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# Evolutionary processes and sources of high-nitrate haze episodes over Beijing, Spring

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### ABSTRACT

Rare and consecutive high-nitrate haze pollution episodes were observed in Beijing in spring 18 2012. We present detailed characterization of the sources and evolutionary mechanisms of this 19 haze pollution, and focus on an episode that occurred between 15 and 26 April. Submicron 20 aerosol species were found to be substantially elevated during haze episodes, and nitrates 21 showed the largest increase and occupation (average: 32.2%) in non-refractory submicron 22 particles (NR-PM<sub>1</sub>), which did not occur in other seasons as previously reported. The haze episode 23 013 (HE) was divided into three sub-episodes, HEa, HEb, and HEc. During HEa and HEc, a shallow 24 boundary layer, stagnant meteorological conditions, and high humidity favored the formation of 25 high-nitrate concentrations, which were mainly produced by three different processes -26daytime photochemical production, gas-particle partitioning, and nighttime heterogeneous 27 reactions — and the decline in visibility was mainly induced by  $NR-PM_1$  However, unlike HEa and 28 HEc, during HEb, the contribution of high nitrates was partly from the transport of haze from the 29 southeast of Beijing — the transport pathway was observed at ~800-1000 m by aerosol Lidar — 30 and the decline in visibility during HEb was primarily caused by PM<sub>2.5</sub>. Our results provide useful 31 information for air quality improvement strategies in Beijing during Spring. 32 © 2016 The Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences. 33 Published by Elsevier B.V.

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### 48 Introduction

49China is currently facing tremendous challenges with respect to 50air quality because of substantial increases in anthropogenic emissions associated with rapid industrialization, motorization, 51and urbanization (Chan and Yao, 2008). Haze pollution has 52become of interest to the public, government, and atmospheric 53 scientists in China. Haze not only reduces visibility and affects 54transportation but also exerts detrimental effects on human 55health. The number of patients with respiratory and cardiovas-56 Q15 Q14 cular diseases increases significantly during haze episodes (HEs) (Zhang et al., 2014b). There is an increase in participation in 58 outdoor activities after the cold winter season and people are 59 more likely to be exposed to the polluted atmosphere during 60 Spring. Therefore, gaining an understanding of the sources, 61 formation mechanisms, and evolutionary processes of haze 62 pollution in spring is of top priority in the development of 63 effective environmental protection policies and making accurate 64 assessments of health and climate impacts. 65

Owing to growing concern about the effects of haze pollution, 66 extensive studies have been conducted in recent years to 67 investigate the sources and formation mechanism of haze in 68

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China (An et al., 2007; Huang et al., 2012b; Li et al., 2010; Liu et al., 69 2013, 2016; Ma et al., 2010; Sun et al., 2006; Tao et al., 2012). 70 However, most previous studies have been conducted in the 71 Summer, Autumn or Winter (Han et al., 2015; Huang et al., 2012a; 72Li et al., 2010; Sun et al., 2012, 2013a, 2013b; Zhao et al., 2013), 016 while studies focusing on spring are limited. It is acknowledged 74 that the aerosol composition and processing, and boundary 75 layer characteristics are significantly different between seasons, 76 77 owing to differing meteorological conditions and source emis-78 sions. Moreover, despite previous efforts made to characterize spring haze pollution (Shen et al., 2016), its sources and 79 evolutionary processes have not yet been fully clarified. 80

Haze pollution often evolves rapidly and dynamically in 81 the atmosphere, and involves complex processes such as 82 direct emissions, secondary formation, aging and mixing, and 83 fog processing. The chemical, physical, and optical properties 84 of aerosol particles may change significantly in a few hours. 85 However, previous studies based on filter measurements are 86 limited by either low time resolution from hours to days or 87 sampling artifacts, such as evaporative loss (Sun et al., 2014). 88 In our work, an Aerodyne Aerosol Chemical Speciation 89 Monitor (ACSM) is employed, which is unique in its ability to 90 provide non-refractory submicron aerosol (NR-PM<sub>1</sub>) species 91 92(organics, sulfates, nitrates, ammonium, and chlorides) in real 93 time with a high time resolution ranging from seconds to 94 minutes (Ng et al., 2011).

