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# Effects of pH and $H_2O_2$ on ammonia, nitrite, and nitrate transformations during $UV_{254nm}$ irradiation: Implications to nitrogen removal and analysis



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# HIGHLIGHTS

- $\bullet$  Conversions among DINs were studied by  $UV_{254nm}$  photolysis with and without  $H_2O_2.$
- Complete oxidation of NH<sub>3</sub> was obtained within 1 h with proper pH & H<sub>2</sub>O<sub>2</sub> controls.
- $\label{eq:holo-oxidation} \bullet \mbox{ Nitrite photo-oxidation depended} strongly on $H_2O_2$ dose but less so on $p$H.$
- Nitrate photo-reduction generally increased with increasing pH.
- Nitrogen balance was achieved in all irradiation experiments.

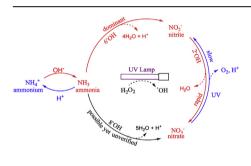
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# ABSTRACT

In order to achieve better removal and analyses of three dissolved inorganic nitrogen (DIN) species via ultraviolet-activated hydrogen peroxide (UV/H<sub>2</sub>O<sub>2</sub>) process, this study systematically investigated the rates of photo-oxidations of ammonia/ammonium (NH<sub>3</sub>/NH $^{+}$ ) and nitrite (NO $^{-}$ ) as well as the photo-reduction of nitrate ( $NO_3$ ) at varying pH and  $H_2O_2$  conditions. The results showed that the mass balances of nitrogen were maintained along irradiation despite of interconversions of DIN species, suggesting that no nitrogen gas  $(N_2)$  or other nitrogen-containing compound was formed. NH<sub>3</sub> was more reactive than NH<sub>4</sub><sup>+</sup> with hydroxyl radical ( $\cdot$ OH), and by a stepwise H<sub>2</sub>O<sub>2</sub> addition method NH<sub>3</sub>/NH<sub>4</sub><sup>+</sup> can be completely converted to  $NO_{\overline{3}}$ ;  $NO_{\overline{2}}$  underwent rapid oxidation to form  $NO_{\overline{3}}$  when  $H_2O_2$  was present, suggesting that it is an intermediate compound linking NH<sub>3</sub>/NH<sub>4</sub> and NO<sub>3</sub>; but once H<sub>2</sub>O<sub>2</sub> was depleted, NO<sub>3</sub> can be gradually photo-reduced back to  $NO_2^-$  at high pH conditions. Other than  $H_2O_2$ , the transformation kinetics of DINs were all dependent on pH, but to varying aspects and extents: the NH<sub>3</sub> photo-oxidation favored a pH of 10.3, which fell within the pK<sub>a</sub> values of  $NH_4^+$  (9.24) and  $H_2O_2$  (11.6); the  $NO_3^-$  photo-reduction increased with increasing pH provided that it exceeds the  $pK_a$  of peroxynitrous acid (6.8); while the NO<sub>2</sub> photo-oxidation remained stable unless the pH neared the  $pK_a$  of  $H_2O_2$  (11.6). The study thereby demonstrates a picture of the evolutions of DIN species together during UV/H<sub>2</sub>O<sub>2</sub> irradiation process, and for the first time presents a method to achieve complete conversion of  $NH_4^+$  to  $NO_3^-$  with  $UV/H_2O_2$  process.

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# 1. Introduction

Ammonium  $(NH_4^+)$ , nitrite  $(NO_2^-)$ , and nitrate  $(NO_3^-)$  are the three most common dissolved inorganic nitrogen (DIN) compounds in water, and their widespread existences may pose many negative impacts, such as eutrophication and acidification (Camargo and Alonso, 2006), on aquatic systems, and cause certain toxicological effects, like methemoglobinemia, to humans (Manassaram et al., 2005; Camargo and Alonso, 2006). Consequently, they have been regulated by many water and wastewater authorities around the world typically including USA, China, and Germany. To mitigate their risks, researchers have taken the initiative in applying measures to transform or remove them from water, including biological nitrification and denitrification (Rittmann and McCarty, 2001), air stripping (Liao et al., 1995), ion exchange (Jorgensen and Weatherley, 2003), and chemical precipitation (Li et al., 1999).

