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Phosphorus Use Efficiency of Bio-Based Fertilizers: Bioavailability and Fractionation

Céline VANEECKHAUTE^{1,2,*}, Joery JANDA¹, Peter A. VANROLLEGHEM², Filip M. G. TACK¹ and Erik MEERS¹

¹Laboratory for Analytical and Applied Ecochemistry, Faculty of Bioscience Engineering, Ghent University, Ghent 9000 (Belgium) ²modelEAU, Département de Génie Civil et de Génie des Eaux, Faculté des Sciences et de Génie, Université Laval, Québec, QC G1V 0A6 (Canada)

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ABSTRACT

Although to date some technologies producing bio-based phosphorus (P) fertilizers have been proposed and implemented, the efficient use of the recovered products is still limited due to legislative constraints and lack of insights in the P release with time and in the corresponding mechanisms. The aim of this work was to evaluate the fertilizer performance in terms of P release and use efficiency of recovered struvite, FePO₄-sludge, digestate, and animal manure as compared to fossil reserve-based mineral triple superphosphate (TSP). First, product physicochemical characteristics and P fractions in the context of European fertilizer legislation were assessed. Next, a controlled greenhouse experiment was set up to evaluate plant reactions as well as changes of P availability in a sandy soil with high P status and a Rheinsand soil with low P status. Soil P fractions were determined in the extracts with water, ammonium lactate and CaCl₂, and in soil solution sampled with Rhizon soil moisture samplers. Based on all results, it is worth conducting long-term field trials to evaluate the P release effect of struvite and digestate as compared to animal manure and TSP on different soil types with varying P status. These products showed promise as sustainable substitutes for conventional P fertilizers and could contribute to a more efficient use of P in agriculture. A refined classification of P application standards/recommendations in terms of soil P status, soil texture, and fertilizer characteristics, next to the crop P demand, is recommended. Moreover, the additional use of Rhizon samplers for determination of direct available P, including dissolved organic P, is proposed for better understanding and categorization of different P fertilizers in environmental and fertilizer legislations.

Key Words: chemical soil analysis, digestate, nutrient recovery, Rhizon samplers, struvite

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The rapid and increasing phosphorus (P) consumption in modern agriculture has raised concerns on both its supply security (Godfray *et al.*, 2010; Elser and Bennett, 2011; Neset and Cordell, 2012; Scholz and Wellmer, 2013) and its impact on the environment (soil P accumulation, leaching, and/or eutrophication) (Syers *et al.*, 2008; Kang *et al.*, 2011; Ranatunga *et al.*, 2013). Consequently, the effective use of soil P and P-containing mineral and organic fertilizers, as well as the cradle-to-cradle recycling of P from municipal, agricultural, and other biodegradable waste sources as green renewable fertilizers with high P use efficiency (*e.g.*, slow-release granules), has become highly important (Syers *et al.*, 2008; Ma *et al.*, 2011; Schröder *et al.*, 2011; Huang *et al.*, 2012; Zhang *et al.*, 2013).

Traditional P removal processes from waste(water) streams often involve the addition of Fe or Al salts, resulting in the production of substantial quantities of Fe/AlPO₄-sludge (Sano et al., 2012). Alternatively, in the past decades, the controlled precipitation of struvite (MgNH₄PO₄ \cdot 6H₂O) through addition of Mg to waste flow has gained interest as a route for P recovery (Latifian et al., 2012; Ryu et al., 2012). Moreover, the anaerobic (co-)digestion of animal manure, sludges, organic biological food waste, and/or energy crops has proven to be an effective technology for bio-energy production and release/mineralization of nutrients, which are concentrated in the remaining digestate (Fehrenbach et al., 2008). During a preceding field-scale assessment, it has been observed that the use of a formulated mixture of digestate with its liquid fraction in agriculture as a substitute for animal manure may stimulate P mobilization in the soil, thereby increasing the use efficiency of soil minerals (to be confirmed) (Vaneeckhaute et al., 2013, 2014). Especially in P saturated regions (e.g., Flanders, Quebec, eastern China, Italy,

^{*}Corresponding author. E-mail: CelineVaneeckhaute@gmail.com.

northern Spain, *etc.*; MacDonald *et al.*, 2011), the extraction of P from agricultural fields is relevant for local reuse (*e.g.*, in the horticultural sector or for plants with high P demand) and/or for industrial purposes. On the other hand, although the use of liquid digestate, with high effective N:P ratio, is interesting in terms of current legislative fertilization standards, its supply of plant-available P may be insufficient, depending on crop P demand and soil P status. Hence, additional fertilization with a source of bioavailable P may be required.

