

Photocatalytic degradation of polycyclic aromatic hydrocarbons on soil surfaces using TiO₂ under UV light

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Abstract

The photocatalytic degradation of phenanthrene (PHE), pyrene (PYRE) and benzo[*a*]pyrene (BaP) on soil surfaces in the presence of TiO₂ using ultraviolet (UV) light source was investigated in a photo chamber, in which the temperature was maintained 30 °C. The effects of various factors, namely TiO₂, soil pH, humic acid, and UV wavelength, on the degradation performance of polycyclic aromatic hydrocarbons (PAHs) were studied. The results show that photocatalytic degradation of PAHs follows the pseudo-first-order kinetics. Catalyst TiO₂ accelerated the photodegradation of PHE, PYRE and BaP significantly, with their half-lives being reduced from 533.15 to 130.77 h, 630.09 to 192.53 h and 363.22 to 103.26 h, respectively, when the TiO₂ content was 0.5%. In acidic or alkaline conditions, the photocatalytic degradation rates of the PAHs were greater than those in neutral conditions. Humic acid significantly enhanced the PAH photocatalytic degradation by sensitizing radicals capable of oxidizing PAHs. Photocatalytic degradation rates of PYRE and BaP on soil surfaces with 2% TiO₂ were different at UV irradiation wavelengths of 254, 310 and 365 nm, respectively. The synergistic effect of UV irradiation and TiO₂ catalysis was efficient for degradation of PAHs in contaminated soil. © 2008 Elsevier B.V. All rights reserved.

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1. Introduction

Polycyclic aromatic hydrocarbons (PAHs) are a class of persistent organic pollutants that are ubiquitous in the environment. They are considered as hazardous pollutants due to their toxicity, mutagenicity and carcinogenicity, and are classified as compounds with significant human health risk [1]. In industrial countries, anthropogenic activities are principal source of PAHs in soil [2,3]. Hundreds of thousands of gallons of used motor oil containing PAHs are disposed of improperly each year into soil [4]. This has prompted an intensive search for effective environmental remediation processes to “clean up” contaminated sites to some predetermined concentration level.

Recently photocatalytic processes involving TiO₂ semiconductor particles under ultraviolet (UV) light irradiation have shown the potential advantages to be used in several cases [5–10]. Solid TiO₂ (in the crystalline form of anatase) is a semi-

conductor that under UV irradiation can promote an electron (e[−]) from the valence band (VB) to the conduction band (CB), leaving a positive hole (h⁺) at the site where the electron was originally captured. When appropriate scavengers are present, oxidation/reductions can take place.

Heterogeneous photocatalysis of organic pollutants using TiO₂ under UV-irradiation and/or solar light has demonstrated successful performance in various remediation systems of polluted soil. The addition of small amounts of TiO₂ (0.5, 1, 2 and 3 wt.%) enhanced the photodegradation of *p,p'*-DDT on soil surfaces significantly, and soil pH, photon flux and humic substances affected the photocatalytic degradation [11,12]. The photocatalytic treatment using TiO₂ combined with solar light was very efficient in the destruction of pesticide Diuron in the top 4 cm of contaminated soil, and the degradation rate was markedly dependent on the irradiation intensity [13]. The contaminated soils containing atrazine, 2-chlorophenol, 2,7-dichlorodibenzodioxin were mixed with TiO₂ and exposed to simulated solar radiation. The organic contaminants were destroyed in a relatively short time [14,15]. Polychlorinated biphenyls (PCBs) in soil can be effectively photodegraded

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in a dispersion containing anionic fluorinated surfactant and TiO₂ [16]. These studies show that photocatalytic processes are effective to decontaminate soils containing dangerous organic chemicals.

The photodegradation of PAHs in water catalyzed by TiO₂ has been extensively studied in the past few years. Catalyst TiO₂ can efficiently photocatalyze the oxidation of PAHs, such as anthracene, fluorene and naphthalene, using artificial or sunlight radiation sources [17]. Phenanthrene with poor aqueous solubility was able to be easily degraded after pre-adsorbed on TiO₂ in an aqueous dispersion under UV irradiation. The pH of the dispersion and Ph/TiO₂ value had little effect on the photooxidation rate of phenanthrene catalyzed by TiO₂ [18]. In fact, PAH compounds exist as a mixture of multi-ring structures (i.e. 2-through 6-membered ring PAHs). These compounds were intentionally photocatalyzed as a mixture, more representative of the way these compounds had encountered in actual field remediation processes. Ireland et al. [4] investigated the photocatalytic degradation of a mixture of 16 PAHs in aqueous suspensions of high surface area TiO₂ illuminated with ultraviolet light. In order to enhance TiO₂ utilization rate, García-Martínez et al. [19] reported the photocatalytic degradation of naphthalene in water using TiO₂, supported on glass Raschig rings as catalyst. Pal and Sharon [20] studied the photocatalytic degradation of saturated aqueous solution of naphthalene and anthracene over thin films of porous TiO₂ particles on glass substrate.

