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Removal of Cs⁺ from water and soil by ammonium-pillared montmorillonite/Fe₃O₄ composite

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ABSTRACT

To remove cesium ions from water and soil, a novel adsorbent was synthesized by following a one-step co-precipitation method and using non-toxic raw materials. By combining ammonium-pillared montmorillonite (MMT) and magnetic nanoparticles (Fe₃O₄), an MMT/ Fe₃O₄ composite was prepared and characterized. The adsorbent exhibited high selectivity of Cs⁺ and could be rapidly separated from the mixed solution under an external magnetic field. Above all, the adsorbent had high removal efficiency in cesium-contaminated samples (water and soil) and also showed good recycling performance, indicating that the MMT/Fe₃O₄ composite could be widely applied to the remediation of cesium-contaminated environments. It was observed that the pH, solid/liquid ratio and initial concentration affected adsorption capacity. In the presence of coexisting ions, the adsorption capacity decreased in the order of Ca²⁺ > Mg²⁺ > K⁺ > Na⁺, which is consistent with our theoretical prediction. The adsorption behavior of this new adsorbent could be expressed by the pseudo-second-order model and Freundlich isotherm. In addition, the adsorption mechanism of Cs⁺ was NH⁴₄ ion exchange and surface hydroxyl group coordination, with the former being more predominant.

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49 Introduction

In order to solve the problems of energy demand and 44 45environmental pollution, many countries are seeking new 46 clean energy alternatives to fossil fuels. Nuclear power, which 47 has provided approximately 18.93% of the world's electricity 48 (International Atomic Energy Agency, 2015), is one of the most 49 promising methods. However, nuclear power plants emit approximately 200 to 300 tons of radioactive waste globally 50each year (S. Yang et al., 2014; Yang et al., 2014a, b). In Q5 Q4 addition, when nuclear power accidents occur, radioactive 52pollutants leak into the environment. For example, in Japan's 532011 Fukushima nuclear accident, approximately 630,000-54

770,000 TB radionuclides were released into the environment, Q6 including thousands of tons of radioactive waste (S. Yang et 56 al., 2014; Yang et al., 2014a, b; Yasunari et al., 2011), which 57 threatened human health and the ecosystems. Meanwhile, 58 the Chinese government decided to build 6 to 8 nuclear power 59 plants each year starting in 2016, which drew more attention 60 to radioactive wastewater (Zhang et al., 2015). Radioactive 61 isotopes of cesium (134 Cs and 137 Cs) are the primary concerns 62 among radioactive wastes, because of their abundant content 63 and a long half-life (with 2.06 and 30 years, respectively) 64 (Zhang et al., 2015). At present, there are many ways to 65 decontaminate radionuclide from wastewater. Precipitation, 66 solvent extraction, membrane dialysis, and adsorption are the 67

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most commonly used methods (Long et al., 2013; Wu et al.,
2009; Yang et al., 2016). Among these methods, adsorption is a
simple operation, economical and feasible.

Clay minerals such as bentonite, illite, kaolinite, montmo-71 rillonites, sepiolite, and zeolite which are natural and low cost 72adsorption materials have been widely used for the removal 73 74 of cesium ions from the solution (Bayülken et al., 2011; Nakano et al., 2003; Poinssot et al., 1999). Montmorillonite 7576 (MMT) belongs to the 2:1 clay family and its layered structure consists of two silica tetrahedral sheets fused to one alumina 77 octahedral sheet. The isomorphous substitution of Al³⁺ for 78 Si^{4+} in the tetrahedral layer and Mg^{2+} or Fe^{2+} for Al^{3+} in the 79 octahedral layer results in a negative surface charge on the 80 clay, and these negative sites can adsorb cations such as Ca²⁺, 81 K⁺ and Na⁺ between the layers through electrostatic attraction 82 (Wu et al., 2009; Yang et al., 2016). Compared with other clay 83 minerals, MMT has a higher adsorption capacity of cesium 84 because of its high cation exchange capacity (CEC), large 85 surface area, stable chemical properties and large reserves in 86 China (J. Wang et al., 2010; T. Wang et al., 2010; Westrich et al., 87 1995). 07

Natural MMT primarily occurs in the form of Ca-MMT or 89 Na-MMT, which has low selective adsorption to Cs⁺. MMT 90 91 pillared with ethylamine (Karamanis and Assimakopoulos, 2007; Long et al., 2013) and aluminum (Karamanis and 92 93 Assimakopoulos, 2007) has been investigated, but the perfor-94 mance of selective adsorption of Cs⁺ was not high. The hydration 95radius of NH⁺₄ is similar to that of Cs⁺. Ca-MMT pillared by NH⁺₄ would preferentially exchange Cs⁺ under other coexisting ions. 96 97 In addition, MMT is usually powder-suspended in the solution, which is difficult to be separated and reused after Cs⁺ sorption. 98 Magnetic separation, which has been used to remove cesium 99 ions from the solution, is convenient to separate the adsorbent 100 from wastewater (Cao et al., 2008; Fan et al., 2011; Hashemian et 101 al., 2015; Oliveira et al., 2004; Pan et al., 2011; Sheha, 2012; 102Yantasee et al., 2007). MMT/Fe₃O₄ has been used to remove 103 heavy metals (Cu²⁺, Pb²⁺, and Ni²⁺) from aqueous solutions 104(Kalantari et al., 2015). However, the capture of Cs⁺ in water and 105soil systems with the use of ammonium-pillared MMT/Fe₃O₄ 106 magnetic composites, and its adsorption mechanism have not 107 been investigated. 108

