IOURNAL OF ENVIRONMENTAL SCIENCES XX (2016) XXX-XXX



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Removal of ammonium ion from water by Na-rich birnessite: Performance and mechanisms

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ARTICLE INFO 8

- 10 Article history:
- Received 8 July 2016 11
- 12Revised 25 November 2016
- 13 Accepted 8 December 2016
- Available online xxxx 14
- Keywords: 32
- 33 Ammonium ion adsorption
- 34 Na-rich birnessite
- 35 Hydroxyl ions
- Negatively charged surface 36

- Electrostatic interaction 37
- 38

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ABSTRACT

Na-rich birnessite (NRB) was synthesized by a simple synthesis method and used as a 15 high-efficiency adsorbent for the removal of ammonium ion (NH4) from aqueous solution. 16 In order to demonstrate the adsorption performance of the synthesized material, the effects 17 of contact time, pH, initial ammonium ion concentration, and temperature were 18 investigated. Adsorption kinetics showed that the adsorption behavior followed the pseudo 19 second-order kinetic model. The equilibrium adsorption data were fitted to Langmuir and 20 Freundlich adsorption models and the model parameters were evaluated. The monolayer 21 adsorption capacity of the adsorbent, as obtained from the Langmuir isotherm, was 22 22.61 mg NH⁺₄-N/g at 283 K. Thermodynamic analyses showed that the adsorption was 23 spontaneous and that it was also a physisorption process. Our data revealed that the higher 24 NH[‡] adsorption capacity could be primarily attributed to the water absorption process and 25 electrostatic interaction. Particularly, the high surface hydroxyl-content of NRB enables 26 strong interactions with ammonium ion. The results obtained in this study illustrate that 27 the NRB is expected to be an effective and economically viable adsorbent for ammonium 28 ion removal from aqueous system. 29

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Introduction 43

Ammonium ion (NH⁺₄) is the primary nitrogen polluting species 4546 in the hydrosphere. Ammonium ion discharged together with 47 municipal, agricultural or industrial effluents is responsible for 48 harmful effects such as eutrophication of lakes and rivers, 49 decreased dissolved oxygen, toxicity to aquatic life, and increased corrosion rate of soil materials (Karadag et al., 2006; 50Yusof et al., 2010). Eutrophication of water bodies, a problem of 51global concern, has been identified as a major environmental 52problem for water resource management (Du et al., 2005; Zheng 53 et al., 2008). In water supply plants, the ammonium ion present 54in raw water will react with chlorine to form disinfection 55byproducts, which could damage the human nervous system 56

and cause deterioration of the taste and odor of water. Hence, it 57 is necessary either to control the ammonium ion contamina- 58 tion in raw water or to remove it before the water is disinfected 59 with chlorine (Cai et al., 2015; Han et al., 2013). Controlling the 60 ammonium ion in the raw water is the preferred solution. 61

A number of methods (e.g., ion-exchange, adsorption) have 62 been developed to remove ammonium ion (Leaković et al., 2000; 63 Liu et al., 2015; Mandowara and Bhattacharya, 2011; McVeigh 64 and Weatherley, 1999; Wang et al., 2014). Among these 65 strategies, adsorption is considered to be promising due to its 66 high-efficiency, cost effectiveness and easy implementation 67 (Zhu et al., 2012). As a result, many natural and synthetic 68 materials have been developed and used as adsorbents for 69 ammonium ion removal. However, these adsorbents were 70

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http://dx.doi.org/10.1016/j.jes.2016.11.015

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Please cite this article as: Cheng, Y., et al., Removal of ammonium ion from water by Na-rich birnessite: Performance and mechanisms, J. Environ. Sci. (2016), http://dx.doi.org/10.1016/j.jes.2016.11.015

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always found to have slow adsorption kinetics, and their 71 adsorption capacities may be limited (Alshameri et al., 2014; 72Kučić et al., 2012; Moradi, 2011; Nielsen, 1996; Rožić et al., 2000). 73 In addition, increased awareness and understanding of the 74 deleterious effects of ammonium ion released from wastewater 75 treatment facilities into natural water systems has resulted in 76 77 stringent laws restricting ammonium ion discharge. Therefore, the development of new adsorbents designed specifically for 78 79 the rapid removal of ammonium ion from water is necessary.

80 Manganese oxides (MnOx) with layered topologies are ubiquitous in the natural environment. Birnessite belongs to 81 the family of layered hydrous manganese oxides. It possesses 82 unique surface charge, cation exchange, and redox properties, 83 which make it highly reactive with respect to sorption phenom-84 ena. Furthermore, birnessite can be synthesized easily under 85 laboratory conditions and, consequently, has often been applied 86 to remove heavy metals (e.g., copper, cadmium, lead and zinc) 87 from water (Eren, 2009; Gadde and Laitinen, 1974; Han et al., 88 2006a, 2006b; Lefkowitz and Elzinga, 2015; Wang et al., 2007; 89 Kanungo et al., 2004; McKenzie, 1980; Villalobos et al., 2005). 90

