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Effects of calcite and magnesite application to a declining Masson pine forest on strongly acidified soil in Southwestern China



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HIGHLIGHTS

· First long-term liming experiment for forests in China.

• Addition of Ca²⁺ to forests may result in nutrient imbalance such as P deficiency.

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ABSTRACT

Liming of strongly acidified soil under a Masson pine (*Pinus massoniana* Lamb.) forest was studied through a seven-year field manipulation experiment at Tieshanping, Chongqing in Southwestern China. To distinguish between the individual effects of Ca^{2+} and Mg^{2+} addition, we separately applied calcite ($CaCO_3$) and magnesite ($MgCO_3$), rather than using dolomite [$CaMg(CO_3)_2$]. Both calcite and magnesite additions caused a significant increase in pH and a decrease in dissolved inorganic monomeric aluminium (Al_i) concentration of soil water. Ecological recovery included increases of herb biomass (both treatments) and Mg content in Masson pine needles (magnesite treatment only). However, the growth rate of Masson pine did not increase under either treatment, possibly because of nutrient imbalance due to phosphorus (P) deficiency or limited observation period. In China, acid deposition in forest ecosystems commonly coincides with large inputs of atmogenic Ca^{2+} , both enhancing Mg^{2+} leaching. Calcite addition may further decrease the Mg^{2+} availability in soil water, thereby exacerbating Mg^{2+} deficiency in the acidified forest soils of southern and southwestern China. The effect of anthropogenic acidification of naturally acid forest soils on P availability needs further study.

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1. Introduction

Acid deposition in China has been a serious environmental problem during the last several decades (Zhao and Sun, 1986; Larssen et al., 2006). Soil acidification, particularly in southern China (Larssen et al., 2006), was found to have severe impacts on tree growth and tree crown condition (Wang et al., 2007). Forest dieback since the early 1980's, attributed to acid rain, has also been reported for this area (Ma, 1991; Hao et al., 1998). Severe forest dieback occurring in central Europe in the mid-1980s was believed to be primarily due to direct, above-ground effects of air pollutants, rather than to indirect effects due to soil acidification (Ulrich et al., 1980; Gorham, 1989; Pitelka and Raynal, 1989). In contrast, forest decline in North America has been

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ascribed to indirect effects such as Al toxicity and depletion of soil base cations (mainly Ca²⁺) in response to long-term soil acidification (Shortle and Smith, 1988; Bernier and Brazeau, 1988; Lawrence et al., 1995).

Reducing atmospheric emissions likely decreases both direct and indirect effects of acid rain on forest ecosystems, and remains a main environmental priority. In addition, remediation of already acidified soils and restoration of degraded forest ecosystems is important to limit further ecological degradation and economic loss. Since the application of calcitic and/or dolomitic limestone increases the base saturation of soils and decreases concentrations of H⁺ and toxic Al species in soil water, 'liming' is considered to be an effective tool for the rehabilitation of naturally or anthropogenically acidified ecosystems, and has been studied and practised for many years in Europe and North America (Hüttl and Zöttl, 1993; Lundström et al., 2003; Löfgren et al., 2009; Moore et al., 2012; Van der Perre et al., 2012). Elevated concentrations of Al³⁺ by soil acidification were found to cause root damage and reduce the uptake of important nutrient cations such as magnesium (Mg^{2+}) in trees (Ulrich et al., 1980; Sverdrup et al., 1992; de Wit et al., 2010). The molar ratio of Al^{3+} to base cation $(Ca^{2+}+Mg^{2+}+K^+)$ (Al/BC) in soil water is commonly used as a risk indicator for forest decline due to acidification (Cronan and Grigal, 1995), and 1.0 is widely used as a critical limit in critical load assessments (Sverdrup et al., 1992). In contrast to expectations, however, lime applied to the soil surface has not always brought an overall improvement in forest growth. Most investigations in Europe have shown that liming does not significantly improve tree growth and tree vitality (Hüttl and Zöttl, 1993; Ljundström and Nihlgård, 1995; Hindar et al., 2003; Lundström et al., 2003; Røsberg et al., 2006), while some studies in North America (Long et al., 2011; Moore et al., 2012) and in Europe (van der Perre et al., 2012) have shown positive impacts. Further, there is no clear relation between elevated Al/BC and forest decline (de Wit, et al., 2010).

