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# A comparison of process performance during the anaerobic mono- and co-digestion of slaughterhouse waste through different operational modes

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#### 43 Introduction

45Anaerobic digestion (AD) technology has gained an increasing 46 attention because of its environmental and economical 47benefits. The worldwide interest for this technology has led 48 to continuous testing and evaluations of different kinds of materials, which are suitable for AD. The first stage in 49evaluations of the feasibility of any substrates to be used for 50AD is the determination of its biomethane potential (BMP). 51This parameter gives valuable information regarding the 52capacity of the substrates to be converted into methane, 53hence, for designing the operational details, as well as for the 54economical evaluation when establishing the new biogas 55

plants (Angelidaki et al., 2009). A simple and reliable method 56 to determine the methane potential of a substrate is a batch 57 anaerobic fermentation assay, so called BMP test. Such 58 method provides useful information regarding the biodegrad-59 ability of the substrates, the possible biogas yield, and makes 60 it possible to estimate the kinetic parameters concerning 61 the degradation rate, as well as the possible toxicity levels 62 (VDI-4630, 2006). Nevertheless, this test provides no in- 63 formation about the stability of the process, which would be 64 important to know for future continuous operation. Further 65 studies, running time-consuming continuous AD assays, are 66 therefore usually required to ensure the long-term effects. 67 Depending on the complexity of the materials treated, the 68

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#### ABSTRACT

The use of consecutive feeding was applied to investigate the response of the microbial 19 biomass to a second addition of substrates in terms of biodegradation using batch tests as a 20 promising alternative to predict the behavior of the process. Anaerobic digestion (AD) of the 21 slaughterhouse waste (SB) and its co-digestion with manure (M), various crops (VC), and 22 municipal solid waste were evaluated. The results were then correlated to previous findings 23 obtained by the authors for similar mixtures in batch and semi-continuous operation 24 modes. AD of the SB failed showing total inhibition after a second feeding. Co-digestion of 25 the SB + M showed a significant improvement for all of the response variables investigated 26 after the second feeding, while co-digestion of the SB + VC resulted in a decline in all of 27 these response variables. Similar patterns were previously detected, during both the batch 28 and the semi-continuous modes. 29

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degradation rate, the interactions between the mixture of 69 substrates and the possible inhibitions of the microbial 70 community; the reactor must be operated using different 71 retention times. It leads to long experimental running (i.e., 72running at the same conditions for a period of at least 3 73 retention times) to evaluate the stability and the performance 74 of the process (Bayr et al., 2012; Zhang et al., 2012). Therefore, 75 decision makers are facing long time delays to get final and 76 77 reliable results applicable for full-scale implementations.

78 Using models describing the AD process is also a valuable and cost-effective tool, which is widely applied to evaluate the 79 system and to predict the biogas production. With the 80 development of the AD models, like ADM1, a joint effort has 81 been dedicated to increase the understanding of this complex 82 biodegradation process (Batstone et al., 2002). However, 83 despite the general acceptance of these models, there are 84 still some emerging areas regarding the characterization of 85 the substrates as well as the resulting interaction of the 86 87 mixtures when the co-digestion process is applied (Batstone, 2013). 88

The use of consecutive feeding (e.g., second feedings) in 89 batch experiments, is an intermediary step between the batch 90 and semi-continuous operation, and it can be a promising 91 92alternative to predict the behavior of the process, i.e., the 93 response of the microbial biomass or the stability and 94 performance of the reactors in a long-term operation, since 95 this test requires a shorter time compared to that needed for 96 running the continuous experiments. Hence, the hypothesis is that consecutive feeding in batch assays would give a fast 97 forecast of the process in order to select the substrate 98 combinations and operational parameters, which would likely 99 be appreciated by both the industrial and research world. 100

The stability of the process is highly dependent on the 101 symbiotic growth and activity of the principal groups of 102 microorganisms involved in the anaerobic degradation chain 103 (Angelidaki et al., 2009). Thus, for monitoring the reactor 104behavior, it is important to determine the specific methano-105 genic activity (SMA). SMA is able to measure the direct enzy-106 matic activity of the microorganisms involved in the last step 107 of the anaerobic degradation, i.e., in the methanogenesis 108 (Sorensen and Ahring, 1993). In fact, evaluating the activity of 109110 the biomass is a key parameter when degrading complex organic matter, reflecting the stability of the reactor during a 111 period of time (Ahring, 1995). 112

In the literature, most of the applications of the SMA test have been focused on measuring the activity of the specific groups of methanogens (i.e., hydrogenotrophic or acetoclastic methanogens) in granular and non-granular sludges by adding specific substrates, such as hydrogen and carbon dioxide, acetate, and methanol (Ahring, 1995; Dolfing and Bloemen, 1985; Sorensen and Ahring, 1993).

