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Achieving partial nitrification by inhibiting the activity of Nitrospira-like bacteria under high-DO conditions in an intermittent aeration reactor

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ABSTRACT

It is generally accepted that a low dissolved oxygen (DO) concentration is more beneficial 16 for achieving partial nitrification than high-DO. In this study, partial nitrification was not 17 established under low-DO conditions in an intermittent aeration reactor for treating 18 domestic wastewater. During the operational period of low-DO conditions (DO: 0.3 ± 19 0.14 mg/L), stable complete nitrification was observed. The abundance of Nitrospira-like 20 bacteria, which were the major nitrite-oxidizing bacteria, increased from 1.03×10^6 to 21 2.64 × 10⁶ cells/mL. At the end of the low-DO period, the batch tests showed that high-DO 22 concentration (1.5, 2.0 mg/L) could inhibit nitrite oxidation, and enhance ammonia 23 oxidation. After switching to the high-DO period (1.8 ± 0.32 mg/L), partial nitrification was 24 gradually achieved. Nitrospira decreased from 2.64×10^6 to 8.85×10^5 cells/mL. It was found 25 that suddenly switching to a high-DO condition could inhibit the activity and abundance of 26 Nitrospira-like bacteria, resulting in partial nitrification. 27

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40 Introduction

Partial nitrification, which avoids the oxidation of nitrite to 42 43nitrate, is a critical step in some biological nutrient removal 44 processes (e.g. SHARON-ANAMMOX, CANON, OLAND) (Van 45 Hulle et al., 2010; Regmi et al., 2014). Compared to traditional 46 nitrification-denitrification nitrogen removal processes, pro-47 cesses with partial nitrification are less expensive since they consume less oxygen and less carbon-source (Ahn, 2006). The 48 establishment of partial nitrification is always the key to the 49implementation of these processes, but it is difficult to 50achieve. At present, several effective methods have been 51developed to achieve partial nitrification, which are mainly 52based on high temperature (Ge et al., 2015), high free 53ammonia concentrations (FA) (Zhang et al., 2014), high free 54

nitrous acid (FNA) (Ge et al., 2015), aeration duration control 55 (Ma et al., 2016), inhibitors and low dissolved oxygen (DO) 56 concentration (Peng and Zhu, 2006). 57

Among these factors, low-DO concentration has been 58 widely reported to successfully achieve partial nitrification 59 in wastewater treatment systems (Peng and Zhu, 2006; Ma 60 et al., 2009, 2011). It is commonly recognized that the activity 61 of ammonium-oxidizing bacteria (AOB) is higher than that of 62 nitrite-oxidizing bacteria (NOB) under low-DO conditions. 63 This is because the oxygen half-saturation constant (K_o) 64 of AOB is lower (Hanaki et al., 1990; Manser et al., 2005; 65 Stenstrom, 1990), resulting in NOB being gradually selected Q3 out due to the stronger inhibition by low-DO conditions. 67 However, Liu and Wang (2013) recently found their reactors 68 performed complete nitrification during long-term operation 69

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under low-DO conditions. Similarly, Park and Nogura (2004) 70 suggested that complete and stable nitrification could be 71 maintained even at a low-DO level (0.12 mg/L). Moreover, 72Regmi et al. (2014) found that the K_o of AOB (1.16 mg-O₂/L) was 73 significantly higher than that of NOB (0.16 mg-O₂/L) in their 74 reactor. These contradictory findings suggest that there is a 75 high probability of not achieving partial nitrification in some 76 low-DO reactors, which can adversely affect processes based 77 78 on partial nitrification. However, there has been limited 79 understanding of how to start-up or recover partial nitrification in a low-DO reactor performing complete nitrification. 80

In our study, an intriguing phenomenon was observed in an 81 intermittent aeration nitrifying reactor treating organic-removed 82 domestic wastewater. Complete nitrification was maintained 83 under the low-DO concentration, but partial nitrification 84 was started up under the high-DO concentration. In order to 85 determine the mechanism of the start-up of partial nitrification, 86 the nitrogen removal performance of the sequencing batch 87 reactor (SBR) was monitored, and the activity and community of 88 nitrifying bacteria were detected by batch experiments and 89 biological methods. 90

1. Materials and methods 92

1.1. Experimental bioreactor set-up and operating conditions 93

94 An SBR with a working volume of 10 L was used which included 95 two operational stages: low-DO stage (DO: 0.3 ± 0.14 mg/L) and high-DO stage (DO: 1.8 ± 0.32 mg/L). The SBR was operated at a 96 97 fixed temperature of 25°C using an electric heater. The oxygen required for the nitrification process was supplied by an air pump 98 99 through a flow meter. A mechanical stirrer was used to ensure a completely mixed status during the nitrification process. The pH, 100 DO and temperature in the reactor were monitored using an 101 on-line multi-parameter sensor (WTW 3420, Germany). The pH 102 was not controlled in the reactor. The operational stage in each 103cycle included four periods, namely 10 min feeding period, 104 variable-length intermittent aeration (10 min aeration/10 min 105anoxic) period, 30 min settling period, and 15 min discharging of 106 5 L supernatant period. The terminal point of nitrification was 107 judged by the real-time control of DO concentration. The 108 endpoint values were 0.7 mg/L in the low-DO stage and 3 mg/L 109in the high-DO stage. The SRT was maintained at 30 days by 110 withdrawing the activated sludge regularly. The mixed liquor 111 suspended solids (MLSS) was maintained at 1500 ± 134 mg/L, 112and the effluent suspended solids (SS) was 24 ± 8 mg/L. 113

