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Model-based evaluation of tetracycline hydrochloride removal and mineralization in an intimately coupled photocatalysis and biodegradation reactor



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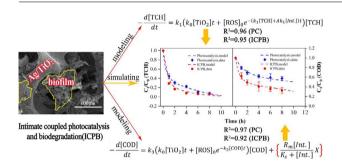
HIGHLIGHTS

- Kinetic models of intimately coupled photo-biological reaction (ICPB) were developed.
- Tetracycline (TCH) removal kinetics by ICPB were successfully validated.
- The root square errors between experimental results and model simulations were at max 2.1%.
- Biodegradation in ICPB played a crucial role in modeling TCH removal and mineralization.

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GRAPHICAL ABSTRACT



ABSTRACT

Intimately coupled photocatalysis and biodegradation (ICPB) shows great potential for treatment of refractory pollutants; however, no kinetics for modeling ICPB performance has been developed and the major challenge is to determine the relationship between photocatalysis and biodegradation. In this work, we developed a simplified kinetic model to predict removal and mineralization of a target pollutant (tetracycline hydrochloride; TCH) by hypothesizing that all of the biodegradable photocatalysis products are immediately bio-utilized. Combined with a second-order photocatalytic kinetic model and Monod-type biodegradation model, we observed the interactions between photocatalysis and biodegradation in ICPB. Parameters in the kinetic equations were estimated using the First Optimization software to fit the experimental data to the proposed model with nonlinear regression. Our experimental results showed that TCH and chemical oxygen demand (COD) removal were as high as 94% and 70% within 8 h, respectively. TCH was transformed to non-toxic intermediates in only 4 h. Significantly, the kinetic models could satisfactorily predict the TCH and COD removal, and agreed well with the experimental data with an $\mathbb{R}^2 > 0.92$. The models confirmed that biodegradation in ICPB played a major role in accelerating TCH and its intermediates removal and mineralization, as the kinetic coefficient k_1 of ICPB was 10% greater than that of photocatalysis alone. The developed models accurately predicted the ICPB efficiencies, and revealed the mechanisms of ICPB operation.

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Nomenclature		R_m	maximum specific c
		K_s	half-saturation cons
[TCH]	TCH concentration (mg L^{-1})	Х	biomass concentrati
[COD _T]	total bulk chemical oxygen demand (mg L^{-1})	RS	reactive species (mo
[COD _{TCF}] theoretical COD of the remaining TCH (mg L^{-1})	[TiO ₂]	TiO ₂ concentration
[COD _{INT}] amount of intermediates (= $[COD_T] - [COD_{TCH}]$, mg L ⁻¹)	C _{mod}	value of model calc
k _o	production rate constant of RS (h^{-1})	Cexp	value of experiment
k_1	second-order rate constant for degradation of TCH (L h^{-1}	P	protocol of photolys
	mol^{-1})	AD	protocol of adsorpti
k_2	second-order rate constant for degradation of Int. (L h^{-1}	В	protocol of biodegra
	mol ⁻¹)	PC	visible-light-induced
k_3	second-order rate constant for degradation of COD (L h^{-1}	ICPB	intimately coupled
	mol ⁻¹)		

1. Introduction

Intimately coupled photocatalysis and biodegradation (ICPB) technology shows great potential for enhancing the removal and mineralization of refractory pollutants [1-3]. In a typical ICPB system, millimeter-sized macro-porous carriers support a photocatalyst on their outer surface and a biofilm within their pores [3]. The concept is that bacteria in the carrier pores are protected from photocatalytic oxidation. This arrangement enables bacteria away from free radical damage but is close enough to rapid degradation of the biodegradable photocatalytic products. ICPB in treatment of bio-refractory pollutants overcomes the weaknesses of incomplete mineralization by photocatalysis alone or inhibition of toxic compounds by bioreactor alone [3–5]. Currently, ICPB has been successfully applied for the treatment of many refractory pollutants, such as phenols [1–3], dinitrotoluene [4] and dyes [5]. Recently, Xiong et al. [6] has used a visible-light-driven ICPB reactor to enhance TCH degradation and mineralization, with which the TCH and COD removal improved by 11% and 20% respectively by ICPB, compared to the photocatalysis alone. This was the first report of treatment of an antibiotic by ICPB. Notably, it was found that biofilm benefitted TCH degradation by reducing interaction of reactive species (RS) and TCH intermediates.

