



# Influence of biofilm thickness on nitrous oxide (N<sub>2</sub>O) emissions from denitrifying fluidized bed bioreactors (DFBBRs)



Ahmed Eldyasti<sup>a,b</sup>, George Nakhla<sup>b,c,\*</sup>, Jesse Zhu<sup>c</sup>

<sup>a</sup> Department of Civil Engineering, Lassonde School of Engineering, York University, Toronto, Ontario, Canada M3J 1P3

<sup>b</sup> Department of Civil and Environmental Engineering, The University of Western Ontario, London, Ontario, Canada N6A 5B9

<sup>c</sup> Department of Chemical and Biochemical Engineering, The University of Western Ontario, London, Ontario, Canada N6A 5B9

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## ABSTRACT

Nitrous oxide (N<sub>2</sub>O) is a significant anthropogenic greenhouse gas emitted from biological nutrient removal (BNR) processes. This study tries to get a deeper insight into N<sub>2</sub>O emissions from denitrifying fluidized bed bioreactors (DFBBRs) and its relationship to the biofilm thickness, diffusivity, and reaction rates. The DFBBR was operated at two different organic and nitrogen loading rates of 5.9–7 kg COD/(m<sup>3</sup> d) and 1.2–2 kg N/(m<sup>3</sup> d), respectively. Results showed that the N<sub>2</sub>O conversion rate from the DFBBR at a biofilm thickness of 680 μm was 0.53% of the total influent nitrogen loading while at the limited COD and a biofilm thickness of 230 μm, the N<sub>2</sub>O conversion rate increased by 196–1.57% of the influent nitrogen loading concomitant with a sevenfold increase in liquid nitrite concentration. Comparing the N<sub>2</sub>O emissions at different biofilm thickness showed that the N<sub>2</sub>O emission decreased exponentially with biofilm thickness due to the retention of slow growth denitrifiers and the limited diffusivity of N<sub>2</sub>O.

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## 1. Introduction

Anthropogenic greenhouse gases (AnGHGs) emissions are globally recognized by the United Nation framework Convention on Climate Change (UNFCCC) (UNFCCC, 2007). GHGs include carbon dioxide (CO<sub>2</sub>), methane (CH<sub>4</sub>), and nitrous oxide (N<sub>2</sub>O). N<sub>2</sub>O, the dominant ozone-depleting substance, is the third most important GHGs with a global warming potential (GWP) of 310 times that of CO<sub>2</sub> (IPCC, 2007; Ravishankara et al., 2009). Wastewater treatment plants (WWTP) employing biological nutrient removal (BNR) including nitrification and denitrification are an important anthropogenic source of N<sub>2</sub>O emissions and are estimated to account for 3.2% of the global anthropogenic N<sub>2</sub>O emission (Sahely et al., 2006; Kampschreur et al., 2009). Considering the widespread use of BNR processes due to the rigorous effluent water quality standards, there is a potential that N<sub>2</sub>O emissions from WWTP will increase (Ahn et al., 2010; Park et al., 2000).

During nitrification, N<sub>2</sub>O can be produced through the aerobic hydroxylamine oxidation by ammonia oxidizing bacteria

(AOB). This hydroxylamine, generated during ammonia oxidation, is oxidized to NO directly under the catalysis of hydroxylamine oxidoreductase (HAO) encoded by the *haoAB* genes, and reduced to N<sub>2</sub>O under the catalysis of *c554* cytochrome (Cyt *c554*) (Chandran et al., 2011). N<sub>2</sub>O also can be produced through chemical decomposition of intermediate from the oxidation of NH<sub>4</sub> to NO<sub>2</sub> (Ritchie and Nicholas, 1972).