Recently, the uses of ultraviolet (UV)-based methods for DINs removal have received significant attention (Pollema et al., 1992; Mack and Bolton, 1999; Lee et al., 2002; Zhang et al., 2005; Zhu et al., 2005; Doudrick et al., 2013; Krasae and Wantala, 2016). In general, these photochemical processes might undergo two reaction routes. One route involves the photo-oxidation of ammonia/ ammonium  $(NH_3/NH_4^+)$  to yield  $NO_2^-$ ,  $NO_3^-$ , and nitrogen gas  $(N_2)$ (Wang et al., 1994; Zhu et al., 2005), and the other route involves photo-reduction of  $NO_3^-$  to form  $NO_2^-$  or  $N_2$  (Mack and Bolton, 1999; Zhang et al., 2005: Doudrick et al., 2013: Krasae and Wantala, 2016). However, earlier studies usually examine ammonia oxidation, nitrite oxidation, and nitrate reduction separately, but seldom look into the concurring evolutions of all three DIN species together in a UV process. In addition, although many mechanistic and radicals formation insights have been provided on NH<sub>3</sub>/NH<sup>+</sup><sub>4</sub> oxidation by UV and hydrogen peroxide  $(H_2O_2)$  (Huang et al., 2008), the reported treatment efficiency appeared not satisfying considering its large consumption of  $H_2O_2$ . Meanwhile, since  $NO_2^-$  formation has been documented in both NH<sub>3</sub>/NH<sup>+</sup><sub>4</sub> photo-oxidation and NO<sub>3</sub> photoreduction processes (Motomizu and Sanada, 1995; Mack and Bolton, 1999; Sharpless and Linden, 2001; Keen et al., 2012), it is also important to know which one is quicker and what conditions are favorable.

Other than water and wastewater treatment, the use of UV methods for analyses of nitrogen-containing compounds is another arena of interest for a long time. To the best of our knowledge, some studies have applied UV (Baxley et al., 2004), UV-activated hydrogen peroxide (UV/H2O2) (Scheuer et al., 1995), UV/TiO2 (Takeda and Fujiwara, 1996), and vacuum UV (Gonzalez and Braun, 1995) to convert dissolved organic nitrogen (DON) into DINs to enable determination of total nitrogen (TN) in water by colorimeter or ion chromatographer (IC) (Westerhoff and Mash, 2002; Ma et al., 2014). Although the oxidation efficiency is not as high as other methods for resistant compounds and nitrogen-enriched samples (Bronk et al., 2000), these UV-based TN analytical methods are superior to high temperature combustion (HTC) and high temperature catalytic oxidation (HTCO) methods in not requiring expensive instruments (e.g., combustor), and are also more advantageous than UV persulfate oxidation (UPO) method because they do not introduce external ions into samples so that it is possible to analyze  $\mu g/L$  (i.e., ppb) level of nitrogen in irradiated water by IC, thus making them appealing to measure traces of nitrogen. However, based upon earlier study, the conversion of NH<sup>+</sup><sub>4</sub> into NO<sup>-</sup><sub>x</sub> remained challenging for UV/H<sub>2</sub>O<sub>2</sub> process, with the maximum efficiency being only 26.4% even at favorable experimental conditions (Huang et al., 2008). So, it is necessary to optimize the process. In addition, because  $NO_2^-$  features the lowest detection limit of all DIN species by spectrophotometric methods (Westerhoff and Mash, 2002; Huang et al., 2008), there were once also desires to convert  $NO_3^-$  into  $NO_2^-$  to allow handy detection of traces of  $NO_x^-$  for water and wastewater operators (Zhang and Wu, 1986; Takeda and Fujiwara, 1993; Motomizu and Sanada, 1995). However, up to date, few applications have been reported out of laboratory for real waters using this approach, suggesting that there are probably some unresolved problems that require clarification and further understandings.