Mathematical modeling serves as a both useful and reliable tool to probe into emerging technologies, while significantly reducing the workload related to process evaluation or optimization [7]. In particular, a kinetic model can be used to: (1) identify the key factors of degradation [8–10], (2) provide improved understanding of the complicated reaction mechanisms [11–13], (3) help to scale up relevant processes, and (4) predict the treatment duration and therefore running cost [14]. However, only the biodegrading kinetics of an ICPB reactor has previously been quantified, using an Aiba self-inhibition model [15]. This model lacks the structure of the real ICPB manifestation, i.e., mutually promotion between the photocatalysis and biodegradation [3–5]. The main challenge to determining an ICPB model is to integrate both photocatalytic and biological degradation, where the immediate consumption of photocatalytic intermediates by biodegradation has to be considered.

To the best our knowledge, no mathematical model with the abovementioned real ICPB manifestation structure has been developed to date, consequently, the modeling evaluation of pollutants degradation in an ICPB remained unknown. In this work, we approached the mathematic model of ICPB for the first time, employing a strategy of hypothesizing that all the biodegradable photocatalysis products were immediately bio-utilized. This in fact constitutes an ideal ICPB reaction [6]. Modeling photocatalysis alone was achieved based on a secondorder kinetic model [6,16–18] and the biodegradation alone was simulated using the well-known Monod model [19]. Then, we developed the ICPB kinetic model by fully considering the transfer of the predicated photocatalytic products into the biodegradation models, under determination of the final removal and mineralization of pollutants.

R_m	maximum specific cell growth rate (h^{-1}) TCH phenol	
K_s	half-saturation constant (mg L^{-1})	
Х	biomass concentration at stable state (mg L^{-1})	
RS	reactive species (mol L^{-1})	
[TiO ₂]	TiO_2 concentration (mg L ⁻¹)	
C _{mod}	value of model calculated results (mg L^{-1})	
Cexp	value of experimental results (mg L^{-1})	
Р	protocol of photolysis alone	
AD	protocol of adsorption alone	
В	protocol of biodegradation alone	
PC	visible-light-induced photocatalysis alone	
ICPB	intimately coupled photocatalysis and biodegradation	

TCH was employed as model pollutant and the experimental data for TCH removal and mineralization were directly compared with the simulated results. This study is important for predicting removal and mineralization of refractory pollutants by ICPB technology, and achieving a direct scale-up and operation of ICPB reactors.

2. Material and methods

2.1. Carriers, photocatalysts and biofilms

We used a sol-gel method to prepare Ag/TiO_2 nanoparticles, which are visible-light responsive photocatalysts. Briefly, Ag/TiO_2 was prepared with titanium tetraisopropoxide [Ti(OCH(CH₃)₂)₄] (Sigma-Aldrich Co. LLC., USA) and silver nitrate (Sigma-Aldrich Co. LLC., USA). Silver reduction and doping were conducted in sodium citrate tribasic dihydrate (Beijing Chemical Works, China) according to a previous report [20].This Ag/TiO₂ sol was used for the following coating procedure.

Polyurethane sponge carriers (2 mm \times 2 mm \times 2 mm) with a porosity of 87% and wet density of 0.89 g/mL (Hayi-diverse, Yixing, China) were coated with Ag/TiO₂ by using the method of evaporation-induced self-assembly (EISA) [21]. Basically, Ag/TiO₂ sol (50 mL) was ultrasonically dispersed in ethanol (150 mL) for 5 min and then the carriers (4 g) were added to the suspension. The carriers in the Ag/TiO₂ sol mixture were further ultrasonically dispersed for another 5–10 min before drying in an oven at 80 °C by stirring with a glass rod every 30 min. After rinsing with deionized water in ultrasonic three times to remove the loosely attached Ag/TiO₂ nanoparticles, porous carriers coated with photocatalysts were ready to use. The Ag/TiO₂ coating amount was 0.39 g-catalyst/g-carrier, with a coating ratio of 28.3% (wt %).

The coated carriers were immersed in an activated sludge (South WWTP, Changchun, China) for the initial attachment of microbes and the biofilm was developed at room temperature (~ 20 °C) with a synthetic wastewater containing sodium acetate (330 mg/L), NH₄Cl (29 mg/L), Na₂HPO₄·2H₂O (8 mg/L), and NaH₂PO₄ (4 mg/L) in an internal loop airlift-driven reactor, which was described in our previous report [22]. When COD of the effluent was stabilized, the biofilm was considered to be mature.

2.2. Reactors and protocols

Photolytic circulating-bed bioreactors of a working volume of 540 mL were used to simulate TCH degradation and mineralization in ICPB reactors with biofilm developed on the Ag/TiO₂ coated sponge carriers, which was described in our previous publications [3], and schematically drawn in Fig. S1. Synthetic wastewater and biofilm carriers were circulated by bubbling air from the bottom of the reactor, supplied by a 35 W aeration pump (SOBO, Weifang, China) and controlled by a flow meter (superficial velocity in the draft tube was

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