Generally, conditions that are conducive to incomplete denitrification and consequently accumulation of nitrites in the liquid phase, such as limited carbon, high dissolved oxygen (DO), competition for carbon by other microbial groups, and cold temperatures result in increased N<sub>2</sub>O emissions. All the above parameters differ significantly in biofilm processes as compared with suspended growth systems. During denitrification, N<sub>2</sub>O is one of the obligatory intermediates in the biochemical reaction by heterotrophic denitrifying bacteria (HDN) where NO<sub>3</sub>-N is reduced to NO<sub>2</sub>-N, NO<sub>2</sub> and N<sub>2</sub>O, with N<sub>2</sub>O finally reduced to N<sub>2</sub> gas (Hu et al., 2013). The reduction of NO<sub>3</sub> to N<sub>2</sub> involves six enzymes and reductase using five electrons (Desloover et al., 2012; Stein, 2011). First, NO<sub>3</sub> is reduced to NO<sub>2</sub> by periplasmic nitrate reductase (NAP) and membrane-bound nitrate reductase (NAR). Second, the reduction of NO<sub>2</sub> to NO involving Cu-containing nitrite reductase (NirK) and cytochrome *cd1* nitrite reductase (NirS) occurs. Using the nitric oxide reductase (NOR), the NO is reduced to N<sub>2</sub>O. Finally, N<sub>2</sub>O is reduced to N<sub>2</sub> with help of nitrous oxide reductase (NOS) as shown in Fig. 1a (Desloover et al., 2012).

\* Corresponding author at: Department of Chemical and Biochemical Engineering, The University of Western Ontario, London, Ontario, Canada N6A 5B9.  
Tel.: +1 519 661 2111x85470; fax: +1 519 850 2921.

E-mail address: [gnakhla@eng.uwo.ca](mailto:gnakhla@eng.uwo.ca) (G. Nakhla).

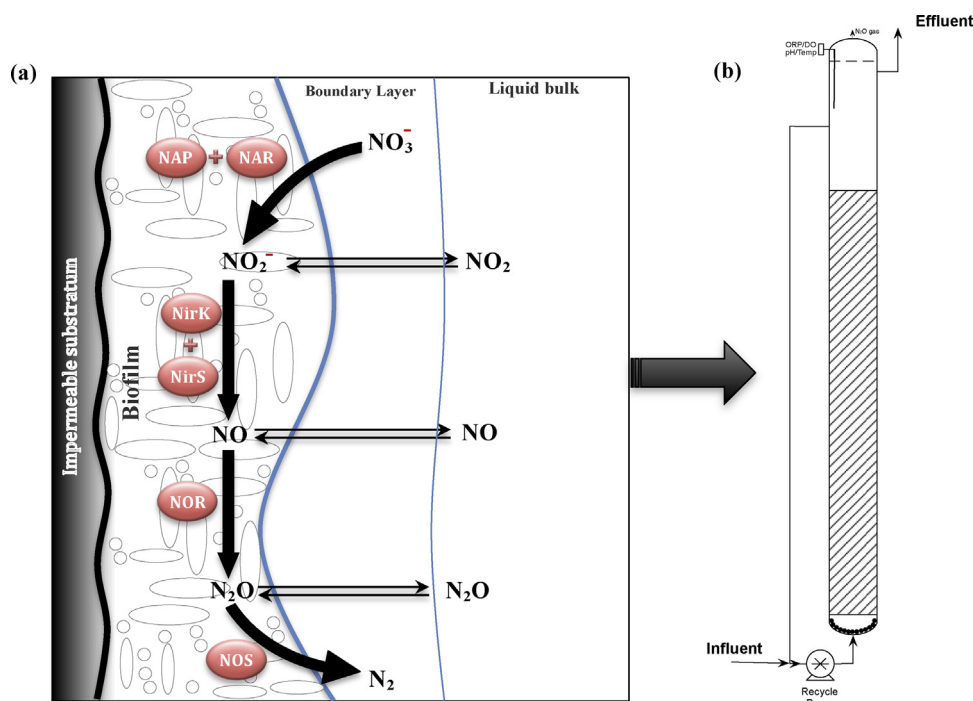


Fig. 1. (a) Conceptual overview of the  $N_2O$  production and consumption pathways of biofilm during denitrification process and (b) schematic diagram of DFBBRs.

During the denitrification process, the emissions of  $N_2O$  from BNR processes have been reported to be reduced by the changing the operation conditions in a step-feed sequencing batch reactor (SBR) and maximizing the anoxic time by decreasing the aerobic carbon breakdown to increase the carbon-to-nitrogen ratio (Hu et al., 2011). Additionally, Park et al. (2007) reported that the addition of immobilized *Alcaligenes faecalis* to the intermittent aeration-activated sludge process would enhance denitrification and mitigate the emissions of  $N_2O$ . Furthermore, Manconi et al. (2006) found that the addition of copper ( $Cu^{2+}$ ) in a range of 10–100  $\mu\text{g/L}$  as a catalyst increased the nitrous oxide reductase (NOS) activity and accelerated the bio-reductions of both nitrite to nitric oxide and nitrous oxide to nitrogen gas (Manconi et al., 2006).

