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Review

Activated carbon adsorption of quinolone antibiotics in water: Performance, mechanism, and modeling

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ABSTRACT

The extensive use of antibiotics has led to their presence in the aquatic environment, and introduces potential impacts on human and ecological health. The capability of powdered activated carbon (PAC) to remove six frequently used quinolone (QN) antibiotics during water treatment was evaluated to improve drinking water safety. The kinetics of QN adsorption by PAC was best described by a pseudo second-order equation, and the adsorption capacity was well described by the Freundlich isotherm equation. Isotherms measured at different pH showed that hydrophobic interaction, electrostatic interaction, and π - π dispersion force were the main mechanisms for adsorption of QNs by PAC. A pH-dependent isotherm model based on the Freundlich equation was developed to predict the adsorption capacity of QNs by PAC at different pH values. This model had excellent prediction capabilities under different laboratory scenarios. Small relative standard derivations (RSDs), i.e., 0.59%–0.92% for ciprofloxacin and 0.09%–3.89% for enrofloxacin, were observed for equilibrium concentrations above the 0.3 mg/L level. The RSDs increased to 11.9% for ciprofloxacin and 32.1% for enrofloxacin at $\mu\text{g/L}$ equilibrium levels, which is still acceptable. This model could be applied to predict the adsorption of other chemicals having different ionized forms.

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68 Introduction

69 The extensive use of antibiotics has been paid more and more
70 attention in the fields of medicine, public health and environ-
71 mental protection. The majority of antibiotics are poorly
72 metabolized and thus excreted by humans or animals, and
73 can directly enter natural waters in large quantities (Gobel et al.
74 2005; Baquero et al. 2008). Also, traditional wastewater treat-
75 ment processes cannot remove these chemicals efficiently
76 (Westerhoff et al. 2005). The presence of antibiotic residues in
77 the environment has constituted an emerging threat to human
78 health and ecology, through entering water supplies, promotion
79 of antibiotic-resistant bacteria and disturbance of ecological
80 balance (Sanderson et al. 2004; Rizzo 2012; Harnisz et al. 2015;
81 Osinska et al. 2016).

82 Quinolones (QNs) are a category of antibiotics sharing the
83 structure of 4-quinolyl. QNs have been detected in aquatic
84 environments throughout the world. The most commonly used
85 antibiotics are ciprofloxacin (CIP), ofloxacin (OFL) and norfloxacin
86 (NOR). The residual concentration of CIP was found to be as high
87 as 600 ng/L in the effluent of a wastewater treatment plant
88 in Spain (Reverte et al. 2003). The average concentrations of
89 OFL and NOR from four wastewater treatment plants in the
90 U.S. were 470 ng/L and 400 ng/L, respectively (Brown et al. 2006).
91 The extremely high concentrations of 100,000 ng/L of CIP and
92 7600 ng/L of OFL were reported in the wastewater of a hospital in
93 Sweden (Huang et al. 2011).

94 In China, QN pollution in aquatic environments may be
95 even more severe (Bu et al. 2013; Sui et al. 2015). Zhang et al.
96 (2015) reported that the proportion of QN usage (17%) among
97 all antibiotics in China was higher than in the U.S. (<8%
98 for humans, <10% for animals) and other countries. Specifi-
99 cally, Zhang et al. (2015) reported that the maximum detected
100 concentrations of several QNs (such as CIP, OFL, and NOR)
101 in aquatic environments (not including wastewater) in
102 China were as high as 7560 ng/L, and the average concentration
103 of all detected QNs was 303 ng/L. The reported QN concentra-
104 tions in aquatic environments from Italy (9 ng/L), U.S. (up to
105 120 ng/L), and Germany (20 ng/L) are much lower than that in
106 China.

107 The use of granular activated carbon (GAC) or powdered
108 activated carbon (PAC) has been proven to be an effective
109 technology for the removal of bulk organic matter as dissolved

organic carbon (DOC) (Mckay et al. 1985). GAC and PAC can
also remove various organic pollutants in water, including
odorants (Li et al. 2015a), disinfection by-product precursors
(Chen et al. 2015; Liao et al. 2015), and organic chemical spills
(Zhang and Chen 2009; Zhang et al. 2011).

115 Adams et al. (2002) reported that PAC adsorption was
116 effective in removing several antibiotics, including carbadox,
117 sulfachlorpyridazine, sulfadimethoxine, sulfamerazine, sulfa-
118 methazine, sulfathiazole, and trimethoprim. More recently,
119 Putra et al. (2009) demonstrated that 30 g/L of PAC could
120 remove 95% of 317 mg/L amoxicillin from pharmaceutical
121 wastewater. The study of Peng et al. (2012) concluded that
122 0.6 mg/L PAC could reduce the ofloxacin and norfloxacin from
123 100 to 1 $\mu\text{mol/L}$ at pH 7.0 ± 0.2 .

124 Adsorbents other than AC have been used to remove QNs
125 from water. Peng et al. (2012) studied the adsorption of NOR and
126 OFL by carbon nanotubes (CNT) and found that the adsorption
127 was affected by hydrophobic effects and molecule structure.
128 Yang et al. (2012) showed that the two main mechanisms for NOR
129 and CNT adsorption by porous resins were hydrophobicity and
130 micropore-filling. Neutral pH generally benefited NOR adsorption
131 by these adsorbents. Ötoker and Akmeahmet-Balciođu (2005) used
132 natural zeolite to remove enrofloxacin (ENR) from water, and
133 reported that the adsorption capacity increased with decreasing
134 pH. Yao et al. (2013) applied sludge-derived biochar to adsorb
135 gatifloxacin, and reported a correlation between the adsorption
136 capacity and the volatiles content of the sludge source. How-
137 ever, these studies are still too limited and inconsistent to
138 enable understanding of the mechanism of adsorption and
139 development of a model to predict the adsorption behavior of
140 QNs.

141 This study evaluated the relationship between the molec-
142 ular structure of QNs and their PAC adsorption behavior at
143 different pH levels. Moreover, a pH-dependent isotherm
144 model (PIM) for the investigation of QN adsorption was
145 introduced. This model was first developed based on the
146 Freundlich adsorption isotherms obtained by experimental
147 data at different pH values (Li et al. 2015a). This model is able
148 to predict the adsorption capacity in pH range of 0–14 and
149 assists in understanding the mechanisms of QN adsorption
150 onto PAC. However, there are still many improvements that
151 can be made for the model, such as verification of ng/L level
152 adsorption and competitive adsorption by organic matters.

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