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² An Alternative Rangeland Management Strategy in an Agro-Pastoral Area ³ in Western China^(L)</sup>

Q2 Hua Limin ^{a,*}, Yang Siwei ^b, Victor Squires ^c, Wang Guizhen ^d

^a Associate Professor, Key Laboratory of Grassland Ecosystem of Ministry of Education/Rangeland Sciences Department, Gansu Agricultural University/Sino-U.S. Centers for Grazing Land Ecosystem
 Sustainability, Lanzhou, 730070, China, Institute of Animal and Veterinary Science, Bijie, 551700, China

^b PhD Student, Key Laboratory of Grassland Ecosystem of Ministry of Education/Rangeland Sciences Department, Gansu Agricultural University/Sino-U.S. Centers for Grazing Land Ecosystem
 Sustainability, Lanzhou, 730070, China, Institute of Animal and Veterinary Science, Bijie, 551700, China

9 ^c Visiting Professor, Key Laboratory of Grassland Ecosystem of Ministry of Education/Rangeland Sciences Department, Gansu Agricultural University/Sino-U.S. Centers for Grazing Land Ecosystem 10 Sustainability, Lanzhou, 730070, China, Institute of Animal and Veterinary Science, Bijie, 551700, China

 ^d Master Student, Key Laboratory of Grassland Ecosystem of Ministry of Education/Rangeland Sciences Department, Gansu Agricultural University/Sino-U.S. Centers for Grazing Land Ecosystem Sustainability, Lanzhou, 730070, China, Institute of Animal and Veterinary Science, Bijie, 551700, China

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ABSTRACT

Governance plays a key role in rangeland management. In China, all rangeland, including pastoral land and agro- 23 pastoral land, is owned by the State. Since 1980, use rights have been granted to households by the Chinese 24 government extending the household contract responsibility system (HCRS). But in the agro-pastoral areas of 25 northwestern (NW) China, the rangeland degradation is more severe than that in pastoral areas. The HCRS is 26 difficult to implement because the limited and fragmented grazing land cannot be contracted to individual 27 households. Thus the pastures in the agro-pastoral areas are grazed as communal pastoral land and the rate of 28 rangeland degradation has accelerated as livestock numbers have grown. Several measures have been introduced 29 in an attempt to reverse this degradation trend, but most failed. This paper reports a 5-year comparison of three 30 rangeland management regimens, including the national "Protecting rangeland by restricting grazing" (PRRG) 31 project under the individual HCRS (PRRG under IHCRS), the Allied Householders Contract Responsibility System 32 (AHCRS) program funded by the World Bank/GEF, and the free grazing on common pasture as the control area 33 (CA) at Mayinggou Village, Yongchang County, Gansu Province in NW China. The results showed significant 34 differences (P < 0.05) between AHCRS and the other two regimens (PRRG under IHCRS and CA) in terms of 35 biomass of palatable forages, cover, and plant diversity index of vegetation but no significant difference (P > 360.05) between PRRG under IHCRS and CA. Reducing the number of livestock in AHCRS also resulted in increased 37 revenue from the livestock turn-off rate compared with that in PRRG under IHCRS and CA. Therefore, AHCRS is a 38 better alternative management regimen for rangelands in agro-pastoral areas. AHCRS can solve the overgrazing 39 problem, maintain or improve household income, and potentially ensure a long-term sustainable rangeland 40 management regimen in agro-pastoral areas in NW China. 41

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43 Introduction

Q4 China has the third largest area of rangelands in the world (Squires 2010). Rangeland can be categorized as either pastoral land, where cropping does not occur, or agro-pastoral lands, where cropland and grazing land are in close proximity (Cheng, 1999). The agro-pastoral land is the transition region of the traditional agricultural region and pastoral region (Hess, 1990; Xu, 1999). In China, the agro-pastoral lands encompass the southeastern edge of Mongolia Plateau, south 50 Loess Plateau, and part of Qinghai-Tibet Plateau (Chen et al., 2004; 51 Shao et al., 2006). The transition zone covers 12 provinces and 140 52 counties with an area of 44 million hectare (Mha) and population of 53 35 million in 2003 (Sun and Shi, 2003). In this region, the rainfall is in 54 the range of 250 to 500 mm, both plant cover and productivity are 55 low, and the ecosystem is susceptible to degradation (Cheng, 2002; 56 Pan et al., 2003). As a result, the farming system is sensitive to environ- 57 mental change and human influence. The degraded and decertified area 58 of rangeland in the agro-pastoral area represents about 50% of the total 59 degraded rangeland area in northwestern China (Cheng, 1999). The 60 rangeland area in agro-pastoral land in Inner Mongolia shrank by 0.41 61 Mha over a 10-year period (1990-2000) (Zhan et al., 2004). Although 62 many areas were illegally converted to croplands for producing more 63 grain crops, the problem of accelerated rangeland degradation in agro- 64

