



Cu-doped Bi₂O₃/Bi⁰ composite as an efficient Fenton-like catalyst for degradation of 2-chlorophenol



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ABSTRACT

Cu-doped BiO_x composites were prepared with and without citric acid. In the presence of citric acid, Bi⁰ and Cu²⁺/Cu⁺ were confirmed to coexist in the composite (BiCu9/Bi⁰) by X-ray diffraction and X-ray photoelectron spectroscopy. The catalyst showed high Fenton activity and stability for the degradation and mineralization of 2-chlorophenol in the pH range of 6–8. H₂O₂ was predominately converted into ·OH radicals in BiCu9/Bi⁰ suspension, as assessed by the electron spin resonance technique. In the composite, Bi⁰ inhibited the oxidation of H₂O₂ by Cu²⁺ into O₂ and ·OH by reducing Cu²⁺, enhancing the reaction of H₂O₂ with Cu⁺ to produce ·OH. Intermediate determination showed that 2-chlorophenol was decomposed mainly by ·OH radicals into hydroxylation products, subsequently decomposed into short-chain organic acids, and eventually decomposed into CO₂ and H₂O.

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1. Introduction

Iron-catalyzed hydrogen peroxide (H₂O₂) decomposition for generation of highly aggressive hydroxyl radicals (·OH) has been extensively developed as Fenton oxidation processes for environmental applications. A variety of catalytic iron species, including Fe²⁺ and Fe³⁺ salts [1,2], iron oxides [3–6], and iron-immobilized zeolites [7], have been exploited for degradation and mineralization of recalcitrant organic pollutants, as well as for disinfection. However, the widespread use of iron-based Fenton oxidation for pollution abatement and disinfection is still limited by the requirement for strict acidic conditions, the requirement for UV irradiation to accelerate the reductive generation of Fe²⁺ and the uncontrolled leaching of iron ions from the catalyst's surface [4,8]. Therefore, non-iron Fenton catalysts, such as Au [9], carbon materials [10], cerium oxides [11], manganese oxides [12], and Cu-containing catalysts [13–16], and their reactivity towards H₂O₂ activation have been explored to overcome these problems.

Among these non-iron catalysts, Cu-containing catalysts have attracted considerable attention because the redox properties of copper are similar to those of iron. Importantly, a Cu-based Fenton-like system could work over a broader pH range than the Fe-based redox system [13,14]. In particular, Cu²⁺/Cu⁺ cycling has been found to result in the generation of reactive oxygen species in the presence of H₂O₂ [17–19]. Compared to Fe³⁺/Fe²⁺ cycling,

which is usually limited by the reduction of Fe³⁺ due to the low rate constants [20], the reduction of Cu²⁺ by H₂O₂ occurs more easily than that of Fe³⁺, and Cu⁺ can react efficiently with H₂O₂ to form ·OH with a higher reaction rate than Fe²⁺ [21,22]. However, the mobilization of Cu⁺ limits its application due to its instability in water. The homogeneous reaction between Cu⁺ and H₂O₂ is severely inhibited by O₂ in water because Cu⁺ is quantitatively oxidized by O₂ to Cu²⁺, which decreases the effective [Cu⁺] available for reaction with H₂O₂ [22]. Therefore, the immobilization of Cu⁺ is required to overcome this problem. Two different methods for preparing copper-containing catalysts have been proposed: one yielding copper-doped samples and the other yielding copper supported on a porous solid substrate. Bismuth oxide (Bi₂O₃) is considered to be a good choice for a solid matrix because of its excellent catalytic, electrical, optical, and conducting properties [23,24]. Moreover, some reactions can proceed through a Bi(III)/Bi(0) or Bi(III)/Bi(I) redox couple in the presence of molecular oxygen [25]. Therefore, the aim of this work was to develop a highly active and stable heterogeneous Fenton-like catalyst by fixing Cu⁺ and Cu²⁺ in a Bi₂O₃ structure.