95 In this study, we employ various high temporal resolution 96 instruments to investigate and comprehensively characterize the sources and evolutionary mechanisms of a rare and 97 lengthy high-nitrate haze pollution episode that occurred in 98 Beijing from April 15 to 26, 2012. The formation mechanisms 99 of the largest increase and occupation species, nitrates, are 100 investigated in detail and the sources of haze pollution, e.g., 101 local sources versus regional transport, are elucidated. In 102particular, the relationship between chemical evolution of the 103submicron aerosol composition, boundary layer meteorolog-104 ical parameters, and visibility is estimated in detail. 105

#### 1. Experimental 106

#### 1.1. Sampling site 108

Field measurements were carried out from April 15 to 26, 2012, 109at the Institute of Atmospheric Physics (IAP), Chinese Acad-110emy of Sciences (39°58'28"N, 116°22'16"E), which is located 111 between the north 3rd and 4th ring roads in Beijing. The 112sampling site is located on the roof of a second-story building 113(~8 m high). All aerosol and gaseous species, including 114 NR-PM<sub>1</sub> species, PM<sub>2.5</sub>, NO, NO<sub>2</sub>, NO<sub>y</sub>, O<sub>3</sub>, SO<sub>2</sub>, CO, and aerosol 115 116extinction coefficients were measured at this site. In addition, meteorological variables, including temperature, relative 117 humidity (RH), wind speed, and wind direction, were obtained 118 119 from the meteorological tower (325 m) at IAP.

#### 1.2. Aerosol and gaseous species measurements 120

NR-PM<sub>1</sub> species (including organics, sulfates, nitrates, ammo-017

nium, and chlorides) were measured in situ with an ACSM (Ng 122

et al., 2011). The aerosol sampling setup and ACSM operations 123

in this study were the same as those in Sun et al. (2012). 124 Following the procedures described in Ulbrich et al. (2009), the 125 positive matrix factorization (PMF) was performed using the 126 PMF2.exe algorithm (v4.2) in robust mode (Paatero and 127 Tapper, 1994) to obtain the ACSM organic aerosol (OA) mass 128 spectra. PMF2 solutions were then evaluated with an Igor 129 Pro-based PMF Evaluation Tool (PET, V 2.04) (Ulbrich et al., 130 2009) following the procedures detailed in Zhang et al. (2011). 131 A two-factor solution that included a hydrocarbon-like OA 132 (HOA) and an oxygenated OA (OOA) with fpeak = 0 was 133 chosen in this study. A detailed ACSM data analysis of Beijing 134 aerosol data can be found in Sun et al. (2013a). 135

PM<sub>2.5</sub> mass was simultaneously measured using a tapered 136 element oscillating microbalance (TEOM series 1400a, Thermo 137 Scientific, USA), and the collocated gaseous species (including 138 CO, SO<sub>2</sub>, NO<sub>2</sub>, NO/NO<sub> $\nu$ </sub>, and O<sub>3</sub>) were measured using various 139 gas analyzers obtained from Thermo Scientific. 140

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(1)

### 1.3. Lidar observation

A dual-wavelength depolarization Lidar (Model: L2S-SM), 142 developed by the National Institute for Environmental Studies 143 (NIES), was operated during this field campaign. The Lidar was 144 continuously operated at 15-min intervals at a height resolu- 145 tion of 30 m, setting a boundary condition at 3 km, and the 146 measured extinction coefficients and depolarization ratio at a 147 wavelength of 532 nm. Detailed information related to the 148 Lidar can be found in Yang et al. (2010). The Fernald inversion 149 method (Fernald, 1984) was used to derive the extinction 150 coefficient with the Lidar ratio (extinction-to-backscatter 151 ratio) set to 50 sr (Liu et al., 2002) in the inversion process. 152

### 1.4. Aerosol optical properties

Optical properties of aerosol particles (scattering and absorp- 154 tion) were measured at Beijing Normal University (BNU), a site 155 located approximately 3 km to the southwest of the IAP. The 156 sampling site is located on the roof of a six-floor building (~20 m 157 above ground level). The light scattering coefficient of dry 158 particles,  $b_{sp}$  (dry), at a wavelength of 532 nm was measured by 159 an Ecotech (Australia) M9003 integrating nephelometer (Liu et 160 al., 2008). A processor-controlled heating system was used to 161 dry aerosol particles in the sampling line (RH <  $\sim$  30%). The 162 nephelometer was routinely calibrated by zero and span 163 checks, and the truncation error was corrected (Liu et al., 2008). 164

A multi-angle absorption photometer (MAAP) (Petzold and 165 Schönlinner, 2004) was operated at an incident light wavelength 166 of 670 nm on the principle of light attenuation due to 167 absorption by aerosols deposited on the quartz fiber filter. The 168 aerosol absorption coefficient,  $b_{\rm ap}$ , was then calculated by 169 converting the mass concentration of black carbon (BC) to  $b_{ap}$  170 using a mass absorption efficiency of 6.6  $\,m^2\!/g$  , as suggested by  $\,_{171}$ the manufacturer. 172

The scattering and absorption coefficients were scaled to 173 values at the wavelength of 550 nm using a power-law 174 wavelength dependence of aerosol optical properties, as given 175 in Eqs. (1) and (2): 176

$$\frac{b_{\rm sp}(\lambda_1)}{b_{\rm sp}(\lambda_2)} = \left(\frac{\lambda_1}{\lambda_2}\right)^{-{\rm SAC}}$$

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