



Abnormal pinocytosis and valence-variable behaviors of cerium suggested a cellular mechanism for plant yield reduction induced by environmental cerium[☆]



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ARTICLE INFO

Article history:

Received 3 February 2017

Received in revised form

9 July 2017

Accepted 11 July 2017

Keywords:

Rare earth elements

Cerium

Plant cell

Life cycle

Valence-variable behavior

ABSTRACT

The environmental safety of cerium (Ce) applications in many fields has been debated for almost a century because the cellular effects of environmental Ce on living organisms remain largely unclear. Here, using new, interdisciplinary methods, we surprisingly found that after Ce(III) treatment, Ce(III) was first recognized and anchored on the plasma membrane in leaf cells. Moreover, some trivalent Ce(III) was oxidized to tetravalent Ce(IV) in this organelle, which activated pinocytosis. Subsequently, more anchoring sites and stronger valence-variable behavior on the plasma membrane caused stronger pinocytosis to transport Ce(III and IV) into the leaf cells. Interestingly, a great deal of Ce was bound on the pinocytotic vesicle membrane; only a small amount of Ce was enclosed in the pinocytotic vesicles. Some pinocytotic vesicles in the cytoplasm were deformed and broken. Upon breaking, pinocytotic vesicles released Ce into the cytoplasm, and then these Ce particles self-assembled into nanospheres. The aforementioned special behaviors of Ce decreased the fluidity of the plasma membrane, inhibited the cellular growth of leaves, and finally, decreased plant yield. In summary, our findings directly show the special cellular behavior of Ce in plant cells, which may be the cellular basis of plant yield reduction induced by environmental Ce.

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1. Introduction

Environmental pollution of heavy metal elements and their effects in plants have attracted great attention from scientists worldwide, especially metal ions in trivalent or higher valences, due to their toxicity and susceptibility to the environment (Abdolali et al., 2015; Araújo and Cedeño-Macias, 2015; Guo and Yang, 2016;

[☆] This paper has been recommended for acceptance by Dr. Chen Da.

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Wu et al., 2016). Rare earth elements (REEs), specifically, fifteen lanthanides, are a set of heavy metal elements in trivalent or higher states (Redling, 2006). They have excellent magnetic susceptibility, radiant energy absorption, luminescence, chemical bonding behavior, and other advantages (Redling, 2006). Because of these special properties, REEs have been widely used in electronics, optics, communications, lasers, agriculture, medicine (such as phosphate binders, treatment of burns, diagnosis and treatment of cancer, and contrast agents), environmental protection, and many other fields (Redling, 2006; Stone, 2009; Tyler, 2004). These uses have led to over-accumulation of REEs in the environment (Biasioli et al., 2012; Gonzalez et al., 2014; Kulaksiz and Bau, 2011; Moreno et al., 2008; Wiseman et al., 2015). In soils worldwide, researchers estimated that the content of REEs has reached 1000 mg kg⁻¹

(Biasioli et al., 2012; Gonzalez et al., 2014). In the atmospheric particulate matter of Mexico and Tokyo, the REE content is approximately 10 ng m^{-3} (Moreno et al., 2008) and $44.1 \text{ } \mu\text{g m}^{-3}$, respectively. In the water of Rhine River rainwater, Germany, the REE content could be up to 610 ng kg^{-1} (Kulaksiz and Bau, 2011). Of all REEs, cerium (Ce) ($4f^{15}d^{16}s^2$) is widely used in industry, agriculture and medicine because it has two stable valence states [trivalence, Ce(III) and quadrivalence, Ce(IV)] in chemical reaction and relatively low toxicity in living organisms (Kostova et al., 2005; Redling, 2006). Therefore, Ce has become the most abundant REE in the environment and in living organisms (Biasioli et al., 2012; Kato et al., 2011; Kulaksiz and Bau, 2011; Moreno et al., 2008; Ni, 1995; Suzuki et al., 2011; Wiseman et al., 2015). It is noteworthy that excess Ce can cause detrimental effects on living organisms from plants to humans that include yield reduction for plants (Redling, 2006). Plants are the primary producer in ecosystems, and Ce is not essential for many physiological processes in plants. Therefore, to protect the ecosystem, it is urgent to elucidate the mechanism of Ce(III) actions in plants to establish a maximum permissible limit for REEs in plants as soon as possible.

The study on the toxic effect of Ce(III) in plants dates back to 1917 (Chien and Ostenhout, 1917). Many advances have been achieved in recent years (Guo and Huang, 2007; Redling, 2006; Schwabe et al., 2015). At the level of whole plant, the major symptom of Ce toxicity is global decreases in plant yield and growth (Guo and Huang, 2007; Redling, 2006; Schwabe et al., 2015). At the physiological level, inhibitory effects of Ce(III) on photosynthesis, respiration, mineral nutrient uptake and metabolism, and hormonal balance are often reported (Guo and Huang, 2007; Redling, 2006; Schwabe et al., 2015). Within cells, Ce toxicity is associated with oxidative stress, membrane damage and alteration of numerous enzymatic activities (Guo and Huang, 2007; Redling, 2006; Schwabe et al., 2015). Despite almost a century of research, the cellular mechanisms remain largely unknown (Guo and Huang, 2007; Redling, 2006; Schwabe et al., 2015) because investigators have known little about the life cycle and behaviors of Ce(III) in plant cells. By using electron microscopic autoradiography (EMARG), we previously observed that Ce(III) can enter the cells of the horseradish leaf (Guo and Huang, 2007). However, in the study of the behaviors of Ce(III) in plant cells, some key questions are still unclear: How does Ce(III) enter plant cells? How does Ce(III) remain in plant cells? Which valence state of Ce deposits in plant cells? How does Ce(III) regulate cellular responses?