It must be understood that only a small proportion (15%-20%) of the total amount of P uptake in the plant ($\pm 2.5 \text{ kg P}_2\text{O}_5 \text{ ha}^{-1} \text{ d}^{-1}$; EFMA, 2000) is directly provided by the fertilizer applied to that crop. The remainder comes from soil reserves. Hence, there must be adequate reserves of readily available P in the soil (Syers et al., 2008). The P status of European soils has been estimated by EFMA (2000). For many countries, about 25% (5%–55%) of soils test as very low and low in readily available P. Such soils require significantly more P application to increase soil reserves and thus soil fertility. On the other hand, in many countries, some 40% (15%–70%) of soils test as high and very high in readily available P. On such soils, when growning the crops that have small, inefficient root systems, but a large daily uptake of P at critical growth stages, it may be necessary to apply more P. On soils with a medium P analysis value, applications need to sustain the P status. This may require a small extra amount of P on top of that removed with the harvested crop (EFMA, 2000).

All the above clearly indicates the relevance of fundamental comparative research on the P release pattern over time of potential sustainable alternatives for chemical P fertilizers produced from natural and exhaustive phosphate rock and/or for animal manure (products with improved P use efficiency). Such an evaluation is essential to determine the agricultural potential of new fertilizers and their responsible application (Erro *et al.*, 2011). Yet, such studies are currently lacking in literature for the above-mentioned biobased products (struvite, digestate, Fe/AlPO₄-sludge), although their production and availability is on the rise (Vaneeckhaute, 2015).

The performance of a fertilizer can be evaluated *via* product fractionation, plant reaction analysis, and/or chemical soil analysis (Sissingh, 1971; Prummel and Sissingh, 1983; Singh *et al.*, 2005; Dekker and Postma, 2008; van Dam and Ehlert, 2008; Millier and Hooda, 2011; Wang *et al.*, 2013). P fractionation of fertilizers is, in general, based on the P solubility in sol-

vents with different strength and selectivity (Frossard et al., 2002; He et al., 2004, 2007). With respect to European legislation, the most important solvents are, ranked from strong to weak: i) mineral acid, ii) neutral ammonium citrate (NAC) solution, and iii) water (EU, 2003). Next to the P solubility, the fertilizer performance is usually expressed as bioavailability indices, such as P use efficiency (PUE). It can be based on the fresh weight (FW) and dry weight (DW) biomass yield, the growth rate, the P uptake (rate), and the degree and rate in which the P status of the soil changes, as determined by chemical methods (van Dam and Ehlert, 2008). Previous studies have shown that the crop response to P fertilization gives insufficient guidance to determine the fertilizer performance (no correlation), while chemical soil analyses can be conclusive (Prummel and Sissingh, 1983; Arendás and Csathó, 2002; van Dam and Ehlert, 2008). Therefore, most studies evaluating P fertilizers to date are based on soil bioavailability indices. Soil measurements can be divided into P capacity and P intensity of the soil, based on the strength of the extraction method. The P intensity gives an indication of the total amount of inorganic P which is directly available for the plant during a short period of time, while the P capacity gives an indication of the amount of P that may be released in the long term (Dekker and Postma, 2008).

In some countries, e.g., the Netherlands, Switzerland, and Norway, fertilizer recommendations are based on the P status of the soil, measured as the ammonium lactate- and water-extractable P (PAL and P_{water} , respectively). The P_{AL} value is a measure of soil P capacity, whereas the P_{water} value reflects a combination of soil P capacity and intensity. In the latest decade, the method of plant-available elements (PAE) has also received increased attention. It concerns a multi-element extraction with $0.01 \text{ mol } L^{-1}$ $CaCl_2$, and hence provides a simple alternative for many extraction procedures that are currently used for single nutrient (van Erp et al., 1998; Houba et al., 2000; Ehlert *et al.*, 2006). With respect to $CaCl_2$ -extractable $P(P_{CaCl_2})$, this measurement gives an indication of soil P intensity (Houba et al., 2000). An important limitation of all these standard methods is that root formation, soil compaction, and mineralization of organic matter are not or not sufficiently accounted for (Ehlert et al., 2006; Soine, 2009; Amoakwah et al., 2013). Underestimations have been observed in literature, especially for the determination of direct available P (Amoakwah et al., 2013; Sánchez-Alcalá et al., 2014). Alternatively, the use of Rhizon soil moisture samplers (SMS) allows assessing the total amount of P

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