Although many studies have investigated liming treatments of forest soils in Europe and North America, the potential of surface liming to remediate acidified forest soils and increase forest growth has not been investigated in long-term experiments in China to our knowledge. In addition to the vegetation (and its sensitivity to acidification), soil, and climatic conditions, the composition of atmospheric deposition in China differs considerably from that in other countries. In particular, the atmospheric deposition of Ca^{2+} is very high (although not as high as SO_4^{2-} on an equivalent basis; Larssen and Carmichael, 2000; Larssen et al., 2011), due to large Ca^{2+} emissions, from both natural (e.g. wind-blown dust) and anthropogenic sources (e.g. cement processing and coal combustion) (Zhao et al., 2011). Since Ca^{2+} is an important acid-neutralizing component of soil, its atmospheric input provides unintentional liming of terrestrial ecosystems. Some studies have indicated that the application of Ca²⁺ could decrease crown thinning, and increase growth rate and seedling density of sugar maple (Juice et al., 2006; Moore et al., 2012).

Here we present a seven-year field manipulation experiment in a Masson pine (*Pinus massoniana* Lamb.) dominated stand growing on highly acidified soil in Tieshanping, Chongqing in southwestern China. Masson pine forests, both natural and planted, are widely distributed in the humid subtropical areas of southern and southwestern China (Wu, 1980), and it has been suggested that Masson pine is very sensitive

to soil acidification (Gao et al., 1992; Zhang et al., 1995; Wang et al., 2007). To distinguish between the individual effects of Ca^{2+} and Mg^{2+} , we used high-purity CaCO₃ and MgCO₃, respectively, rather than dolomite $[CaMg(CO_3)_2]$, which contains both Ca^{2+} and Mg^{2+} at a fixed ratio, and is commonly applied to acidified soils in Europe. The effect of Mg²⁺ on tree growth is of particular importance because Mg is an important nutrient, and its deficiency has often been associated with poor tree growth on acidified soils (Horsley et al., 2000; de Wit et al., 2001, 2010). Soil Mg^{2+} deficiency may be a major problem in acidified forest soils in China since long-term Mg²⁺ leaching may have been aggravated by the large fluxes of atmogenic Ca²⁺. The objective of the present study was to quantify the long-term responses of a widely-distributed subtropical forest on acid soil in China to different liming treatments. Based on the documented high Al³⁺ concentration in soil water (Larssen et al., 2006) and severe forest defoliation (Wang et al., 2007), we hypothesized that in particular the magnesite treatment would improve the Mg status and overall health of the forest.

2. Materials and methods

2.1. Experiment location and design

A field manipulation experiment was carried out in a Masson pine (P. massoniana Lamb.) stand planted in 1962 following the clear cutting of a natural Masson pine forest at Tieshanping (106°41.24′E, 29°37.42′ N). Tieshanping is located about 25 km northeast of the centre of Chongqing City, at an elevation of about 450 m above sea level (Fig. 1). The soil is locally called yellow earth (corresponding to a Haplic Acrisol in FAO; Surhone et al., 2010), with a thin (<1 cm) O horizon and pH about 4. The annual mean temperature and precipitation were 18.2 °C and 1105 mm, respectively, from 1971 to 2000 (Meteorological Station of Chongqing). The stand is homogeneous, dominated by Masson pine and some associated species such as Cunninghamia lanceolata (Lamb.) Hook. and Eurya groffi Merr. The stand density is about 800 stem ha⁻¹. Dominant species of ground vegetation are Camellia oleifera Abel, Randia cochinchinensis (Lour.) Merr., and Loropetalum chinense (R. Br.) Oliver in the underbrush layer; Dicranopteris pedata (Houtt.) Nakaike, Woodwardia japonica (L. F.) Sm., Dryopteris

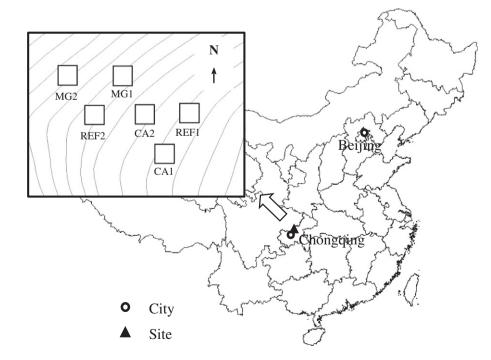


Fig. 1. Location of the experimental site and schematic presentation of sampling plots. Squares on the inset figure represent sampling plots. Contour interval of the contour map is 1 m.

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