In this paper, a composite catalyst with Cu⁺/Cu²⁺ and Bi⁰/Bi³⁺ was successfully synthesized by a citric acid-based sol–gel route. The performance of the catalyst was evaluated by degrading 2-chlorophenol (2-CP), which is a common model contaminant for the advanced oxidation process, a germicide/disinfectant, and an important precursor for synthesizing pesticides and other chlorophenols [26]. The Fenton-like degradation pathway of 2-CP was also investigated. The catalytic mechanism of Cu-doped

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$\text{Bi}_2\text{O}_3/\text{Bi}^0$ was systematically studied on the basis of the effect of multivalent Cu and Bi^0 on the generation of $\cdot\text{OH}$ from H_2O_2 decomposition.

2. Experimental

2.1. Materials

Bismuth nitrate ($\text{Bi}(\text{NO}_3)_3 \cdot 5\text{H}_2\text{O}$), copper nitrate ($\text{Cu}(\text{NO}_3)_2 \cdot 3\text{H}_2\text{O}$), ethylene glycol, citric acid, 2-methoxyethanol, and H_2O_2 (30%, w/w) were purchased from Sinopharm Chemical Reagent Co., Ltd. 5,5-Dimethyl-1-pyrroline-N-oxide (DMPO) was supplied by Sigma. 2-CP was obtained from Acros. All chemicals were at least analytical grade and were used as received without further purification. Deionized water was used throughout this study.

2.2. Preparation of catalysts

Cu-doped $\text{Bi}_2\text{O}_3/\text{Bi}^0$ was prepared by a citric acid-based sol-gel route. The synthesized samples were designated as $\text{BiCu}_x/\text{Bi}^0$, where x indicated the Bi/Cu molar ratio. In the following, synthesis description for $\text{BiCu}_9/\text{Bi}^0$ is given. Typically, 9 mmol $\text{Bi}(\text{NO}_3)_3 \cdot 5\text{H}_2\text{O}$ and 1 mmol $\text{Cu}(\text{NO}_3)_2 \cdot 3\text{H}_2\text{O}$ were dissolved in 20 mL of 2-methoxyethanol, followed by adding 20 μL of 0.1 M HNO_3 , 10 mmol citric acid (as a complexant) and 10 mL ethylene glycol (as a dispersant). After stirring for 1 h at 60 °C, the synthesis medium was incubated at 100 °C for 10 h to form the gel. Finally, the gel was calcined at 500 °C for 2 h in air. As a reference, BiO/Bi^0 was prepared as described above without $\text{Cu}(\text{NO}_3)_2 \cdot 3\text{H}_2\text{O}$ added. Cu_2O was also synthesized using a previously reported process [27]. To clarify the formation and effect of Bi^0 , BiO and BiCu_9 were prepared as described above without the addition of citric acid.

2.3. Characterization

Powder X-ray diffraction (XRD) patterns were recorded using a Scintag-XDS-2000 diffractometer with $\text{Cu K}\alpha$ radiation ($\lambda = 1.540598 \text{ \AA}$). The generator voltage and tube current used were 40 kV and 40 mA, respectively. The X'Pert HighScore Plus software (version 2.2.4) was used to semi-quantitatively determine the weight fraction of constituents. X-ray photoelectron spectroscopy (XPS) data were measured via an AXIS-Ultra instrument from Kratos using monochromatic $\text{Al K}\alpha$ radiation (225 W, 15 mA, 15 kV) and low-energy electron flooding for charge compensation. To compensate for surface charge effects, binding energies were calibrated using the C1s hydrocarbon peak at 284.8 eV. Electron spin resonance (ESR) spectra were obtained using a Bruker model ESP 300E electron paramagnetic resonance spectrometer.

2.4. Procedures and analysis

Unless indicated otherwise, 0.07 g of catalyst powders was dispersed in 50 mL of a 0.078 mM 2-CP solution at natural pH (approximately 6.0). The solution pH was not adjusted or controlled because it did not vary by more than 0.2 pH units during the catalytic reaction. The suspensions were stirred for approximately 20 min to establish an adsorption/desorption equilibrium between the solution and catalyst. Afterwards, the reaction was started by adding 20 mM H_2O_2 as an oxidant. At given time intervals, part of the suspension sample was withdrawn and filtered to remove catalyst powders for analysis. The concentrations of 2-CP and the intermediates catechol, p-benzoquinone and phenol were measured using high-performance liquid chromatography (1200 series; Agilent) with a Zorbax SB-Aq column (5 μm , $4.6 \times 250 \text{ mm}$; Agilent). The analysis was carried out at 275 nm

using a 60/40% v/v mixture of methanol/ultrapure water as the mobile phase. The total organic carbon (TOC) of the solution was analyzed using a TOC-V_{CPH} analyzer (Shimadzu). The measurements of chloride ions (Cl^-) and short-chain organic acids were conducted using a Dionex model ICS 2000 ion chromatograph (IC) equipped with an IonPac AS11-HC analytical column ($4 \times 250 \text{ mm}$) and using 40 mM KOH as an eluent.