To the best of our knowledge, studies on the adsorption of 91 ammonium ion by manganese oxide in water solution have 92 been limited. Birnessite is the predominant naturally occur-93 94 ring manganese oxide in most environmental settings (Zhu et al., 2010). Thus, in this study, we examined the adsorption 95 96 of ammonium ion from water by a birnessite-type manganese 97 oxide. The adsorption kinetics and equilibrium isotherms of 98 the uptake of ammonium ion on manganese oxide were studied. Additionally, the effects of initial ammonium ion 99 concentration, contact time, pH, and the major ions present in 100 groundwater on the adsorption process were investigated. 101 The structural characterization of the adsorbents was used to 102preliminarily reveal the underlying sorption mechanism. 103

104 1. Materials and methods

106 **1.1. Materials**

Manganese chloride (MnCl₂·4H₂O), potassium permanganate
(KMnO₄), sodium hydroxide (NaOH), ammonium chloride
(NH₄Cl), calcium chloride (CaCl₂), and magnesium chloride
(MgCl₂) were of analytical reagent grade. All solutions were
prepared with deionized distilled water. NH⁺₄-N solutions of
different concentrations were prepared by dissolving the
required amounts of NH₄Cl in deionized distilled water.

114 **1.2. Preparation of absorbent**

Birnessite-type manganese oxides with interlayer K⁺ and Na⁺ 115116 ions have been synthesized by redox reactions between KMnO₄ and MnCl₂ under alkaline conditions. In this study, 117 the absorbent sample was prepared based on the method 118 reported by Luo and Suib (1997). Specifically, 160 mL of 119 120 0.5 mol/L MnCl₂·4H₂O was added into 180 mL of 5.55 mol/L NaOH under constant stirring. After 2 hr of cooling, 160 mL of 1211220.2 mol/L KMnO₄ was added dropwise under constant stirring. The manganese oxide formed was allowed to settle and the 123 supernatant was replaced with water several times until no 124 more manganese ion was detected in the washing water. The 125

resultant sample was freeze-dried, ground, and sieved 126 through a 100 mesh sieve before being used.

Birnessite is a layered Mn oxide consisting of stacked 128 sheets of edge-sharing MnO_6 octahedra with water molecules 129 in the interlayer space (Elzinga, 2011; Drits et al., 1998; 130 Nakayama et al., 2008). Furthermore, due to the Mn site 131 vacancies and substitution of Mn(III) for Mn(IV), the manganese oxide layer is always negatively charged (Elzinga, 2011), 133 and thus in order to counterbalance the charge defects in 134 MnO_6 sheets, cations (*e.g.*, Na⁺) will also be intercalated in the 135 interlayer region. 136

1.3. Characterization of adsorbent 137

The manganese oxide sample solids were characterized by 138 attenuated total reflectance Fourier transform infrared (ATR- 139 FT-IR) spectroscopy, powder X-ray diffraction (XRD), scanning 140 electron microscopy (SEM) and nitrogen porosimetry to deter- 141 mine MnO_x mineralogy. Samples were randomly selected for 142 characterization. To resolve the bands of functional groups and 143 determine their frequency, spectral analysis was done for MnO_x 144 sample solids before and after NH⁺₄ adsorption using an FT-IR 145 spectrometer (Thermo-Nicolet IS50) in the range of 500-146 4000 cm⁻¹. The structures of the samples were examined by 147 XRD at room temperature using a Rigaku Ultiman IV diffrac- 148 tometer with Cu K α radiation (λ = 1.5606 Å) operated at 40 kV 149 and 40 mA, with data collected in the range of 5 to 80° at a scan 150 rate of 5°/min and a step size of 0.02°. Phases were identified 151 using JCPDS files. The surface and pore properties of the 152 adsorbent were studied using nitrogen adsorption experiments 153 using an Autosorb-1 (Quantachrome Instruments) at 77 K. The 154 specific surface area of the adsorbent was calculated using the 155Brunauer-Emmett-Teller (BET) method. SEM images were 156 taken on a Hitachi S-3400 N scanning electron microscope. 157 The voltage used was 5 kV for the SEM studies. Zeta potential 158 was conducted by a Malvern Instruments Zetasizer Nano ZS 90. 159

1.4. Batch adsorption experiments

Two different types of sorption experiments were performed: 161 kinetics experiments, with reaction times ranging from 0 to 162 300 min, and isotherm experiments at 283, 293, 303 and 313 K 163 to establish the maximum ammonium ion sorption after 164 30 min of equilibration, where the concentrations of NH₄⁺-N 165 ranged from 2 to 50 mg/L. For both types of experiments, 166 portions of the adsorbent samples were accurately weighed 167 into 250-mL glass conical flasks to yield a final solid 168 concentration of 0.5 g/L. The final volumes were adjusted to 169 100 mL with deionized distilled water after appropriate 170 volumes of ammonium ion stock solution were added to 171 yield the pre-selected NH⁺₄-N concentrations. The initial 172 concentrations of NH₄⁺-N in the kinetics experiments were 2- 173 25 mg/L. The suspensions were shaken at 150 r/min on a 174 water bath shaker at 298 K. 175

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To explain the adsorption behavior of manganese oxide, 176 the effect of pH on the NH_4^+ adsorption capacity of manganese 177 oxide was evaluated. The pH values were adjusted using 178 dilute NaOH or HCl. The initial concentrations of NH_4^+ -N and 179 manganese oxide used in these experiments were 10 mg/L 180 and 0.5 g/L, respectively. To study the effect of cations (Na⁺, 181

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