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# Evidence of an increasing $NO_2/NO_X$ emissions ratio from road traffic emissions

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#### Abstract

A statistical analysis of roadside concentrations of nitrogen oxides (NO<sub>X</sub>) and nitrogen dioxide (NO<sub>2</sub>) in London shows that from 1997 to 2003 there has been a statistically significant downward trend (at the p = 0.004 level) in NO<sub>X</sub> averaged across a network of 36 sites. Conversely, there has been no statistically significant trend in the concentrations of NO<sub>2</sub> over the same period. Hourly modelling using a simple constrained chemical model shows that the NO<sub>2</sub>/NO<sub>X</sub> emissions ratio from road traffic has increased markedly from a mean of about 5–6 vol% in 1997 to about 17 vol% in 2003. Calculations show that if the NO<sub>2</sub>/NO<sub>X</sub> emissions ratio had remained the same as that towards the beginning of each time series, 14 out of the 36 sites would have shown a statistically significant downward trend in NO<sub>2</sub> at the p = 0.10 level compared with only five that did. The increase in the NO<sub>2</sub>/NO<sub>X</sub> emissions ratio from road traffic in recent years has therefore had a significant effect on recent trends in roadside NO<sub>2</sub> concentrations. It is shown that the increased use of certain types of diesel particulate filters fitted to buses is likely to have made an important contribution to the increasing trends in the NO<sub>2</sub>/NO<sub>X</sub> emissions ratio. However, it is unlikely that these filters account for all of the observed increase and other effects could be important, such as the increased penetration of diesel cars in the passenger car fleet and new light- and heavy-duty engine technologies and management approaches.

Keywords: Primary NO2; Nitrogen dioxide; Diesel particulate filter; Trend analysis; Mann-Kendall; London; Urban air quality

### 1. Introduction

### 1.1. Background

Nitrogen dioxide (NO<sub>2</sub>) represents an important urban pollutant in most European cities. Concerns over the health impacts of NO<sub>2</sub> have prompted legislation at a national and international level. In Europe, the EU First Daughter Directive (99/30/EC) sets an annual mean limit of  $40 \, \mu \mathrm{g \, m^{-3}}$ , and an hourly limit of  $200 \, \mu \mathrm{g \, m^{-3}}$  not to be exceeded on more than 18

\*Tel.: +44 113 3437522; fax: +44 113 3434127. E-mail address: d.c.carslaw@its.leeds.ac.uk. occasions each year. The Directive requires compliance by 1 January 2010. In the UK, recent projections show that many urban areas will not meet the annual mean limit by 2010 (AQEG, 2004). In the UK, there is also a system of Local Air Quality Management (LAQM) which recognises that national or international measures will not always be the best way to deal with localised air quality problems. The LAQM process has led to 132 (out of 407) local authorities declaring air quality management areas, for which further effort will be required to meet air quality limits (AQEG, 2004). Of the 132 authorities, 92% were for NO<sub>2</sub>, 45% for PM<sub>10</sub> and 4% for sulphur dioxide. The declarations for NO<sub>2</sub> were all related to road traffic sources; although in some cases

industrial sources were considered important. LAQM in the UK shows NO<sub>2</sub> to be an important pollutant and road traffic to be the most important source affecting concentrations.