In addition, the possible cation leaching (Bi, Cu) was monitored by inductively coupled plasma optical emission spectrometry (ICP-OES) on an Optima 2000 (PerkinElmer, Inc.) instrument. To test the stability and recyclability of $\text{BiCu}_9/\text{Bi}^0$, the catalyst was filtered, washed with water, dried at 70 °C and used again in the second cycle. This process was repeated several times.

All of the above experiments were repeated in triplicate, and the data are presented as the average of the triplicates, with a standard deviation of less than 5%.

3. Results and discussion

3.1. Characterization of catalysts

The XRD patterns of the synthesized samples were shown in Fig. 1. The characteristic peaks of $\alpha\text{-Bi}_2\text{O}_3$ (JCPDS 71-0465) and Bi^0 (JCPDS 85-1329) were present in all of the samples prepared in the presence of citric acid. The lattice parameters decreased with increasing copper content in Bi_2O_3 up to a Bi/Cu ratio of 9:1 as shown in Table 1, indicating that Cu^{2+} was incorporated into the Bi_2O_3 structure to replace Bi^{3+} , due to the smaller ionic radius of Cu^{2+} (72 pm) than that of Bi^{3+} (96 pm) [28]. However, further increase in Cu content resulted in the occurrence of the impurities Bi_2CuO_4 (JCPDS 80-1906) and Cu_2O (JCPDS 78-2076) in $\text{BiCu}_3/\text{Bi}^0$ (Fig. 1). Therefore, there was an upper limit of Cu content that could be incorporated inside the Bi_2O_3 structure. When citric acid was not added in the preparation process of BiO and BiCu_9 , pure $\alpha\text{-Bi}_2\text{O}_3$ was produced and no Bi^0 was found (e, f in Fig. 1), indicating that the formation of Bi^0 depended on the presence of citric acid in the course of synthesis.

XPS analysis was used to confirm the oxidation states of the Bi and Cu elements on the surface of $\text{BiCu}_9/\text{Bi}^0$. The binding energy of $\text{Bi } 4f_{7/2}$ was 158.4 eV, which was assigned to Bi^{3+} [29]. Surface Bi^0 was not observed, but it appeared in the XRD patterns of $\text{BiCu}_9/\text{Bi}^0$, indicating that Bi^0 was in the interior of the catalyst and coated with Bi_2O_3 . Peaks corresponding to $\text{Cu } 2p_{3/2}$ were observed at 932.4 and 934.2 eV for reduced copper species and Cu^{2+} ,

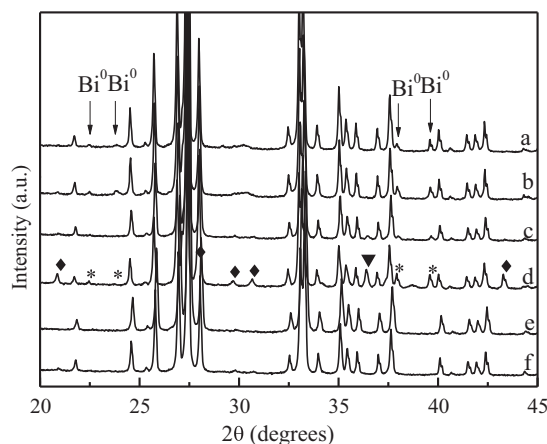


Fig. 1. XRD patterns of (a) BiO/Bi^0 , (b) $\text{BiCu}_{19}/\text{Bi}^0$, (c) $\text{BiCu}_9/\text{Bi}^0$, (d) $\text{BiCu}_3/\text{Bi}^0$, (e) BiO and (f) BiCu_9 . Some characteristic peaks in the patterns: (*) Bi^0 , (◆) Bi_2CuO_4 , (▼) Cu_2O .

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