition of NH<sub>2</sub>OH in the soils greatly exceeded production of N<sub>2</sub>O(g) through decomposition of nitrite (NO<sub>3</sub>), and no formation of nitrogen oxide (NO) was measured. Results indicated that the abiotic reaction between NH<sub>2</sub>OH and NO<sub>2</sub> was limited. Soil from different ecosystems were amended with NH<sub>2</sub>OH and monitored to assess biotic and abiotic sources of N<sub>2</sub>O (g) formation (Heil et al., 2015). Soil parameters including Fe and Mn were measured for each soil type. In sterilized soil, N<sub>2</sub>O(g) formation was not completely inhibited indicating that abiotic NH<sub>2</sub>OH transformation occurred. It was concluded that the Fe in soil was weakly correlated with NH<sub>2</sub>OH-related N<sub>2</sub>O formation, and a higher correlation between Mn and  $N_2O(g)$  formation, despite lower Mn concentrations relative to Fe. It was proposed that the difference in redox potential of the two redox pairs, Fe2+/Fe3+ and Mn2+/Mn4+, favored the reaction of NH<sub>2</sub>OH and helped explain why lower levels of Mn can exert a higher rate of NH<sub>2</sub>OH oxidation than Fe. It was also proposed that Fe might be complexed too tightly in soil to be available as a reaction partner with NH2OH.

Recent developments in sensitive methods to accurately measure low concentrations of  $NH_2OH$  in soil have enabled a more quantitative determination of  $NH_2OH$  abundance in soils (Liu et al., 2014), and to disentangle the roles of biotic and abiotic fate of  $NH_2OH$  in soil (Liu et al., 2016). Liu et al. (2016) used multiple regression analysis and concluded that Mn was an important factor explaining  $N_2O(g)$  emission rates from

soil, emphasizing the importance of MnO<sub>2</sub> transformation of NH<sub>2</sub>OH to N<sub>2</sub>O(g) in the Norway spruce forest ecosystem. A negative correlation in N<sub>2</sub>O(g) emission rates was reported for soil samples containing organic matter and pH values near or above the pK<sub>a</sub> of NH<sub>2</sub>OH (pK<sub>a</sub> = 5.95). Under these conditions the de-protonated form of NH<sub>2</sub>OH reacted with carbonyl groups to form oximes. Thus, NH<sub>2</sub>OH became less available for the oxidation by MnO<sub>2</sub>.

Studies conducted to better understand the role of abiotic nitrification in the global nitrogen cycle have involved an array of soil types, Fe and Mn content, soil physiochemical characteristics, reagents amended to soil, redox potential, methods of analysis, pH, and organic carbon content (Zhu-Barker et al., 2015; Heil et al., 2016). The range in the concentration of Mn and Fe, and the undifferentiated mineral forms of the Mn and Fe species used in these studies have also had a measurable impact on NH<sub>2</sub>OH transformation, N<sub>2</sub>O(g) production, and the formation of nitrite (NO<sub>2</sub><sup>-</sup>) and nitrate (NO<sub>3</sub><sup>-</sup>) reaction intermediates (Bremner et al., 1980; Zhu-Barker et al., 2015; Heil et al., 2016). Often, acid digested soil samples, used to extract metals from the soil, are analyzed but may not accurately reflect available metals involved in NH<sub>2</sub>OH nitrification. Detailed information regarding geochemical composition, elemental composition, and surface characteristics of the Mn- and Fe-bearing minerals would be beneficial to understand their correlation with NH2OH nitrification.

### 1.2. Reactions and mechanism for Mn- and Fe-mediated abiotic $NH_2OH$ transformation

Given the reports of abiotic transformation of  $NH_2OH$  to  $N_2O$ , and  $NO_2^-$  and  $NO_3^-$  reaction intermediates (Bremner et al., 1980), a series of balanced oxidation-reduction reactions involving Mn and Fe mineral species was proposed as a technical guideline for the study (reactions 1–9). Gibbs free energy calculations were used to assess the thermodynamic feasibility of these reactions. Based on data provided in Diakonov et al. (1994) and Dean (1979), the thermodynamic analysis indicated that all reactions are energetically favorable and a summary of the analysis and calculations are provided in the Supporting information (Section S.1 Summary of Thermodynamic Calculations).

#### 1.2.1. Mn-bearing minerals

There are several crystallographic structures (i.e., polymorphs) of  $\mathsf{MnO}_2.$  Pyrolusite, the most common naturally occurring mineral form of  $\mathsf{MnO}_2,$  was used in this study. Amorphous  $\mathsf{MnO}_2(s),$  is poorly structured and more amenable to dissolution than pyrolusite, and will be referred to as amorph- $\mathsf{MnO}_2(s).$  Contrasting results between the two mineral forms provides insight regarding the potential role of surface characteristics,  $\mathsf{Mn}$  content, and Fe impurities found in the pyrolusite.

In nitrification reactions involving Mn, the *proposed* overall reaction (reaction 1) represents the sum of the *proposed* specific reactions (reactions 2–4) where  $MnO_2(s)$  initiates  $NH_2OH$  transformation and is converted to  $N_2O(g)$ . In specific reactions, intermediates include  $NO_2^-$  and  $NO_3^-$  (reactions 2–4), where  $NH_2OH$  is oxidized to  $NO_2^-$  (reaction 2), and subsequently  $NO_2^-$  is oxidized to  $NO_3^-$  (reaction 3). The final step involves the reduction of  $NO_3^-$  to  $N_2O(g)$  by reduced Mn. During the reactions, =Mn(IV) is reduced to undifferentiated reduced forms of Mn, represented as  $Mn^{2+}$  and MnO. Reduced Mn has been represented as undifferentiated MnO (reaction 5) (Bremner, 1997). The reduction step may be carried out by surface reactions involving =Mn(II), or soluble  $Mn^{2+}$ . In environmental systems exposed to air, =Mn(II) or  $Mn^{2+}$  could become oxidized (=Mn(IV)) by oxidized species in the test system, including  $O_2(g)$  or dissolved oxygen.

#### 1.2.1.1. Overall reaction.

$$\begin{split} 2NH_2OH + MnO_2(s) + MnO + 1/2O_2 + 4H^+ &\rightarrow N_2O + 2Mn^{2+} \\ + 5H_2O \end{split} \tag{1}$$

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