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^{*} Correspondence: Hua Limin, No.1, Yingmen Village, Anning District, Lanzhou, 730070, China.

E-mail address: Hualm@gsau.edu.cn (H. Limin).

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pastoral land has been exacerbated by the imposition of new institutional arrangements relating to land tenure and use rights (Williams,
1996; DaLintai and Gaowa, 2010).

Governance is the key to promotion of sustainable use and manage-68 ment of natural resources and to achieve the goals of conservation of 69 natural resources and environment, alleviation of poverty, and sustain-70able utilization of natural resources (Ostrom, 1990; Acheson, 2006). 71 With the implementation of economic policy opening up and reform 72and cropland tenure reform in the better-watered agricultural areas, 06 the household contract responsibility system (HCRS) (jia ting cheng 74 75bao ze ren zhi) was extended to the pastoral and agro-pastoral areas to 76solve the problem of unfettered common-use grazing (Feng, 1988; Cao and Wang, 1995; Yang 2008). HCRS is a policy designed to boost ag-08 07 ricultural productivity and was first adopted in China in 1981. A key part 78 of HCRS was to assign use rights to individual households from land that 7980 was formerly communally used (Brandt et al., 2002). HCRS has been the fundamental tool for rangeland management in China since the 1980s 81 and raised productivity and increased animal husbandry output from 82 83 the 1980s to 1990s (Wang et al., 2010). According to the Grassland¹ Law of the People's Republic of China (2002 Amendment), the use rights 84 85 to rangeland belonging to the State could be contracted by individual 86 householders (dan hu cheng bao) or cooperating householders (lian hu cheng bao) (Banks, 2003; Cao et al., 2011; Li et al. 2011). The individual 09 88 HCRS (IHCRS) (dan hu cheng bao ze ren zhi) implies that a household has use rights to a specific area of pasture and can run privately owned live-89 90 stock there. By contrast, under cooperating HCRS, each household has private ownership of livestock, but their pasture is shared and managed 91by the cooperating households. From 2001-2011, the contracted range-92land area by individual households and cooperating households in 93 China increased to 273 Mha, which accounts for 83% of total rangeland 94 available nationwide. The proportion of rangeland contracted to indi-95 vidual households is now about 80.5% of the total contracted rangeland 96 97 area in China (China Grass Internet, 2013). The IHCRS has been the dominant management type in the rangeland region of China. However, 98 99 the policy of IHCRS is difficult to apply in agro-pastoral areas because of the contradiction between population pressure and the relatively small 100 areas of residual rangeland. In agro-pastoral regions in China, the aver-101 age rangeland area was 0.23 ha per household, which is too small to be 102103 operated as functional rangeland by individual households (Xu et al., 104 2012). As a result, the rangeland in many parts of the agro-pastoral regions was used as communal grazing land and the overstocking rate 105 was higher (42.07%) than that (23.37%) in the pastoral regions (Xu 106 107 et al., 2012). IHCRS did not solve the dilemma of private livestock on 108 state-owned land in the agro-pastoral regions. Severe overgrazing has 109become a major challenge because of the pressure to maintain or improve livelihoods that was believed by households to come more 110quickly from bigger herds (Richard et al., 2006). 111