120 The objective of this work was to investigate the response of the microbial biomass to a second feeding addition in terms 121of methane yield, degradation kinetics and SMA. The process 122was evaluated under the batch operation mode when mono-123 digesting slaughterhouse wastes or during its co-digestion 124 125with other waste fractions from the agro-industrial activities. To the best of our knowledge, the utilization of consecutive 126 feeding to predict the performance of the co-digestion of dif-127128 ferent agricultural waste streams under practical conditions

has not yet been reported. The methane yield, degradation 129 kinetics, and SMA were determined and then compared with 130 the results obtained previously by the authors in both the 131 batch assays and in the semi-continuous operation modes for 132 similar mixtures (Pagés-Díaz et al., 2014, 2015) in order to 133 establish a correlation between process performance and 134 operation modes. The goal was to examine whether there is a 135 possibility for the use of consecutive feeding in batch assays 136 as a quick prediction method for determining the expected 137 process performance under a semi-continuous operation. 138

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L.	Materials	and	methods	

#### 1.1. Substrates and inoculum

Solid cattle slaughterhouse waste (SB) and its mixtures with 142 animal manure (M), various crops (VC), and the organic 143 fractions of the municipal solid waste (MSW) were prepared. 144 According to the previous results obtained from the batch and 145 semi-continuous operations (Pagés-Díaz et al., 2014, 2015), 146 mixtures of SB + M (50%: 50%), SB + VC + MSW (33%: 33%: 147 33%), and SB + VC (50%: 50%), each based on the wet weight 148 (ww), were investigated. Detailed information of the compo- 149 sition, and the preparation of the materials are presented in 150 Pagés-Díaz et al. (2014). The characteristics of the SB and its 151 different mixture combinations are summarized in Table 1. 152 The inoculum (i.e., biomass) with 3.3% total solid (TS) and 1.9% 153 volatile solid (VS) used in the assays, was obtained from a 154 full-scale thermophilic ( $55 \pm 1^{\circ}$ C) co-digestion biogas plant 155 (Borås Energy and Environment AB, Borås, Sweden) treating 156 MSW. The inoculum was first filtered with a 2-mm sieve to 157 remove the indigested particles and then stored for stabilization 158 at  $(55 \pm 1)^{\circ}$ C during three days, before starting up the assays. 159

#### 1.2. Batch experimental set-up

The experimental set-up was performed in accordance with a 161 method previously described by Hansen et al. (2004). Four 162 set-ups, one for the mono-digestion of the SB and three for the 163 co-digestion assays (i.e., SB + M, SB + VC + MSW, and SB + VC) 164 were performed, all in triplicates. Controls, which included 165 either only the inoculum (blank reactor) to determine the 166 gas production from the inoculum itself, or the cellulose 167 (Cellulose Fibrous Long, Sigma Aldrich, Germany) as a substrate 168 (positive control) to ensure the quality of the inoculum, were 169 also run in parallel. Glass bottles of 2 L were used to carry out 170 the tests. Each bottle was fed with 400 mL of inoculum and 171 fresh substrate to keep the ratio of inoculum to substrate 172 equal to 2 (g VS/g VS). The reactors were then sealed with tick 173 rubber septa and aluminum caps (Apodan Nordic, Copenhagen, 174 Denmark). Finally, the headspace of each reactor was flushed 175 with a mixture of 80% N<sub>2</sub> and 20% CO<sub>2</sub> (V/V) during 3 min to 176 attain anaerobic conditions. The reactors were then incubated 177 at thermophilic conditions (55  $\pm$  1°C) during the experimental 178 period (incubator MMM-group, Einrichtungen GmbH, Venticell) 179 and manually shaken every day. 180

Gas samples were regularly taken from the headspace 181 and analyzed by gas chromatography (GC). To avoid an 182 overpressure inside the bottles, the gas was released after 183

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