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#### **Short Communication**

# Confirmation of elevated arsenic levels in groundwater of Myanmar



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#### HIGHLIGHTS

- Of 55 wells sampled in Myanmar, the groundwater of only 24 contained <10 µg/L As.
- Reductive dissolution of Fe oxyhydroxides is inferred as the mechanism of As release.
- Observations in the Ayeyarwady are consistent with predictions and field-kit data.
- Well testing is urgently needed in order to help reduce the human exposure.

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#### ABSTRACT

Millions of villagers across South and Southeast Asia are exposed to toxic levels of arsenic (As) by drinking well water. In order to confirm the field-kit results that Myanmar is also affected, a total of 55 wells were tested in the field in January 2013 and sampled for laboratory analysis across seven villages spanning a range of As contamination in the lower Ayeyarwady basin. Elevated concentrations of As (50–630 µg/L) were measured in wells up to 60 m deep and associated with high levels of Fe (up to 21 mg/L) and low concentrations of SO<sub>4</sub> (<0.05 mg/L). Concentrations of As <10 µg/L were measured in some shallow (<30 m) grey sands and in both shallow and deep orange sands. These results indicate that the main mechanism of As release to groundwater in Myanmar is the reductive dissolution of Fe oxyhydroxides, as in the neighboring Bengal, Mekong, and Red River basins. Concentrations of As in groundwater of Myanmar are therefore unlikely to change rapidly over time and switching to existing low-As wells is a viable way of reducing exposure in the short term. However, only 17 of the 55 well owners interviewed correctly recalled the status of their well despite extensive testing in the region. A renewed effort is thus needed to test existing wells and new wells that continue to be installed and to communicate the health risks of exposure to As for infants, children, and adults.

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#### 1. Background

The most dramatic impacts of drinking well water elevated in arsenic (As) such as cancerous skin lesions and loss of limb were recognized in India in the mid-1980s (Chakraborty and Saha, 1987). Public health scientists have since shown that As exposure also increases infant mortality (Rahman et al., 2010), reduces intellectual function in children (Wasserman et al., 2004), and increases adult mortality (Argos et al., 2010) due to cardiovascular disease (Chen et al., 2011) and cancers of the lung, liver, and bladder (Smith et al., 2000). It is now estimated that over 100 million in Pakistan, Nepal, India, Bangladesh, Cambodia, Vietnam, and China are chronically exposed to

As by drinking groundwater that does not meet the World Health Organization guideline of 10  $\mu$ g/L (Ravenscroft et al., 2009).

Elevated concentrations of As in groundwater have been predicted for the lower Ayeyarwady basin on the basis of geologic and climatic factors (Fig. 1a; Winkel et al., 2008) and confirmed by field-kit testing of over 200,000 wells and a smaller number of laboratory measurements (Tun et al., 2003; WRUD/UNICEF, 2006; MOH/UNICEF, 2013) by the Ministry of Health and the Water Resources Utilization Department with support from UNICEF over the past decade. There is a broad consensus that the reductive dissolution of iron (Fe) oxyhydroxides is a key factor leading to naturally elevated As concentrations in anoxic groundwater over large expanses of South and Southeast Asia (Ravenscroft et al., 2009; Fendorf et al., 2010). In some semi-arid areas such as the Indus River basin, however, other mechanisms operating under oxic conditions instead have been shown to result in the accumulation of As in groundwater (Farooqi et al., 2007). Central Myanmar is

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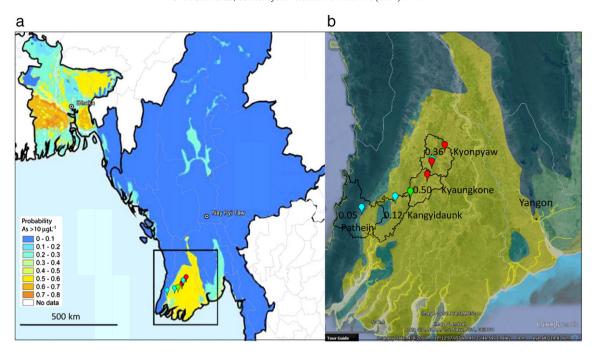