112 In order to find a better management approach for rangeland in the agro-pastoral region, a potentially viable alternative was trialed in 113114 Mayinggou Village, Yongchang County, Gansu Province in NW China 115in 2007 when the World Bank/GEF supported a rangeland conservation project there. Each community of the village has use rights to its own 116 winter pasture (closer to the village) and summer pasture (often far 117 118 from the village). However, the pasture was allocated by the village 119 committee to households, more or less at random. Each allocation has a use right certificate, but there was no clear boundary between these 120individually assigned pastures, so trespass grazing was a problem. 121 Entry and exit dates were either not well defined or ignored, and the 122123pastures received no rest. Therefore, the key issue of rangeland degradation in the agro-pastoral village was the failure of the existing rangeland 124management regimen. There was an opportunity to devise a new range-125land management arrangement. A participatory rural appraisal was 126127conducted (Hua and Zhang, 2012), revealing that most households

considered the uncontrolled grazing on communal rangeland was the128biggest contributor to rangeland degradation in this village. After several129discussions between the project officers and farmers, an alternative130Allied Households Contract Responsibility System (CRS) (*lian hu ze ren*131cheng bao zhi) was implemented in the village in accordance with the132households' approval (World Bank, 2004).133

Since 2000, the Protecting Rangeland by Restricting Grazing (PRRG) 134 (*tui mu huan cao*) project funded by the Ministry of Agriculture (MOA) 135 has been conducted in China (Huang and Wang, 2004; Li, 2011). The Q10 PRRG project was predicated on the prior registration of IHCRS for all 137 households in the project area and on demarcation of individual pasture 138 allocations (Liu, 2010). For the funding support from the central 139 government, the local government of Yongchang Country extended 140 the IHCRS to other communities of Mayinggou Village and the pasture 141 belonging to the participating households from the village was divided 142 into each household's portion on the basis of the number of people in 143 the household. 144

In this study, the rangeland recovery and changes in stocking pressure within the three management regimens were used to assess the impact on rangeland recovery and family income of reassigning grazing use rights to groups of cooperating households in the agro-pastoral rangeland in NW China. There was opportunity to compare outcomes under three contrasting management regimens: 1) the nationally sponsored PRRG under IHCRS, 2) communally grazed rangeland without institutional management, and 3) a new regimen, Allied Households Contract Responsibility System (AHCRS), developed in western Gansu.

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Study Site	and	Methods	
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Study Site

Mayinggou Village is located in the middle section of Qi Lian Moun- 156 tain in the Hexi Corridor of western Gansu Province (Fig. 1). The village 157 has 518 households (HH) with 2 155 people of Han nationality. Each 158 household has up to six people. The available areas of rangeland and 159 cropland are 8 677 ha and 260 ha, respectively. This accounts for 160 30.6% and 2.9% of the total land area, respectively. The other land type 161 is primarily sparse woodland with low shrub and plantation forestry 162 for woodlots and shelter belts. The altitude is 2 178 to 2 515 m. The cli-163 mate is temperate with a semi-alpine zone at higher altitude. Annual 164 rainfall ranges from 170 mm to 320 mm and temperature from 1.5°C 165 to 7.0°C (Yang, 2010). The rainfall is mainly distributed in July, August, 166 and September. The rangeland types include alpine meadow, temperate 167 grassland, semi-arid rangeland, and arid rangeland. The livestock num- 168 ber in 2007 was 3 240 Sheep Unit.² According to the record of the village 169 committee, sheep are the dominant livestock in the village (Yang, 170 2010). The main crops are barley, beans, and wheat. The dominant 171 plants on the rangeland are Leymus secalinum, Agropyron cristatum, 172 Poa pratensis, Stipa sareptarta var. krylovii, Deyeuxia scabrescens, Stipa 173 purpurea, Stipa breviflora, and Artemisia frigida. 174

Various rangeland regimens existed in past decades in the village. Before the 1980s, all rangeland and livestock was State owned, but the village had the right to use the pastoral lands and manage the livestock 177 enterprise. From 1980–2007, the State-owned livestock was redistributed 178 to individual householders, and the rangeland was owned by the village 179 as common pasture. Since 2007, with the extension of HCRS in pastoral 180 land, partial rangeland in four communities has been redistributed 181 to individual householders. In 2007, the World Bank/GEF funded a 182 rangeland conservation project to solve the rangeland degradation 183 based on a reformed rangeland regimen in two demonstration communities of this village. 185

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¹ Grassland in China involves rangeland and sown grassland.

 $^{^2~}$ Sheep Unit:a 50 kg sheep with a half-year lamb at foot eating 1.8 kg per day dry forage with 14% of moisture.

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