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Source apportionment of perfluoroalkyl substances in surface sediments from lakes in Jiangsu Province, China: Comparison of three receptor models

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ABSTRACT

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Introdu

48 Introduction

The mass production and widespread usage of perfluoroalkyl substances (PFASs) have begun since the late 1940s (Kim et al., 2012). PFASs are ubiquitous in various environmental media, such as water, soils, sediments, sewage sludge, biota and human bodies (Wang et al., 2015; Zhao et al., 2012). It was inferred that a large proportion of PFASs would be released to 55 the surface waters, and sediments, as the natural environ- 56 ment of benthic organisms, are considered as one of the most 57 important environmental sinks of PFASs (Prevedouros et al., 58 2006). PFASs have been found in sediments from several 59 countries and regions, and PFAS concentrations have reached 60 up to 800 ng/g dry weight (dw) (Ahrens et al., 2015; Campo 61

Receptor models have been proved as useful tools to identify source categories and 18

quantitatively calculate the contributions of extracted sources. In this study, sixty surface 19

sediment samples were collected from fourteen lakes in Jiangsu Province, China. The total 20

concentrations of C_4 - C_{14} -perfluoroalkyl carboxylic acids and perfluoroactane sulfonic acid 21

 $(\Sigma_{12}$ PFASs) in sediments ranged from 0.264 to 4.44 ng/g dw (dry weight), with an average of 22

1.76 ng/g dw. Three commonly-applied receptor models, namely principal component 23

analysis-multiple linear regression (PCA-MLR), positive matrix factorization (PMF) and 24

Unmix models, were employed to apportion PFAS sources in sediments. Overall, these three 25

models all could well track the \sum_{12} PFASs concentrations as well as the concentrations 26

explained in sediments. These three models identified consistently four PFAS sources: the 27

textile treatment sources, the fluoropolymer processing aid/fluororesin coating sources, the 28

textile treatment/metal plating sources and the precious metal sources, contributing 28.1%, 29

37.0%, 29.7% and 5.3% by PCA-MLR model, 30.60%, 39.3%, 22.4% and 7.7% by PMF model, and 30 20.6%, 52.4%, 20.2% and 6.8% by Unmix model to the \sum_{12} PFASs, respectively. Comparative 31 statistics of multiple analytical methods could minimize individual-method weaknesses 32 and provide convergent results to enhance the persuasiveness of the conclusions. The 33

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findings could give us a better knowledge of PFAS sources in aquatic environments.

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et al., 2015; Zhou et al., 2013). Source apportionment of PFASs
in sediments is of great significance for pollution control and
ecological protection of aquatic environments.

Qualitative and semiquantitative methods have been 65 widely employed to identify PFAS sources. Xiao et al. (2012) 66 adopted cluster analysis, correlation analysis, ANOVA and per 67 capita discharge to distinguish different PFAS patterns in 68 69 influent samples from thirty-seven wastewater treatment 70 plants in a multi-city survey. Murakami et al. (2009) used ratio 71 methods to evaluate PFAS contributions from street runoff and wastewater to the aquatic environments. A few methods, 72 e.g., principal component analysis (PCA), multiple linear 73 regression model (MLR), positive matrix factorization (PMF) 74 and Unmix models, have been applied to quantitative source 75apportionment of PFASs. However, only one or two of these 76 77 models are usually applied to source apportionment of PFASs, and it is not enough to provide comprehensive information 78 for PFAS sources (Kuroda et al., 2014; Qi et al., 2016). 79

Jiangsu Province, located in the lower reaches of the Yangtze 80 River and Huaihe River, is traversed by the Beijing-Hangzhou 81 Grand Canal from north to south. It forms part of the Yangtze 82 River Delta urban agglomeration, which is one of the six 83 world-class city clusters. As one of the most intensively 84 85 industrialized provinces in China, Jiangsu Province contributed 86 the largest portion of perfluorooctane sulfonic acid (PFOS) 87 emissions in China (Xie et al., 2013). PFOS concentrations in 88 Taihu Lake in Jiangsu Province even exceeded US EPA standards 89 of 0.2 µg/L for PFOS in drinking water (US EPA, 2009; Yang et al., 2011). Our published work showed that lake sediments from 90 Jiangsu Province have also suffered the heaviest PFAS pollution 91(mean 1.76 ng/g dw) among all provinces studied (Qi et al., 922016). In this study, PCA-MLR, PMF and Unmix models were 93 applied to identify source categories and quantitatively calcu-9495late source contributions of PFASs in lake sediments from Jiangsu Province. Results from three models were evaluated and 96 compared in order to improve source apportionment of PFASs. 97 Comparative statistics could enhance the persuasiveness of the 98 conclusions and offer us a better knowledge of PFAS sources in 99 lake sediments. 100

