



Uncoupling of ammonia oxidation from nitrite oxidation: Impact upon nitrous oxide production in non-cropped Oregon soils



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ABSTRACT

The factors controlling the relative contributions of ammonia- (NH₃) oxidizing archaea (AOA) and bacteria (AOB) to nitrification and nitrous oxide (N₂O) production in soil remain unclear. A study was conducted to examine the contributions of AOA and AOB to nitrification, nitrite (NO₂⁻) accumulation, and NO₂⁻-affected N₂O production in three non-cropped Oregon soils. Nitrification potential rates in the three soils ranged seven-fold from 0.15 to 1.08 μmol N g⁻¹ d⁻¹, with AOA contributing 64–71% of the total activity. AOA- and AOB-driven NO₂⁻ accumulation represented 8–100% of total NO₂⁻ + NO₃⁻ accumulation, persisted over 48 h, and was accompanied by acetylene-sensitive, ammonium- (NH₄⁺) stimulated N₂O production. Ammonium- and NO₂⁻-dependent N₂O production occurred when both AOA and AOB, or AOA alone were active. By adding the NO₂⁻-oxidizing bacteria, *Nitrobacter vulgaris*, to soil slurries to increase NO₂⁻-oxidizing capacity, both NO₂⁻ accumulation and N₂O production were prevented, while the overall rate of nitrification was unaffected. Yields of N₂O-N amounted to 0.05 ± 0.01% of total NO₂⁻ + NO₃⁻-N accumulation in the presence of supplemental NH₄⁺, and 0.28 ± 0.11% in the presence of both supplemental NH₄⁺ + NO₂⁻. Regression analysis of the N₂O production against NO₂⁻ accumulation over 24 h revealed a positive, non-linear relationship for N₂O production by both AOA plus AOB and by AOA alone. Values of V_{max} ranged 12-fold from 0.05 to 0.62 nmol N₂O g⁻¹ d⁻¹, and predicted K_m values for NO₂⁻ ranged 15-fold from 0.02 to 0.30 μmol NO₂⁻ g⁻¹ soil. These findings provide new insights into the impact of NO₂⁻ accumulation in soils on N₂O production by both AOA and AOB, and show that NO₂⁻ accumulation primarily drives N₂O formation in these soils, and increases N₂O yield by both AOA and AOB.

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1. Introduction

Nitrification is the process whereby ammonia (NH₃) is oxidized sequentially to nitrite (NO₂⁻) and nitrate (NO₃⁻). The first step of nitrification is carried out by NH₃-oxidizing bacteria (AOB) and thaumarchaea (AOA) (Arp and Stein, 2003; Leininger et al., 2006; Vajjala et al., 2013). Several studies have shown that the process of NH₃ oxidation can be a major source of aerobically produced N₂O, and can contribute 36–57% of total N₂O production from soils (Kool et al., 2011; Wrage et al., 2001; Zhu et al., 2013). Whereas AOA

and AOB are generally abundant and widely distributed in soils (Leininger et al., 2006; Prosser and Nicol, 2012; Taylor et al., 2012, 2013), few studies have examined the relative contributions of AOA and AOB to soil nitrification (Chen et al., 2013; Daebeler et al., 2015; Giguere et al., 2015; Taylor et al., 2010, 2013; Wessén et al., 2010; Lu et al., 2015). Furthermore, despite the activities of AOA and AOB having the potential to produce N₂O (Kozłowski et al., 2014; Poth and Focht, 1985; Santoro et al., 2011; Shaw et al., 2006; Stieglmeier et al., 2014; Stein, 2011), to our knowledge there is only one study in the literature that has examined the relative contributions of AOA and AOB to nitrifier-dependent N₂O production in soil (Hink et al., 2016). There is considerable interest in determining the factors that influence the proportion of NH₃ oxidized that is transformed to N₂O, and if the relative contributions of AOA and AOB might influence the latter value (Jung et al.,

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2013; Mørkved et al., 2007; Shaw et al., 2006; Stieglmeier et al., 2014).

There is a growing body of evidence that aerobic N₂O production in soil may be associated with NO₂⁻ accumulation (Maharjan and Venterea, 2013; Venterea, 2007; Venterea et al., 2015). Several studies have demonstrated that NO₂⁻ accumulates in soil under conditions where NH₃-oxidizing activity is stimulated (Müller et al., 2006), and/or NO₂⁻-oxidizing activity is negatively affected by additions of urea (Burns et al., 1995; Chapman and Liebig, 1952; Ma et al., 2015; Shen et al., 2003; Venterea, 2007) or anhydrous NH₃ (Maharjan and Venterea, 2013; Venterea et al., 2015). Production of N₂O by AOB has been demonstrated to be stimulated by NO₂⁻ (Shaw et al., 2006) and most AOB possess both NO₂⁻ nitrite reductase (NirK) and nitric oxide reductase (NorB) which enable them to carry out NO₂⁻-dependent N₂O production (Cantera and Stein, 2007; Kozłowski et al., 2014). In the case of AOA, although they possess the gene encoding for NirK (Spang et al., 2012; Walker et al., 2010), the gene encoding for nitric oxide reductase has not been detected (Hatzenpichler, 2012; Kozłowski et al., 2016). Although it has been suggested that AOA can abiotically produce N₂O, the isotopic signature of N₂O produced from AOA enrichments suggests that NO₂⁻ is involved in N₂O production (Jung et al., 2013; Stieglmeier et al., 2014), and a positive relationship was observed between NO₂⁻ concentration and N₂O production by marine AOA enrichment cultures (Santoro et al., 2011).