Most of the studies on  $N_2O$  emissions during denitrification from BNR systems were conducted predominantly for suspended growth system, i.e. activated sludge system and sequencing batch reactors (SBR), and few studies have been conducted to investigate  $N_2O$  emissions from particulate bioparticles systems. Gaëlle et al. (2006) studied the  $N_2O$  emissions from biofilters during nitrification (Biostyr<sup>®</sup>) and denitrification (Biofor<sup>®</sup>) processes at a nitrogen loading rate of 2.2  $\text{kg TN/m}^3 \text{ d}$  and found that 0.4–1% of the oxidized ammonium was emitted as  $N_2O$  during nitrification and 0.2–1.3% of the nitrate removed during the denitrification stage. They found that the denitrification rates and  $N_2O$  emissions are directly correlated with the quantities of added carbon source (COD/N ratios ranged from 1.9 to 5), without any correlation to the biofilm thickness. In biofilm processes, nitrification and denitrification are maintained predominantly in the attached biomass (Eldyasti et al., 2010), which would hypothetically contribute to  $N_2O$  emissions. During nitrification,  $N_2O$  may be produced during the oxidation of hydroxylamine and/or the reduction of nitrites. As shown in Fig. 1a,  $\text{NO}_3\text{-N}$  is reduced to  $\text{NO}_2\text{-N}$  and  $\text{N}_2\text{O}$ , with  $\text{N}_2\text{O}$  finally reduced to  $\text{N}_2$  gas. Since the aforementioned processes are microbially mediated and strongly influenced by substrates diffusion in and out of the biofilm, understanding the contribution of biofilm thickness to the  $N_2O$  emissions would help reduce the  $N_2O$  emissions from biofilm processes.

Among the biological processes for the municipal and industrial wastewater, the denitrifying fluidized bed bioreactor (DFBBR) system is a promising particulate bioreactor for the biological nutrient removal (BNR) and proves to be economic and efficient, due to a large specific surface area (SSA) ranging from 2000  $\text{m}^2/\text{m}^3$  to 4000  $\text{m}^2/\text{m}^3$  which sustains a very high biomass (biofilm) concentrations of up to 40,000  $\text{mg VSS/L}$  (Eldyasti et al., 2010, 2012). As a result of much higher bioparticle density per unit reactor volume and smaller media size, the DFBBR exhibits very different bioreactor properties i.e. biofilm thickness, detachment rates, and attrition rates than conventional biofilm processes (Eldyasti et al., 2010, 2012).

Since literature studies explored  $N_2O$  emissions from bioparticles system without correlating the  $N_2O$  emissions to the biofilm thickness at high specific surface area, the primary goal of this study is to investigate  $N_2O$  emissions from a DFBBR characterized by high SSA of 2880  $\text{m}^2/\text{m}^3$  with a focus on the contribution of biofilm thicknesses to  $N_2O$  emissions from biofilm. This work also investigates the relationship of maximum reaction rates and diffusivity through denitrification biofilm and conducts a comprehensive nitrogen mass balance. This work for the first time provides an insight into  $N_2O$  emission and conversion rates in DFBBR, which will help improve the design and operation of such systems. This study also aimed at examining the relationship between the  $N_2O$  emission and nitrite concentrations during denitrification at two different carbon to nitrogen ratios. Furthermore, after the limited carbon phase, the DFBBR was tested at chemical oxygen demand (COD)-nitrogen ratio (COD/N) of 5 again for 50 days to investigate the dynamics of  $N_2O$  emissions for the particulate biofilm. Additionally,  $N_2O$  emissions from DFBBR were further compared to those in other BNR systems.

## 2. Materials and methods

### 2.1. System description

An anoxic fluidized bed bioreactor (Fig. 1b) comprising a plexiglass reactor with a total working volume of 507 ml, height of

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