102 1. Materials and methods

103 **1.1. Standards and reagents**

Seventeen PFASs including C_4 - C_{14} -perfluoroalkyl carboxylic 104 acids (PFCAs) and C4, C6, C8 and C10-perfluoroalkane sulfonic 105acids were targeted. A mixture of seventeen native PFASs 106and nine stable isotope-labeled surrogate internal standards 107 in $2 \mu g/mL$ solution mixtures were both purchased from 108 Wellington Laboratories (Guelph, ON, Canada). Perfluoro-1-109[1,2,3,4,5,6,7,8⁻¹³C₈]octanesulfonate and perfluoro-n-[1,2,3,4,5, 110 6,7,8-¹³C₈]octanoic acid (50 μg/mL, 99%) were purchased from 111 Cambridge Isotope Laboratories (Andover, MA, USA). Detail 112 113 information is shown in Appendix A.

114 **1.2. Sample collection**

115 The sampling campaign was carried out in 2013. In total, sixty

116 surface sediment samples from fourteen lakes of China were

analyzed (Fig. 1). Geographic and limnological features of 117 fourteen lakes are listed in Appendix A Table S1. Details on 118 sample collection are displayed in Appendix A. 119

1.3. Sample extraction and instrumental analysis

The samples were pretreated as previously published proce-121 dures with minor modifications and optimizations, using 122 methanol extraction and cleanup by Oasis WAX-SPE cartridges 123 (Zhou et al., 2013). PFAS analysis was accomplished using a 124 ultra-high performance liquid chromatography coupled to a 125 negative electrospray ionization tandem mass spectrometer 126 (UPLC-ESI-MS/MS, Xevo TQD, Waters Corp., Milford, MA, USA) 127 operated in the quantitative multiple reaction monitoring 128 mode. Detail information on sample extraction, instrument 129 analysis and quality assurance/quality control is shown in 130 Appendix A. 131

1.4. Statistical analysis

PFAS concentrations are reported on a dry weight basis. 133 Twelve PFASs, including C₄–C₁₄-PFCAs and PFOS, were further 134 analyzed, with high detection frequencies of 87%-100%. At 135 least 93% of concentrations were quantifiable for each PFAS. 136 In further analysis, all concentrations lower than the limits of 137 detection (LODs) and limits of quantification (LOQs) were 138 reported as half of the LODs and LOQs, respectively. The total 139 concentrations of C₄-C₁₄-PFCAs and PFOS were represented 140 by the \sum_{12} PFASs. Raw data were obtained from the MassLynx 141 V4.1 workstation (Waters Corp., Milford, MA, USA) and 142 processed in the Microsoft Office 2010 software (Microsoft 143 Inc., Redmond, WA, USA). Pearson correlation analysis was 144 conducted with the SPSS 22.0 software (SPSS Inc., Chicago, IL, 145 USA). Data for Pearson correlation analysis conform to a 146 normal distribution. PCA-MLR, PMF and Unmix models were 147 run with the Matlab R2014b software (MathWorks Incor., 148 Natick, MA, USA), US EPA PMF 5.0 and Unmix 6.0 software 149 packages, respectively. 150

1.5. Receptor models

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PCA-MLR, PMF and Unmix models are three multivariate 152 factor analysis receptor models, and none of them need 153 source categories of pollutants in advance. In source appoint- 154 ment, it was assumed that PFAS compositions of each source 155 did not change as they moved from source to receptor and did 156 not react with each other, as described in the previous study, 157 and so forth (Watson, 1984). Generally, they can be described 158 by the following Eq. (1): 159

$$c_{ij} = \sum_{k=1}^{n} \mathbf{x}_{in} f_{nj} + \mathbf{e}_{ij}$$
⁽¹⁾

where, c_{ij} is the concentration of ith species for the *j*th sample; 160 x_{in} is the ith species concentration from the *n*th source; f_{nj} is 162 the contribution of the *n*th source to the *j*th sample; e_{ij} is the 163 error or uncertainty (Yang et al., 2013). Details regarding these 164 three receptor models were given in their respective user 165 manuals.

PCA-MLR model explains data variables by fewer independent $\,167$ factors. Prior to statistical analysis, all data were transformed $\,168$

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