In this study, we reported some special cellular behaviors of Ce(III) in plant cells by using interdisciplinary methods of EMARG, X-ray photoelectron spectroscopy (XPS), confocal laser scanning microscopy (CLSM), gas chromatography (GC), and scanning electron microscope (SEM), etc. Our study provides a new opportunity to probe into the mechanism of Ce(III) action in plants.

2. Materials and methods

2.1. Materials

The $^{141}\text{CeCl}_3$ solution was prepared by dissolving $^{141}\text{Ce}_2\text{O}_3$ powders (Beijing Atom High Tech Co. Ltd.) in the concentrated HCl and H_2O_2 . The nuclear emulsion was purchased from the Technical Institute of Physics and Chemistry, Chinese Academy of Sciences. CeCl_3 (purity > 99.99%) and other chemical reagents were purchased from Sigma-Aldrich China, Inc. (Shanghai, China).

2.2. Plant culture and treatment

Horseradish [*Armoracia rusticana* (Lam.) Gaerth.] is an important commercial crop, and it is a typical representative of edible

plants. It is widely used as a flavouring in Japan, Korea, Europe, America, etc. because it contains specific pungency (Veitch, 2004). In addition, horseradish is rich in horseradish peroxidase, which is an important industrial product (Veitch, 2004). Horseradish tubers were obtained from the Planting Base of Exported Horseradish in Dafeng City, Jiangsu Province, China, in which the Ce content in the airborne particulate matter of this base was 30 mg kg^{-1} , determined by inductively coupled plasma mass spectrometry (VG Plasma Quad Plus, Fisons Instruments, UK). Horseradish culture and treatment with Ce(III) were performed based on the previous reports (Wang et al., 2009, 2014). Horseradish tubers were planted in $30 \times 30 \text{ cm}^2$ pots without any fertilizer. The basic properties of the soil were as follows: sandy, pH 6.56, $15.71 \text{ me } 100 \text{ g}^{-1}$ cation exchange capacity, 0.81% soil organic matter, $0.031 \text{ } \mu\text{g g}^{-1}$ Ce. The horseradish plants were cultured in a glasshouse at the following condition: $25 \pm 1 \text{ }^\circ\text{C}$, 16 h light: 8 h dark cycle, $300 \text{ } \mu\text{mol m}^{-2} \text{ s}^{-1}$ light, and 70% relative humidity. At the four-leaf stage, some horseradish plants were evenly sprayed with 30, 80 or $300 \text{ } \mu\text{M}$ CeCl_3 solution until the drops began to fall, meanwhile, the soil was covered with plastic wrap and filter paper, which were called the CeCl_3 -treated plants; some horseradish plants were sprayed with the deionized water, which were called as the control plants. The amount of CeCl_3 solution sprayed on each horseradish plant was approximately 50 mL. Eight days after being sprayed, some CeCl_3 -treated and control plant leaves were collected for determining the net photosynthesis rate (P_n) [by a portable gas exchange system (CIRAS-1, PP Systems International Ltd, UK)], the chlorophyll content (Lichtenthaler, 1987), cell growth, fatty acids in the plasma membrane. The determination methods of the P_n and chlorophyll content were shown in the Supporting Information (SI).

2.3. Cell growth

To observe the size of the plant leaf cells, 100 mM Na-phosphate buffer (pH 7.0) containing 4% glutaraldehyde was used for fixing the fourth fully expanded rosette leaves of the plant. After fixation, the leaf samples were put into a series of concentrations of ethanol for dehydration. After infiltration, embedding, and cutting into sections, the plant leaf cells were observed under a microscope (Wang et al., 2014).

The protoplasts of horseradish leaves were isolated based on previously reported methods (Benoit et al., 2009; Wang et al., 2005). After fixation with 1% glutaraldehyde in 50 mM buffer solution (0.8 M mannitol and 50 mM Tris-HCl, pH 6.8) for 2 h at $4 \text{ }^\circ\text{C}$, the protoplasts were washed several times with the above-mentioned buffer solution. After critical point drying, the protoplasts were mounted over the stubs with double-sided conductivity tape, and a thin layer of gold metal was applied over the samples using an automated sputter coater (K500X, Quorum Technologies Ltd, United Kingdom). The samples were examined under the condition of low vacuum with a secondary electron detector at 15 kV using SEM (JSM-5600LV, JEOL, Japan) at various magnifications (Wang et al., 2010a).

2.4. EMARG observation of subcellular distribution of Ce(III)

Horseradish leaves were treated with 0, 30 or $80 \text{ } \mu\text{M}$ CeCl_3 solution that contained $^{141}\text{CeCl}_3$, respectively, and the radioactivities of these solutions were all $75 \text{ } \mu\text{Ci mL}^{-1}$. After treating with $^{141}\text{Ce(III)}$ -containing Ce(III) solution for 12 h, 48 h and 8 d, the leaves were sampled and then cut into regular ultra-thin ($\sim 60 \text{ nm}$ thickness) sections of $1.5 \times 2 \text{ mm}$ size by using Reichert Ultracut E ultramicrotome. These ultra-thin sections were painted with nuclear emulsion, and these sections were subsequently developed and fixed. The subcellular distribution of $^{141}\text{Ce(III)}$ was subsequently

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