In this study, ammonium-pillared MMT/Fe₃O₄ was synthe-109 110 sized by the in situ co-precipitation technique. The sorption of Cs⁺ on MMT, Fe₃O₄ and MMT/Fe₃O₄ composite was measured, 111 and the adsorption equilibrium, kinetics, thermodynamics 112and their applicability in real samples were also discussed. In 113addition, the adsorption mechanism was investigated by 114 means of batch experiments, adsorption model and charac-115terization techniques. 116

118 1. Experimental section

119 **1.1. Materials**

The chemicals $FeCl_2$ ·4H₂O, $FeCl_3$ ·6H₂O, NaCl, KCl, MgCl₂, and Q8 CaCl₂ of A.R. grade were obtained from Sinopharm Chemical Reagent Co., Ltd., China. Ethanol, NH₃·H₂O (25 wt.%) and hydrochloric acid of A.R. grade were purchased from Beijing Beihua Fine Chemicals Co., Ltd., China. Nonradioactive 124 cesium chloride (CsCl) of extra-pure grade was purchased 125 from Amresco, USA. Pure water generated by a Synergy UV Q9 ultra-pure water system (Millipore, USA) was used to prepare 127 the solution and synthesize the novel adsorbent. 128

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1.2. Preparation of MMT/Fe₃O₄ composite

Fe₃O₄ was synthesized by the co-precipitation method as 130 reported previously (Jolivet et al., 2004) with a slight modifi- 131 cation. The synthetic process of the MMT/Fe₃O₄ magnetic 132 particles was similar to that of Fe₃O₄. The composite 133 mechanism is in situ aggregation (Zhao et al., 2011). First, 134 3.1736 g of FeCl₂· $4H_2O$ (0.016 mol) was dissolved in 300 mL of 135 deionized water; 3.00 g of MMT was added into the mixture to 136 ultrasonic bath for 15 min, and the mixture was stirred for 137 30 min. Next, 7.5684 g of FeCl₃·6H₂O (0.028 mol) was dis- 138 solved in the suspension by stirring the mixture solution for 139 30 min at 80°C under nitrogen atmosphere. Then, 40 mL of 140 NH₃·H₂O was injected into the mixture rapidly, stirred under 141 nitrogen for another hour, and cooled at room temperature. 142 The obtained black precipitates were magnetically separated 143 and washed repeatedly with distilled water and ethanol. 144 Finally, the MMT/Fe₃O₄ magnetic fluid underwent vacuum 145 freeze-drying for 12 hr. 146

1.3. Characterization

Fourier transform infrared (FT-IR) spectrum was recorded Q10 with KBr over the spectral range from 4000 to 400/cm at a 149 resolution of 4/cm (Affinity-1, Shimadzu, Japan). TGA curves Q11 were measured using a TG-DSC instrument from 25 to 800°C 151 under N₂ at a heating rate of 20°C/min (TGA/DSC1, 152 Mettler-Toledo, Switzerland). Powder X-ray diffraction (XRD) 153 patterns were recorded on a powder XRD (X Pert Pro MPD, 154 Panalytical, Holland) with Cu K α radiation over the 20 range of 155 5°-80° at a scanning rate of 0.033°/sec. X-ray photoelectron 156 spectrometry (XPS) was performed on a Scanning X-ray 157 Microprobe (ESCSLAB 250Xi, ThermoFisher, USA), and the 158 binding energy (BE) was calibrated using the C (1s) peak at 159 284.6 eV. The surface micromorphology was analyzed by 160 scanning electron microscopy (SEM) equipped with energy 161 dispersive spectroscopy (S-4800/EX-350, Hitachi, Japan). 162

1.4. Adsorption experiments

The adsorption behaviors of inactive cesium (cesium-133) were 164 determined by using a batch mode of operations. A certain 165 amount of sorbents were ground into the powdered material and 166 then dispersed in the aqueous cesium chloride solution. After the 167 predetermined time of shaking at 200 r/min on a rotatory shaker, 168 the adsorbent suspended in aqueous solution was separated 169 with a 0.22 μ m cellulose acetate filter membrane. The residual 170 concentrations of metal ions remaining in the solution were 171 measured using an atomic absorption spectrophotometer 172 (ICE3500, ThermoFisher, USA), and the concentration of ammo-173 nium ion was determined by UV spectrophotometer (UV-2600, 174 Shimadzu, USA).

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