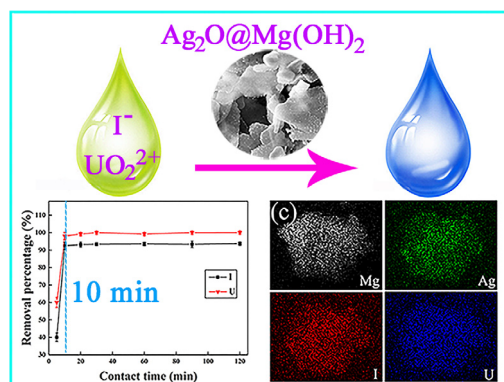




## Regular Article

One-step synthesis of  $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$  nanocomposite as an efficient scavenger for iodine and uraniumYuan-Yuan Chen<sup>a</sup>, Sheng-Hui Yu<sup>a</sup>, Qi-Zhi Yao<sup>b,\*</sup>, Sheng-Quan Fu<sup>c</sup>, Gen-Tao Zhou<sup>a,\*</sup><sup>a</sup> CAS Key Laboratory of Crust-Mantle Materials and Environments, School of Earth and Space Sciences, University of Science and Technology of China, Hefei 230026, PR China<sup>b</sup> School of Chemistry and Materials Science, University of Science and Technology of China, Hefei 230026, PR China<sup>c</sup> Hefei National Laboratory for Physical Sciences at Microscale, University of Science and Technology of China, Hefei 230026, PR China

## GRAPHICAL ABSTRACT



## ARTICLE INFO

## Article history:

Received 9 August 2017

Revised 18 September 2017

Accepted 20 September 2017

Available online 21 September 2017

## Keywords:

 $\text{Mg}(\text{OH})_2$  nanoplates $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$  nanocomposite

One-step method

Simultaneous removal

 $\text{I}^-$  and  $\text{UO}_2^{2+}$ 

## ABSTRACT

$\text{Ag}_2\text{O}$  nanoparticles anchored on the  $\text{Mg}(\text{OH})_2$  nanoplates ( $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$ ) were successfully prepared by a facile one-step method, which combined the  $\text{Mg}(\text{OH})_2$  formation with  $\text{Ag}_2\text{O}$  deposition. The synthesized products were characterized by a wide range of techniques including powder X-ray diffraction (XRD), field emission scanning electron microscopy (FESEM), transmission electron microscopy (TEM), selected area electron diffraction (SAED), and nitrogen physisorption analysis. It was found that  $\text{Ag}_2\text{O}$  nanoparticles anchored on the  $\text{Mg}(\text{OH})_2$  nanoplates show good dispersion and less aggregation relative to the single  $\text{Ag}_2\text{O}$  nanoaggregates. In addition, iodide ( $\text{I}^-$ ) removal by the  $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$  nanocomposite was studied systematically. Batch experiments reveal that the nanocomposite exhibits extremely high  $\text{I}^-$  removal rate ( $<10$  min), and  $\text{I}^-$  removal capacity is barely affected by the concurrent anions, such as  $\text{Cl}^-$ ,  $\text{SO}_4^{2-}$ ,  $\text{CO}_3^{2-}$  and  $\text{NO}_3^-$ . Furthermore,  $\text{I}^-$  and  $\text{UO}_2^{2+}$  could be simultaneously removed by the nanocomposite with high efficiency. Due to the simple synthetic procedure, the excellent removal performances for iodine and uranium, and the easy separation from water, the  $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$  nanocomposite has real potential for application in radioactive wastewater treatment, especially during episodic environmental crisis.