### 1.2. Recent trends in roadside $NO_X$ and $NO_2$ concentrations in London

Trends in the concentration of  $NO_X$  and  $NO_2$  have been calculated for 36 air pollution monitoring sites in London. Thirty-two of these sites are classified as kerbside (<1 m from the kerb) or roadside (<5 m from the kerb); four are defined as urban background sites, but are strongly influenced by local road sources. Trends have been considered from January 1997 to December 2003. Fig. 1 shows the monthly mean trend in  $NO_X$  and NO<sub>2</sub> for all sites. Fig. 1 shows that concentrations of NO<sub>X</sub> show a strong seasonal component, with concentrations being consistently higher during the winter months. There is no clear seasonal component to monthly mean NO<sub>2</sub> concentrations. Also apparent from Fig. 1 is that concentrations of  $NO_X$  have declined from 1997 to 2003. For NO<sub>2</sub> there is no clear trend in concentration. In order to quantify whether the trends in  $NO_X$  and  $NO_2$  are statistically significant, a seasonal Mann-Kendall test has been applied (Hirsch et al., 1982). The test, which has proved popular for water quality studies, is a non-parametric test that can be applied to data that are not normally distributed and where there are missing data. Applying the test to the data in Fig. 1 shows that there is a statistically significant downward trend in NO<sub>X</sub> at the p = 0.004level. For NO2, no statistically significant trend was detected (p = 0.63). The Mann-Kendall technique only detects the presence of a trend and does not quantify it. An approach often used in conjunction with the Mann-Kendall technique is the calculation of a slope using the method of Sen (Gilbert, 1987). Calculating the Sen slope for  $NO_X$ , showed that the concentration

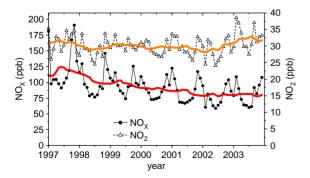


Fig. 1. Trends in monthly mean concentrations of  $NO_X$  and  $NO_2$  averaged across 36 air pollution monitoring sites in London (1997–2003). A 12 month moving average trend has been fitted to both datasets.

decreased on average by 6.1 ppb yr<sup>-1</sup>. Individual site trends are considered later in Table 2.

Despite a comparatively large reduction in concentrations of NO<sub>X</sub> at roadside locations, concentrations of NO<sub>2</sub> have remained relatively constant. It is important to consider why there has been no statistically significant downward trend in roadside NO2 concentrations. It was known that exceptional meteorological conditions during 2003 led to elevated concentrations of many pollutants (Fuller, 2004), but it has not been established how these conditions affected NO<sub>2</sub> concentrations. This paper addresses the following issues. First, a simple chemical modelling technique is applied to the data from air pollution measurement sites to quantify the trend in the road transport NO<sub>2</sub>/NO<sub>X</sub> emissions ratio (assumed to relate to primary NO<sub>2</sub> emissions, i.e., directly emitted NO<sub>2</sub>) on a monthly basis. Second, the results from this technique are checked against other data to ensure their validity. Finally, trends in NO<sub>2</sub>/NO<sub>X</sub> ratios are considered in relation to factors affecting this ratio, such as road transport emissions control technologies.

#### 2. Method

### 2.1. Estimating the $NO_2/NO_X$ emissions ratio from ambient measurements

Carslaw and Beevers (2005b) describe a technique for estimating the  $NO_2/NO_X$  ratio from road transport emissions based on an analysis of ambient measurements. In London, there is a large number (>30) of roadside monitoring sites that measure hourly concentrations of  $NO_X$  and  $NO_2$  using the chemiluminescence technique. London is therefore an excellent location in which to consider the road transport contribution to  $NO_X$  and  $NO_2$  concentrations. The increment in  $NO_X$  and  $NO_2$  concentrations above nearby background values is dominated by contributions from road traffic sources adjacent to the roadside monitoring sites.

Briefly, the increment in  $NO_X$  and  $NO_2$  concentration at a roadside monitoring site above a nearby background site is partitioned into NO<sub>2</sub> that is chemically derived through the reaction between NO and ozone (O<sub>3</sub>) and that which is emitted directly by road vehicles. The technique is constrained by assuming that the difference between roadside and background NO<sub>X</sub> concentrations each hour is due to  $NO_X$  emissions from road vehicles adjacent to the roadside monitoring site. A simple set of chemical equations is used to describe the time-dependent change in NO, NO2 and O3 concentrations as vehicle plumes are mixed with background air. By considering different values of the assumed NO<sub>2</sub>/  $NO_X$  emissions ratio from road traffic and the time available for the NO-O<sub>3</sub> reaction to take place, the best agreement between hourly NO2 concentrations between

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