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# Alkali resistance promotion of Ce-doped vanadium-titanic-based NH<sub>3</sub>-SCR catalysts

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## ABSTRACT

The effect of K deactivation on V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub> and Ce-doped V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub> catalysts in the selective catalytic reduction (SCR) of NO<sub>x</sub> by NH<sub>3</sub> was studied. Ce-doped V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub> showed significantly higher resistance to K deactivation than V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub>. Ce-doped V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub> with K/V = 4 (molar ratio) showed 90% NO<sub>x</sub> conversion at 350°C, whereas in this case V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub> showed no activity. The fresh and K-poisoned V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub> and Ce-doped V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub> catalysts were investigated by means of *in situ* diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS), NH<sub>3</sub>-temperature progress decomposition (NH<sub>3</sub>-TPD), X-ray photoelectron spectroscopy (XPS) and H<sub>2</sub>-temperature program reduction (H<sub>2</sub>-TPR). The effect of Ce doping on the improving resistance to K of V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub> were discussed.

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## Introduction

Selective catalytic reduction (SCR) of nitrogen oxides (NO<sub>x</sub>) by NH<sub>3</sub> is one of the most successful technologies for the control of NO<sub>x</sub> emission from power plant flue gas and diesel exhaust. The SCR system for coal-fired power plants is commonly in the high-dust (HD) configuration, in which the SCR catalyst is upstream of the electrostatic dust precipitator and processes high dust flue gas with a temperature range of 300–400°C. Vanadium-titanic-based catalysts (e.g., V<sub>2</sub>O<sub>5</sub>-WO<sub>3</sub>/TiO<sub>2</sub>), the most widely used SCR catalysts, can provide high performance NO<sub>x</sub> reduction in the temperature range 300–400°C (Părvulescu et al., 1998; Xu et al., 2017; Zhu et al., 2017). However, a major problem in practical application of V<sub>2</sub>O<sub>5</sub>-WO<sub>3</sub>/TiO<sub>2</sub> catalysts has

been their strong deactivation by deposition of alkali and alkaline earth metals in fly ash. For example, 1% K<sub>2</sub>O-doped V<sub>2</sub>O<sub>5</sub>-WO<sub>3</sub>/TiO<sub>2</sub> gives almost no SCR activity at temperatures from 250 to 400°C (Kamata et al., 1999). The effect of alkali and alkaline earth metals on both tungsten-free and tungsten-containing vanadium-titanic-based catalysts has been well studied (Chen et al., 2010; Due-Hansen et al., 2009; Kamata et al., 1999; Klimczak et al., 2010; Lietti et al., 1993; Lisi et al., 2004; Nicosia et al., 2007, 2008; Tang et al., 2010; Wu et al., 2013). The Brønsted acidity of catalysts and the amount of NH<sub>3</sub> adsorbed on the surface can be decreased by alkali metals like K, which cause the decrease of SCR activity (Kamata et al., 1999; Lietti et al., 1993). Nicosia et al. (2008) explained the deactivation of V<sub>2</sub>O<sub>5</sub>/WO<sub>3</sub>-TiO<sub>2</sub> catalysts by alkali and alkali earth metals by a

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mechanism whereby the poisoning element (Ca, K) occupies the non-atomic hole sites of the (010)  $V_2O_5$  surface, such that both Brønsted acid and  $V^{5+} = O$  sites are blocked.

Besides the effect on surface acidity, the reducibility of  $V_2O_5/TiO_2$  catalysts could be another reason for the poisoning effect of  $Na^+$  and  $Ca^{2+}$  (Tang et al., 2010). Yu et al. (2013) found that the isolated vanadia species over the  $V_2O_5-WO_3/TiO_2$  catalyst were more reactive with potassium than polymeric vanadia species. Chen et al. (2010) proposed that decreases in  $NH_3$  adsorption, surface chemisorbed oxygen and reducibility of surface vanadium species could be the main factors in the poisoning effect of alkali metal.

In our previous study, cerium-titanium ( $Ce/TiO_2$ ,  $CeTiO_x$ ,  $CeWTiO_x$ ) catalysts were developed and showed high activity for  $NH_3$ -SCR (Shan et al., 2011, 2012; Xu et al., 2008). Chen et al. (2009) and Peng et al. (2012) reported that substitution part of W by Ce in  $V_2O_5/WO_3-TiO_2$  (0.1 or 0.4 wt.% of  $V_2O_5$ ) can promote SCR activity in the temperature range 200–450°C and enhance the catalyst's alkali poisoning resistance. Xu et al. (2015) studied the addition of Ce and Sb to the  $V_2O_5/TiO_2$  (5 wt.% of V) can improve the low temperature activity for  $NH_3$ -SCR of  $NO_x$ . Here, the promotional effect of Ce-doping on the alkali metal resistance of the  $V_2O_5/WO_3-TiO_2$  catalyst was investigated. The effect of K-poisoning on  $V_2O_5/WO_3-TiO_2$  and Ce-doped  $V_2O_5/WO_3-TiO_2$  were studied by means of activity measurements, X-ray photoelectron spectroscopy (XPS),  $NH_3$ -temperature progress decomposition ( $NH_3$ -TPD), *in situ* diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS), and  $H_2$ -temperature program reduction ( $H_2$ -TPR).