Nonetheless, only one study has examined the relative importance of AOA and AOB driven NH₃ oxidation to N₂O production (Hink et al., 2016), and no study has examined the importance of NO₂⁻ accumulation on AOA- and AOB-dependent N₂O production. Indeed, Hink et al. (2016) measured both AOA- and AOB-dependent N₂O production over a 28-d incubation of a cropped UK sandy loam soil and found KCl-extractable NO₂⁻ levels to be undetectable. We have identified Oregon soils with significant nitrification contributions from both AOA and AOB (Taylor et al., 2013; Giguere et al., 2015), and that also accumulate NO₂⁻ when nitrification is stimulated by NH₄⁺ additions. In addition, with our recent discovery of the selective AOB inactivator, 1-octyne (Taylor et al., 2013), we have formulated the following objectives. These are: to determine to what extent AOA and AOB-driven NH₃ oxidizing activities contribute to N₂O production, and to determine the influence of NO₂⁻ accumulation on AOA and AOB-driven N₂O production.

2. Materials and methods

2.1. Soil sampling and location

Three locations in Oregon (Pendleton, Madras, and Klamath Falls) were selected for this study and are described in detail elsewhere (Giguere et al., 2015). At each location, 4 replicates of cropped and non-cropped soils were sampled from adjacent sites on the same soil series Pendleton (Walla Walla silt loam), Madras (Madras loam), and Klamath (Fordney loamy fine sand). A preliminary survey showed that non-cropped soils accumulated NO₂⁻ after nitrification was stimulated with 1 mM NH₄⁺ additions as described elsewhere (Giguere et al., 2015; Taylor et al., 2012).

2.2. Soil slurry design

Soils were removed from 4 °C storage and composite 5-g portions of soil were added to 125-ml Wheaton bottles, wet to approximately field capacity, capped loosely with butyl stoppers, and pre-incubated at room temperature (21 °C) for 24 h. Each bottle received 15 ml of water, was amended depending on the experiment, and was capped tightly. Soil slurries were shaken

continuously at 200 rpm at 25 °C. Gas samples for N₂O analysis were collected through the butyl stoppers at 24 and 48 h for all experimental incubations. Acetylene (6 μM_{aq}) was used to inhibit ammonia-oxidizing activity. Previous studies of these soils found no evidence of acetylene-insensitive nitrification, implying that all ammonia oxidation was chemolithotrophic (Giguere et al., 2015; Taylor et al., 2013). Octyne (4 μM_{aq}) was used to inactivate AOB activity, leaving AOA activity unaffected (Giguere et al., 2015; Hink et al., 2016; Lu et al., 2015; Taylor et al., 2013, 2015). Octyne was prepared by adding 40 μl liquid octyne to a Wheaton bottle with a 155 ml headspace, with several glass beads and over-pressured with 100 ml air, and a 2.8 ml aliquot was added to each sample bottle.

2.3. Analysis of NO₂⁻, NO₃⁻, NH₄⁺, pH and N₂O

Initial pH measurements were made in a 2:1 soil water slurry and ranged from 7.2 to 7.6.

Concentrations of NO₂⁻ and NO₃⁻ were determined as described elsewhere (Miranda et al., 2001; Taylor et al., 2013). Briefly, aliquots of soil slurries were sampled from sealed Wheaton bottles, centrifuged, and were immediately analyzed. Nitrite was measured colorimetrically using Griess reagents, and NO₃⁻ was measured using a vanadium reduction assay in which NO₃⁻ is reduced to NO₂⁻ and the total NO₂⁻ + NO₃⁻ measured (Miranda et al., 2001). The NO₃⁻ concentration was calculated as the difference between NO₂⁻ + NO₃⁻ and NO₂⁻ accumulations. Nitrification rates were calculated as the net accumulation of NO₂⁻ + NO₃⁻ above the acetylene controls. Detection limits for NO₂⁻ were 0.02 μmol NO₂⁻ g⁻¹ soil, and 0.05 μmol NO₃⁻ g⁻¹ soil for NO₃⁻.

NH₄⁺ extractions were conducted independently from NO₂⁻ or NO₃⁻ by extracting 5 g portions of soil in 15 ml 2 M KCl for 1 h. Extracts for NH₄⁺ analysis were frozen until analysis and measured colorimetrically as described by Mulvaney (1996).

N₂O concentration in the gas phase was determined using a Varian Model 3700 gas chromatograph equipped with an electron capture detector as described previously (Mellbye et al., 2016). Total N₂O production from the soil was calculated as described by Tiedje (1994) using the equation

$$M = C_s(V_g + V_l \alpha) \quad [1]$$

where, M is total N₂O, C_s is N₂O concentration in the gas phase, V_g is total gas volume, V_l is volume of the liquid and α is the Bunsen absorption coefficient for N₂O at 25 °C (0.544). The detection limits for N₂O production were 0.015 nmol g⁻¹ soil. Rates of N₂O formation were calculated as the difference between the acetylene control N₂O levels and N₂O accumulation at 24 h and 48 h N₂O yields were calculated using the equation

$$\frac{N_2O - N}{(NO_2^- - N + NO_3^- - N)} \quad [2]$$

2.4. Incubations to establish the impact of NH₄⁺, and NO₂⁻ on N₂O production by AOB + AOA and AOA alone

An experiment was conducted to examine the effect of supplemental NH₄⁺ and NO₂⁻ on nitrification activity and N₂O production by the combination of AOA + AOB (-octyne) and by AOA alone (+octyne). Soil slurry incubations for each of the 3 soils were conducted in the presence or absence of supplemental 1 mM NH₄⁺ and in the presence or absence of supplemental 1 mM NO₂⁻. NO₂⁻ and NO₃⁻ concentrations were measured at 0, 6, 24, and 48 h.

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