In order to achieve better transformation efficiency, an in-depth understanding of the kinetics and influencing factors is needed. Based on existing information, •OH was believed to be responsible for the NH<sub>3</sub>/NH<sup>‡</sup> photo-oxidation in TiO<sub>2</sub> or H<sub>2</sub>O<sub>2</sub>-based methods (Lee et al., 2002; Huang et al., 2008). Hence, the additions of H<sub>2</sub>O<sub>2</sub> are expected to promote transformation of NH<sub>3</sub>/NH<sup>‡</sup>. However, contrary to expectations, lower-than-expected removals of NH<sub>3</sub>/ NH<sup>‡</sup> via UV/H<sub>2</sub>O<sub>2</sub> was observed even with sufficient dosages H<sub>2</sub>O<sub>2</sub> (Huang et al., 2008), suggesting that scavenging effect of •OH was probably present and therefore a higher NH<sub>3</sub>/NH<sup>‡</sup> oxidation may require consideration of other important factors and/or other approaches (Zhu et al., 2005). Hence, the influences of pH, H<sub>2</sub>O<sub>2</sub>, H<sub>2</sub>O<sub>2</sub> dosing method, and feeding NH<sub>3</sub>/NH<sup>‡</sup> concentration on NH<sub>3</sub>/NH<sup>‡</sup> removal deserve to be further explored.

Herein, we systematically examined three photolysis processes step-by-step, including: 1) photo-oxidation of NH<sub>3</sub>/NH<sub>4</sub><sup>+</sup> to yield  $NO_2^-$  and  $NO_3^-$  in the presence of  $H_2O_2$ ; 2) photo-oxidation of  $NO_2^-$  to produce  $NO_3^-$  with  $H_2O_2$ ; and 3) photo-reduction of  $NO_3^-$  to form  $NO_2^-$  without  $H_2O_2$ . Low pressure UV lamps emitting 254 nm wavelength (UV<sub>254nm</sub>) were employed as the energy source, and H<sub>2</sub>O<sub>2</sub> was used as the only oxidant to yield •OH. In addition, although emphasis were put on the influences of pH and H<sub>2</sub>O<sub>2</sub>, the effects of H<sub>2</sub>O<sub>2</sub> dosing mode and feeding DIN concentration was also evaluated. Noteworthy is that since earlier studies have already provided the details of DIN transformation mechanisms and radical formation pathways (Mack and Bolton, 1999; Huang et al., 2008), this study mainly intended to enhance the DIN transformation efficiency and kinetics, instead of repeating the efforts on mechanistic explorations, despite that the results of this study may also help enhance current understanding of these processes.

### 2. Materials and methods

### 2.1. Chemicals

The three dissolved inorganic nitrogen (DIN) compounds used in this study were ammonium chloride, sodium nitrite, and sodium nitrate, all in analytical grades. A stock solution of H<sub>2</sub>O<sub>2</sub> (20 g/L) was made by using a 30% reagent grade solution. Sodium hydroxide (NaOH, 1 or 0.1 M) and sulfuric acid (H<sub>2</sub>SO<sub>4</sub>, 0.1 M) were used to adjust water pH in case of need. All solutions were prepared by ultrapure water with electric resistance of 18.2 M $\Omega$  cm and  $\leq$ 10 µg/ L organic carbon.

#### 2.2. Analytical methods

NH<sub>3</sub>/NH<sup> $\frac{1}{4}$ </sup> was detected by a spectrophotometer (DR 3900, Hach Co., Ltd.) using the EPA method 350.2, with a method detection limit (MDL) of 0.02 mg-N/L. NO<sup> $\frac{1}{3}$ </sup> and NO<sup> $\frac{1}{2}$ </sup> were measured by an ion chromatograph (IC2010, Tosoh Inc., Japan) equipped with a SuperIC-AZ column (Tosoh Inc.) and a 100 µL injection loop

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