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## 1. Introduction

Recently, the concern on nuclear waste leakage has drawn increasing attention globally since the nuclear accident at Fukushima in 2011 [1]. The radioactive iodine is one kind of the major by-products of uranium and plutonium fission, and the half-life of radioiodine isotope differs from about 8 days ( $^{131}\text{I}$ ) to  $1.6 \times 10^7$  years ( $^{129}\text{I}$ ) [2–4]. In addition, radioactive iodine is also widely applied to the diagnosis and treatment of various diseases [5,6]. For instance,  $^{131}\text{I}$ , often used in hyperthyroidism, thyroid cancer diagnosis and metabolic therapies [7], is habitually dumped into domestic sewer system, and thus is one of the radionuclides most often detected in the medical sewage [8–10]. Because of the non-selective uptake of iodine isotope by human body, radioactive iodine can be accumulated in the human thyroid gland and lead to increase of metabolic disorders, mental retardation, and thyroid cancer, especially for children [11]. Studies on iodine speciation in aqueous environment have demonstrated that pH, salinity, natural organic matter (NOM) and microorganisms are involved in regulating iodine biogeochemical processes [11,12]. Dominant iodine species include iodide ( $\text{I}^-$ ), iodate ( $\text{IO}_3^-$ ), and organo iodine in the natural environment [13,14]. It has been proposed that  $\text{I}^-$  is the major species in deep radioactive waste repositories and is stable almost over the entire pH range [4,15]. Due to the poor removal abilities of many nanomaterials to anions, the implementation in developing nanomaterials for the efficient immobilization of radioactive iodine anions, especially  $\text{I}^-$ , from aqueous environment is lagged [15].

Various nanomaterials, such as organic anion exchange resins [16], clay minerals [15,17], Ag-based materials [18–27], layered double hydroxide [28–31], and compounds containing  $\text{Hg}^{2+}$  [32],  $\text{Bi}^{3+}$  [4,33], and  $\text{Cu}^+$  [34,35], have been recently studied for the treatment of water contaminated by radioactive  $\text{I}^-$ . However, organic anion resin are unstable under high dose of radiation and elevated temperatures [36], and many inorganic materials often display relatively low removal capacities and slow uptake kinetics [4,24]. Among these materials,  $\text{Ag}_2\text{O}$  is a potential capture agent for radioactive  $\text{I}^-$  because of the efficient  $\text{I}^-$  immobilization by  $\text{Ag}_2\text{O}$  to form  $\text{AgI}$  precipitate. Nevertheless, it has been found that the smaller the nanoparticles are, the higher the tendency of aggregation becomes, which stems from high surface free energy. The formation of aggregates could significantly decrease the surface area of  $\text{Ag}_2\text{O}$ , thereby leading to a sharp deterioration of activity and efficiency [24]. Therefore, using  $\text{Ag}_2\text{O}$  directly for the treatment of radioactive  $\text{I}^-$  is impractical. Anchoring  $\text{Ag}_2\text{O}$  nanoparticles onto the supported-materials could reduce the aggregation of  $\text{Ag}_2\text{O}$  nanoparticles, and thus may improve their performance in  $\text{I}^-$  uptake. Recently, many  $\text{Ag}_2\text{O}$ -based nanocomposites have been reported for the efficient removal of radioactive  $\text{I}^-$  [19–24]. For example,  $\text{Ag}_2\text{O}$  anchored titanate nanotubes, nanofibers and nanolamina have been reported to act as scavengers for highly efficient and selective capture of  $\text{I}^-$  [19,20]. In addition,  $\text{Ag}_2\text{O}$  grafted sodium niobate nanofibers also showed good removal performance for  $\text{I}^-$  [24]. These findings have provided new insights into the efficient capture of radioactive iodine. However, some shortcomings of the reported supported-materials, such as complexity of preparation, high cost and sensitivity to solution pH, still limit the wide application of  $\text{Ag}_2\text{O}$ . Thus, the search for new substrate that is facilely synthesized and stable under complicated conditions is still crucial and urgent.

As a nontoxic, low cost and environmentally friendly material, nanoscale brucite (nano- $\text{Mg}(\text{OH})_2$ ) is widely applied in the environmental remediation [37–43]. For instance, Li et al. reported that self-supported flowerlike  $\text{Mg}(\text{OH})_2$  could recycle rare earth elements from low concentration wastewater via ion-exchange [37].

