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Sediment pollution characteristics and in situ control in a deep drinking water reservoir

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Sediment release inhibition

ABSTRACT

Sediment pollution characteristics, *in situ* sediment release potential, and *in situ* inhibition of sediment release were investigated in a drinking water reservoir. Results showed that organic carbon (OC), total nitrogen (TN), and total phosphorus (TP) in sediments increased from the reservoir mouth to the main reservoir. Fraction analysis indicated that nitrogen in ion exchangeable form and NaOH-extractable P (Fe/Al-P) accounted for 43% and 26% of TN and TP in sediments of the main reservoir. The Risk Assessment Code for metal elements showed that Fe and Mn posed high to very high risk. The results of the in situ reactor experiment in the main reservoir showed the same trends as those observed in the natural state of the reservoir in 2011 and 2012; the maximum concentrations of total OC, TN, TP, Fe, and Mn reached 4.42 mg/L, 3.33 mg/L, 0.22 mg/L, 2.56 mg/L, and 0.61 mg/L, respectively. An *in situ* sediment release inhibition technology, the water-lifting aerator, was utilized in the reservoir. The results of operating the water-lifting aerator indicated that sediment release was successfully inhibited and that OC, TN, TP, Fe, and Mn in surface sediment could be reduced by 13.25%, 15.23%, 14.10%, 5.32%, and 3.94%, respectively.

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43 Introduction

Various pollutants adsorb to sediments that accumulate at 44 45the bottom of reservoirs. These sediments accumulate over 46 long periods of time and can act as new pollutant sources to the overlying water (Sun et al., 2009; Młynarczyk et al., 2013). 03 48 The forms and types of pollutants usually vary in different sediments because bonding forces vary with respect to grain 49size (Ruttenberg, 1992; Bo et al., 2003), which results in 50different release times and release potentials. 51

As is well known, nitrogen and phosphorus are the most common macronutrients for phytoplankton growth. Furthermore, the nitrogen and phosphorus transport processes between sediments and overlying water play a crucial role in the aquatic ecological environment (Huang et al., 2015). 56 Meanwhile, heavy metal pollution is an environmental crisis 57 that accompanies rapid economic development in many 58 countries. Metal elements constitute an environmental 59 threat by creating serious human health hazards and affect- 60 ing ecological food chains (Souza and Wasserman, 2015; 61 Xuelu and Chen-Tung Arthur, 2012). To clarify the transport 62 processes, first we should have a clear understanding of 63 the nitrogen, phosphorus, and metal element fractions 64 and release potentials in sediments. During the stratifica- 65 tion period, sediment release in Grand Lake (Nikolai et al., 66 2014) yielded total nitrogen (TN) and total phosphorus (TP) 67 concentrations in the hypolimnion of 1.1 mg/L and 0.17 mg/L, 68 respectively. Soluble Fe and Mn also reached problematic 69

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levels during the summer stratification in Carvins Cove
Reservoir (Nikolai et al., 2014). Thus, sediments act as a source
of pollution.

Many laboratory studies have reported on sediment 73 release of N, P, Fe, and Mn (Wang et al., 2014; Wu and Hua, 74 2014; Zhang et al., 2014). Considering the sediment form 75 changes as well as pressure and temperature variations that 76 occur in real settings, laboratory release experiments are not a 77 78 reliable reflection of sediment release in reservoirs. In the 79 present study, a new in situ reactor was used to study the release characteristics of pollutants in a canyon-shaped 80 deep-water reservoir. Using the intact sediment and true 81 pressure, the in situ reactor was used to produce more credible 82 data. Of the in situ sediment remediation strategies, sediment 83 capping has been applied in Lake Taihu (Xu et al., 2012) for 84 internal P control, but this method could not be applied in the 85 Jinpen Reservoir for deposition carried by runoff. Additives 86 that were used in Suzhou Creek (Qian et al., 2009) successfully 87 stabilized the metal elements, but this may also cause new 88 problems in drinking water reservoirs, such as increased iron 89 from additives containing ferrihydrite. Dredging for sediment 90 remediation is always toxic to phytoplankton and autotrophic 91 bacteria (Nayar et al., 2004). In this study, sediment release 92 93 control using water-lifting aerators (WLAs) was shown to be 94 clean and environmentally safe. We also demonstrated sig-95 nificant results: concentrations of organic carbon (OC), TN, TP, 96 Fe, and Mn in the sediment were reduced during the operation 97 of WLAs. The goal of this study was to understand the pollution status and release potential of reservoir sediments 98 and to improve the ability of WLAs to control the release of 99 100 internal sediments.

102 1. Materials and methods

103 1.1. Sampling sites

Jinpen Reservoir (34°42′ 0-34°13′N; 107°43′-108°24′E) is the 104 main drinking water reservoir of Xi'an City in Shaanxi 105Province, China. Jinpen Reservoir is located at the foot of the 106 Qinling Mountains, 90 km from Xi'an City, as shown in Fig. 1. 107 108 It is a canyon-shaped deep-water reservoir, and the main reservoir length extends 3.5 km. The total capacity of the 109reservoir is 2.0×10^8 m³. When full, the surface area of the 110 reservoir is 4.55 km². The main function of the reservoir is 111 urban water supply; agricultural irrigation, power generation, 112and flood control are accessory functions. 113

The Heihe River, which originates in the Qinling Mountains, 114is the main water supply for Jinpen Reservoir. The river is 115 $91.2 \; \text{km}$ long with a catchment area of $1418 \; \text{km}^2\text{,}$ and the 116 117 catchment is largely undeveloped and consists primarily of forested mountains. Along the reservoir, six representative 118 sampling sites were chosen. S1 (33°58'26.91"E, 108°08'50.02"N, 119 depth 1 m) was located at the reservoir mouth; S2 (34°00'16.7"E, 120121 108°10′38.67″N, depth 5 m) and S3 (34°01′14.24″E, 108°10′58.1″N, depth 20 m) were upstream from the reservoir; S4 (34°01′58.17″E, 122 123108°11′13.78″N, depth 40 m) was near the middle of the reservoir; S5 (34°02'35.95"E, 108°11'52.83"N , depth 65 m) and S6 (34°02' 124 42.20"E 108°12'22.76"N 70, depth 70 m) represented the main 125126reservoir.



Fig. 1 - Position of Jinpen Reservoir and sampling sites (S1-S6).

1.2. Field work

Sediment samples were collected using a core sampler 128 (HL-CN, Hengling technology Ltd., Corp., China; Ni and 129 Wang, 2015). Sediment columns that showed no evidence of 130 disturbance were retained. One centimeter segments of the 131 sediment columns were collected, transferred to pre-cleaned 132 polythene bags, and sealed. Sediment samples were kept at 133 -4° C until laboratory processing (Zahra et al., 2013). 134

The *in* situ reactor utilized is a proprietary device specially 135 designed for the investigation of migration and transforma-136 tion of substances at the sediment-water interface, and its 137 schematic diagram is shown in Fig. 2. The reactor is a 138 stainless steel cylinder with its bottom open. The *in* situ 139 reactor was installed at the water-sediment interface using a 140 windlass with a wire rope, and the lower part of the reactor 141 was embedded into the sediment by its own weight. The 142 water in the reactor was separated from the surrounding 143 water, and dissolved oxygen (DO) consumption and sediment Q4



Fig. 2 - Schematic diagram of the in situ reactor.

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