However, there are few studies investigating the photocatalytic degradation of PAHs on soil surfaces using TiO₂ as the catalyst under UV irradiation.

The present paper focuses on the possible usefulness of heterogeneous photocatalytic processes using TiO₂ for the degradation of PAHs present in soil. Phenanthrene (PHE), pyrene (PYRE) and benzo[a]pyrene (BaP) were chosen as representative compounds for PAHs. Furthermore, the main influencing factors, such as photocatalyst concentration, soil pH, humic acid (HA) and wavelength of UV irradiation, were taken into consideration.

2. Materials and methods

2.1. Chemicals

The test PAHs, namely PHE, PYRE, and BaP, were purchased from Fluka, Germany, and were used without further purification. Methanol (HPLC grade) was purchased from Shandong Yuwang Company, China. Hexane and dichloromethane were purchased from Tianjin Concord Technical Company, China. Particles of TiO₂ were purchased from Degussa (P25, anatase, surface area ~50 m² g⁻¹, mean diameter 20 nm). Humic acid was purchased from Tianjin Jinke Company, China.

2.2. Experimental soil

Surface soil sample (top 10 cm) was collected from the Ecological Station of the Shenyang, Institute of Applied Ecology, Chinese Academy of Sciences. After being air-dried, the soil

Table 1
Characteristics of the experimental soil

pH	TOC (%)	Texture (%)			Bulk density (g cm ⁻³)
		Sand	Silt	Clay	
6.8	1.78	21.4	46.5	32.1	2.53

sample was passed through a 1 mm sieve. To prepare a sterile soil, the sieved soil was autoclaved at 121 °C for 30 min twice, and stored in dark before use. The characteristics of the soil are presented in Table 1.

The soil samples were spiked with methanol solutions of PHE, PYRE and BaP, respectively, mixed thoroughly and then air-dried for the evaporation of methanol. The concentrations of PHE, PYRE, BaP were calculated to be 40 mg kg⁻¹ in the soil, respectively.

The catalyst load was made by adding desired amount of TiO₂ to the PAH spiked soil, followed by shaking for 30 min in a 500 ml closed stainless steel reactor.

To study the influence of soil pH, the pH of PAH contaminated soil sample was adjusted to 4.2 or 9.7 using NaOH or H₂SO₄, and TiO₂ addition amount was 2%.

Effects of humic acid on PAH degradation were also examined. One hundred grams of HA was dissolved in 0.1 M NaOH solution, diluted with 1000 ml distilled water as storage solution. Different dosages of HA storage solution were added to PAH contaminated soil samples. The additive concentrations of HA in the soil samples were 5, 10, 20 and 40 mg kg⁻¹, respectively. After being air-dried, the soil samples were mixed with 2 wt.% TiO₂.

2.3. Photodegradation chamber

Photodegradation studies were performed in a chamber as shown in Fig. 1. In parallel, two arrays of nine UV lamps were fixed in the top and middle of the chamber, respectively, with a distance of 60 mm between two lamps in the same array. The distance between the lamps and soil samples was 150 mm. The UV irradiation intensity was 1071 μW cm⁻². The UV lamps (Phillips ATLD 20 W, Model UVA, UVB and UVC) can be changed for wavelength variation, and available wavelengths in the experiment were 254, 310 and 365 nm, respectively. Petri dishes containing the experimental soil samples were placed on the shelves for photo irradiation. Temperature within the chamber can be adjusted through refrigerator, heater and fans in the chamber. Chamber temperature was 30 °C throughout all the experiments.

2.4. Photocatalytic degradation experiments

In the photocatalytic degradation experiments, 5 g of soil sample was evenly spread on the Petri dishes at three replicates. Light proof Petri dishes containing also 5 g of soil sample were set as control for the measurement of non photocatalyzed PAHs loss throughout all the experiments. Effect of TiO₂ amount, soil pH, humic acid and wavelength on the photocatalytic degrada-

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