The influent to the nitrifying SBR was domestic wastewater 114from which the organics had been removed in an activated 115sludge bioreactor by direct aeration for a short time (35 min). 116 117 The characteristics of the influent were as follows: COD: 40-55 mg/L, NH₄⁺-N: 50.2-80.4 mg/L, NO₂⁻-N: 0.1-5.4 mg/L, NO₃⁻-N 118

0.04-1.24 mg/L, and pH 7.2-7.8. Activated sludge taken from a 119 pilot-scale SBR (volume: 7000 L) was used as the inoculum. 120

1.2. Batch experiments 121

Batch experiments were carried out to investigate the effect of DO 122 fluctuation on the nitrifying performance of the activated sludge 123 acclimated to the stable DO concentration. The experiments were 124 conducted in a lab-scale SBR with a working volume of 2.5 L at 125 the end of the low-DO stage and the high-DO stage, separately. 126 The sludge was taken from the operating SBR (10 L) and washed 127 three times with deionized water before being transferred to the 128 2.5 L SBR. The synthetic wastewater was prepared from a 129 solution of NH₄Cl and NaNO₂. The initial NH₄⁺-N and NO₂⁻-N 130 concentrations were 40 mg/L and 20 mg/L, respectively. Differ- 131 ent DO concentrations (0.3, 0.5, 1.0, 1.5 and 2.0 mg/L) were 132 applied in the batch experiment. A variable frequency air pump 133 was used to adjust air flow automatically according to the DO 134 probes, in order to ensure a stable DO concentration. The initial 135 MLSS in each 2.5 L SBR was kept at 2000 mg/L. The temperature 136 was maintained at 24°C and the pH was kept at 7.2 by adding 137 NaHCO3 solution. The batch experiments were operated in 138 triplicate for each DO concentration. 139

1.3. Analytical methods 140

 $NH_{4}^{+}-N$, $NO_{2}^{-}-N$ and $NO_{3}^{-}-N$ concentrations were measured using a 141 Quickchem® 8500 system (Hach Company, Lachat Instruments, 142 USA), while the chemical oxygen demand (COD), SS and MLSS, 143 MLVSS were analyzed using standard methods (American Public 144 Health Association, 1998). 145

1.4. DNA extraction and quantitative PCR assay

Total DNA was extracted from 0.1 g of lyophilized sludge 147 using a Fast DNA Spin Kit for Soil (QBIOgen Inc., Carlsbad, CA, 148 USA), following the manufacturer's instructions. The DNA 149 samples were stored at -20°C. The concentration of DNA was 150 measured using a Nanodrop ND-1000 UV-Vis Spectrophotometer 151

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(NanoDrop Technologies, Wilmington, DE, USA). Quantitative PCR was performed with a Stratagene Mx3000P 153 QPCR System (Agilent Technologies, USA), using the fluorescent 154 dye SYBR-Green approach (TAKARA, Dalian, China). Three 155 coupled primers (Table 1) were used to amplify the AOB and 156 NOB (Nitrobacter, Nitrospira), namely amoA-1f/amoA-2f, FGPS872f/ 157 FGPS1269r and NSR1113f/NSR1264r (Park et al., 2010; Degrange 158 and Bardin, 1995; Geets et al., 2007). Each PCR amplification 159 mixture (20 µL) consisted of 10 µL SYBR® Premix Ex Taq™ 160 (Takara, Dalian, China), 0.4 µL ROX Reference Dye50, 1 µL bovine 161 serum albumin (25 mg/mL), 0.2 μ L of each primer (10 μ mol/L) and 162 a 2 µL template of DNA (1–10 ng). The PCR amplification program 163 performed denaturation for 3 min at 96°C followed by 40 cycles of 164

Target	Primer name	Sequence (5'-3')	Temp. (°C)
AOB	amoA-1f/amoA-2r	ggggtttctactggtggt/cccctckgsaaagccttcttc	53
Nitrobacter	FGPS872f/FGPS1269r	ctaaaactcaaaggaattga/ttttttgagatttgctag	56
Nitrospira	NSR1113f/NSR1264r	cctgctttcagttgctaccg/gtttgcagcgctttgtaccg	53
	EUB338f/Ntspa0685M	actcctacgggaggcagc/cgggaattccgcgctc	53

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