A strategy was proposed by Liu et al. to simultaneously realize both the recycle of  $\text{Mg}(\text{OH})_2$  nanoadsorbent and enrichment of diluted  $\text{Cr}(\text{VI})$  from wastewater, which was based on the reversible phase transformation between platelet-like  $\text{Mg}(\text{OH})_2$  nanoparticles and rod-like  $\text{MgCO}_3 \cdot 3\text{H}_2\text{O}$  bulk crystals [39]. In addition,  $\text{Mg}(\text{OH})_2$  micro/nanorods also showed highly selective adsorption capacity and fast adsorption rate for the removal of dilute anionic dye, and the adsorbed dye was successfully desorbed by carbonation, resulting in a  $\sim 4000$  fold enrichment of the dye solution [40]. It's well established that nano- $\text{Mg}(\text{OH})_2$ , having a hexagonal symmetry structure [44], usually exhibits nanoplate-like morphology with a large specific surface area [45], which can potentially provide an ideal substrate for  $\text{Ag}_2\text{O}$  nanoparticle loading. Moreover, nano- $\text{Mg}(\text{OH})_2$  could be facilely synthesized by the hydrolysis of periclase ( $\text{MgO}$ ), which is abundant in nature, and thus could be helpful to greatly reduce the preparation cost. In particular, Chen et al. recently reported that a high uranium removal capacity of  $\text{Mg}(\text{OH})_2$  could be found when the initial uranium concentration crosses over a threshold [46]. Therefore,  $\text{Mg}(\text{OH})_2$  nanoplates can not only act as a substrate for  $\text{Ag}_2\text{O}$  nanoparticles, but also remove uranium efficiently from solution. That is,  $\text{Ag}_2\text{O}$  coupled with  $\text{Mg}(\text{OH})_2$  can be potentially applied to the simultaneous removal of iodine and uranium from radioactive wastewater. However, to the best of our knowledge, no research has been reported on the removal of radioactive iodine, especially on the simultaneous removal of iodine and uranium, by  $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$  nanocomposite.

Herein,  $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$  nanocomposite is successfully prepared by a facile one-step method, which combined the  $\text{Mg}(\text{OH})_2$  formation with  $\text{Ag}_2\text{O}$  deposition, and the removal ability of the nanocomposite to  $\text{I}^-$  is systematically investigated. In addition, considering the fact that multiple radionuclides can coexist in radioactive wastewater, the simultaneous removal of  $\text{I}^-$  and  $\text{UO}_2^{2+}$  by the nanocomposite is also investigated.

## 2. Materials and methods

### 2.1. Materials

All chemicals are commercially available and used as received without further purification.  $\text{AgNO}_3$ ,  $\text{MgO}$  and  $\text{NaI}$  were purchased from Sinopharm Chemical Reagent Co., Ltd., and are of analytical grade.  $\text{UO}_2(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$  was obtained from Beijing Chemical Industry. Analytical grade nitric acid ( $\text{HNO}_3$ ) and sodium hydroxide ( $\text{NaOH}$ ) were used to adjust the pH whenever necessary. Deionized water was used in all experiments.

### 2.2. Preparation of $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$ nanocomposite

The  $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$  nanocomposite was synthesized using a facile one-step method at room temperature. In a typical procedure, 0.46 g (2.71 mmol)  $\text{AgNO}_3$  was first dissolved into 50 mL of deionized water. Then 0.40 g of  $\text{MgO}$  (10 mmol) was slowly added into the  $\text{AgNO}_3$  solution under continuously stirring. After 12 h of stirring, the precipitate was collected by centrifugation, washed three times with deionized water, and then dried in the vacuum at 60 °C for 12 h. The finally obtained precipitate was designated as sample SOMH-1. For comparison, different amounts of  $\text{AgNO}_3$  (0.095 and 0.85 g) were used to obtain  $\text{Ag}_2\text{O}@\text{Mg}(\text{OH})_2$  nanocomposite with different  $\text{Ag}_2\text{O}$  loadings by the same procedures, which were labeled as samples SOMH-2 and SOMH-3, respectively. The loading of  $\text{Ag}_2\text{O}$  in the nanocomposite is determined by inductively coupled plasma atomic emission spectroscopy (ICP-AES, optima 7300 DV) after sample dissolved by  $\text{HNO}_3$  solution. In addition,  $\text{Mg}(\text{OH})_2$  nanoplates were prepared under the same experimental

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