## 1. Experimental

### 1.1. Catalyst preparation

The catalysts were prepared by a wet impregnation method.  $TiO_2$  powder containing 5 wt.%  $WO_3$  was used in this work. Ammonium metavanadate was dissolved in an oxalic acid solution. The  $TiO_2$  powder was impregnated in the mixed solution by stirring for 1 hr. After rotary evaporation, the sample was dried at 110°C overnight and then calcined at 500°C in air for 3 hr.  $V_2O_5/WO_3-TiO_2$  catalysts with 1 wt.%  $V_2O_5$  were prepared and denoted as VWTi. Ce-doped  $V_2O_5/WO_3-TiO_2$  catalysts with 6 wt.% Ce were prepared by adding a cerium nitrate solution into the solution of ammonium metavanadate and oxalic acid, and followed by a process similar to the preparation of  $V_2O_5/WO_3-TiO_2$  catalysts. Ce-doped  $V_2O_5/WO_3-TiO_2$  catalysts were denoted as Ce-VWTi. The K-containing samples were prepared by impregnating the dry powder with  $KNO_3$  solution for 8 hr, then dried in air at 100°C overnight and calcined at 500°C for 3 hr. The Ca and Mg poisoning samples were prepared by the same procedures according to the molar ratio of  $Ca/V = Mg/V = 4$ . All catalysts were ground and sieved to 40–60 mesh for activity testing.

### 1.2. Catalytic activity measurements

The reaction conditions were controlled as follows: 500 ppm  $NO$ , 500 ppm  $NH_3$ , 5 vol.%  $O_2$ ,  $N_2$  balance; 150 mg catalyst, total flow rate of 500 mL/min and gas hourly space velocity (GHSV) =

100,000  $hr^{-1}$ . The effluent gas was analyzed using an FTIR spectrometer (Nexus 670, Nicolet, USA) equipped with a heated, low volume multiple-path gas cell (2 m), which can continuously analyze the  $NO$ ,  $NO_2$ ,  $N_2O$  and  $NH_3$  in the effluent gas. The spectra were collected when the SCR reaction reached a steady state.  $NO_x$  conversion was calculated as follows (Liu et al., 2009):

$$NO_x \text{ conversion} = \left(1 - \frac{[NO]_{out} + [NO_2]_{out}}{[NO]_{in}}\right) \times 100\% \quad (1)$$

### 1.3. Catalyst characterization

The *in situ* DRIFTS experiments were performed on an Fourier Transform infrared spectroscopy (FTIR) spectrometer (Nexus 670, Nicolet, USA) equipped with a mercury cadmium telluride (MCT) detector cooled by liquid nitrogen and an *in situ* DRIFTS reactor cell with ZnSe window connected to a purging/adsorption gas control system, collecting 100 scans at spectral resolution of  $4 \text{ cm}^{-1}$ . The reaction temperature was controlled precisely by an Omega programmable temperature controller. Each sample was pretreated at 350°C in a flow of 20 vol.%  $O_2/N_2$  for 1 hr, and cooled down to 50°C in  $N_2$ . The sample was saturated at 50°C with 0.5%  $NH_3$  in  $N_2$  for about 30 min, and purging with  $N_2$  at 50°C for 1 hr, then heating up to 150°C. The spectra were recorded at 150°C.

Ammonia temperature-programmed desorption ( $NH_3$ -TPD) experiments were used to determine the  $NH_3$  storage capacity of SCR catalysts. The test sample (100 mg) was pretreated at 400°C for 30 min by 20 vol.%  $O_2/N_2$ , and then cooled down to 150°C.  $NH_3$  was introduced until the adsorption on the sample was saturated. The sample was then sufficiently purged with  $N_2$  to remove excess adsorbate from the surface of the sample. The TPD was conducted by heating the sample in  $N_2$  from 150 to 500°C at a rate of  $10^\circ\text{C}/\text{min}$  and the  $NH_3$  in the outlet gas was analyzed by an online NEXUS 670-FTIR (Nexus 670, Nicolet, USA) spectrometer.

The catalysts were analyzed using X-ray photoelectron spectroscopy (XPS) to identify the surface nature. The XPS data were taken on using Al K $\alpha$  radiation (AXIS Ultra, Kratos Japan). The binding energy was corrected using the energy of adventitious carbon (284.7 eV).

$H_2$ -TPR experiments were performed in a flow of 10%  $H_2/Ar$  mixture ( $50 \text{ cm}^3/\text{min}$ ) over 50 mg of catalyst using a heating rate of  $10^\circ\text{C}/\text{min}$ . The consumption of  $H_2$  was detected by thermal conductivity detector (TCD) (AutoChem II 2920, Micromeritics, USA).

## 2. Results and discussion

### 2.1. SCR activity evaluation

Commonly, the operating temperature window of vanadium based catalysts for reducing  $NO_x$  from stationary sources is between 300 and 400°C. To investigate the effect of K doping on the VWTi and Ce-VWTi catalysts, the  $NH_3$ -SCR activity of fresh and K-doped catalysts was tested. Fig. 1a–b shows the  $NH_3$ -SCR activity and of VWTi and Ce-VWTi catalysts with different K loadings at the temperature range from 275 to 450°C under a fixed GHSV of 100,000  $hr^{-1}$ . As can be seen in

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