Fig. 1. (a) Map from Winkel et al. (2008) showing the probability of encountering groundwater with  $>10 \, \mu g/L$  As groundwater in Myanmar and Bangladesh and Myanmar. (b) Close-up of the lower Ayeyarwady basin combined with satellite imagery from Google Earth. Also shown are the boundaries of four townships labeled with their names and, according to MOH/WRUD field-kit data, the proportion of wells with  $>10 \, \mu g/L$ , starting from the north with Kyonpyaw (19,301 wells tested), Kyaungkone (6,996), Kangyidaunk (3,696), and Pathein (3,342). Symbol color indicates the average As content of wells  $<60 \, \text{m}$  in each of the 7 villages sampled: light blue  $<10 \, \mu g/L$ ; green  $10-50 \, \mu g/L$ ; red  $>50 \, \mu g/L$ .

also a semi-arid region, although it is located well upstream of the focus of the present study in the lower Ayeyarwady basin.

#### 2. Methods

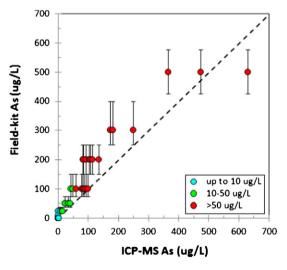
Test results mapped by the Water Resources Utilization Department of the Ministry of Agriculture and Irrigation were made available for the planning of a short sampling campaign to verify the results and identify the mechanisms underlying the release of As to groundwater. A total of 55 wells were sampled on 23-27 January 2013 in 7 villages distributed along a 100-km transect that extends across four townships of the Ayeyarwady region, According to previous field-kit testing, the proportion of wells in those four townships containing > 10 µg/L As ranges from 0.05 to 0.50 (Fig. 1b). Villages are referred to hereon according to their distance in kilometers along the transect, starting from the north. In the field, As concentrations were measured visually with the ITS Econo-Quick arsenic kit and the Wagtech portable kit using a digital reader (George et al., 2012). In the laboratory, acidified groundwater was analyzed by high-resolution inductively coupled plasma mass spectrometry on a Thermo Element2 for As, Fe, and other groundwater constituents (Cheng et al., 2004). The reproducibility was better than  $\pm 3\%$ , the detection limit < 0.1 µg/L, and the accuracy was verified against reference standards NIST1640A and NIST1643. Unacidified samples were analyzed for sulfate (SO<sub>4</sub>) and other anions using a Dionex DX2000 ion chromatograph in gradient mode equipped with an AS-11HC column (see Supplemental Table for these and additional measurements). The reproducibility for SO<sub>4</sub> was better than  $\pm 2\%$  and the detection limit < 0.05 mg/L. Before sampling, the owner of a well or a household member was informed of the purpose of the testing and asked about the status of the well with respect to As.

### 3. Results

Visual field-kit measurements identified a total of 23 wells that met the WHO guideline for As of 10  $\mu g/L$ , another 9 wells with As concentrations up to the national standard of 50  $\mu g/L$ , and 23 wells with > 50  $\mu g/L$  As. As documented in previous studies, the ITS kit results were generally

consistent with laboratory measurements (Fig. 2) as well as field data using the considerably more costly and elaborate Wagtech procedure (George et al., 2012; van Geen et al., 2014). Relative to laboratory measurements, the ITS kit shifted 3 samples with <10  $\mu$ g/L up to the next less safe category and another 2 samples containing 10–50  $\mu$ g/L up to the unsafe category. The ITS kit did not underestimate the As content of any of the 55 samples relative to these categories.

In samples from the three most affected villages located at km 00, 21, and 32, in Kyonpyaw and Kyaungkone townships, laboratory measurements of As in wells <50 m deep in each village average 80–380  $\mu$ g/L (23 wells sampled altogether). None of these samples contain >1 mg/L SO<sub>4</sub> whereas most contain >5 mg/L Fe (Fig. 3). At the other end of the spectrum, As concentrations samples collected from wells <50 m deep in the least affected villages in Kyonpyaw and Pathein townships at



**Fig. 2.** Comparison of As concentrations in water samples measured by ICP-MS and the field ITS Econo-Quick field kit. Vertical error bars indicate the range of As concentrations for a particular measurement, calculated by assuming a mid-range cutoff between adjacent readings.

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