



## Minimization of nitrous oxide emission in a pilot-scale oxidation ditch: Generation, spatial variation and microbial interpretation



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### HIGHLIGHTS

- Distinct spatial variations and low emission factor were found in oxidation ditch.
- Nitrifier denitrification contributed to 89.78% of the total N<sub>2</sub>O generation.
- Denitrifiers performing N<sub>2</sub>O reducing capabilities dominated microbial community.
- NOB abundance played a key role in the low nitrite and N<sub>2</sub>O accumulations.
- N<sub>2</sub>O emissions significantly increased when encountered ammonia or aeration shock.

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### ABSTRACT

Nitrous oxide (N<sub>2</sub>O) emission from wastewater treatment plants (WWTPs) has received increasing attention. This paper presented how N<sub>2</sub>O emission was significantly reduced in a pilot-scale Carrousel oxidation ditch under reasonable nitrification and denitrification. N<sub>2</sub>O emission from the reactor was found as low as 0.027% of influent nitrogen, which was much less than that from other processes. Further measurements on spatial variation of N<sub>2</sub>O emission in the alternative aerobic/anoxic zones with help of a series of batch experiments demonstrated that about 90% of the emission was contributed by nitrifier denitrification (ND). Moreover, the taxonomic analysis based on high through-put 16S rRNA gene sequencing revealed that the high abundance of denitrifying bacteria and nitrite-oxidizing bacteria (NOB) was responsible for low nitrite accumulations and consequent low N<sub>2</sub>O emissions. However, N<sub>2</sub>O generation would be greatly increased upon the normal operation being shocked by either ammonia overload or aeration failure of the oxidation ditch system.

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## 1. Introduction

Nitrous oxide (N<sub>2</sub>O) is the third largest greenhouse gas in the atmosphere after carbon dioxide and methane (IPCC, 2013). Apart from its significant global warming potential of 296 times (100-year period) as strong as that of carbon dioxide, N<sub>2</sub>O is currently the single most important destructive source resulting in stratospheric ozone depletion (Ravishankara et al., 2009). Atmospheric N<sub>2</sub>O concentration has increased continuously at an average rate of 0.73 ± 0.03 ppb per year during the past three decades, and anthropogenic sources were supposed to be responsible for this historic increase, as N<sub>2</sub>O emissions from natural sources were quite stable (IPCC, 2013).

N<sub>2</sub>O emission from wastewater treatment plants (WWTPs) has received increasing attentions since it was confirmed to be one significant atmospheric N<sub>2</sub>O source (Aboobakar et al., 2013; Foley et al., 2010). A survey revealed that domestic wastewater treatment contributed to 14% of the overall N<sub>2</sub>O emissions in U.S. (Ahn et al., 2010). To date, many efforts have been made to quantify N<sub>2</sub>O emissions from WWTPs but a wide range of emission factors (0.03–28.2%) have been reported (Table S1). Generally, N<sub>2</sub>O emissions measured in full-scale WWTPs was lower compared with those in laboratory studies, which was probably due to low microbial diversity facilitated by synthetic wastewater consisting of single carbon source.

N<sub>2</sub>O generation mainly occurs in biological nitrogen removal (BNR) in WWTPs and commonly comprises three pathways, the contributions of which were still controversial: (1) nitrifier denitrification (ND): N<sub>2</sub>O produced as a by-product when ammonia-oxidizing bacteria (AOB) use nitrite as an alternate electron acceptor instead of O<sub>2</sub> (Kampschreur et al., 2008; Yu et al.,

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2010); (2) heterotrophic denitrification (HD):  $N_2O$  accumulated during nitrate or nitrite reduction by heterotrophic denitrifiers due to imbalanced activities of nitrogen-reducing enzymes (Hanaki et al., 1992; Lu and Chandran, 2010); (3) Chemical reactions:  $N_2O$  produced from oxidizing  $NH_2OH$  by nitrite or  $O_2$  and decomposition from  $NOH$  or  $N_2O_2H_2$ , but this process is considered to occur rarely due to insufficient concentrations of these nitrification intermediates (Wunderlin et al., 2012).

Process design and operational parameters of WWTPs dominate the magnitude and variability of  $N_2O$  emissions. For instance, reactors achieving partial nitrification or simultaneous nitrification and denitrification (SND) always produce considerable amount of  $N_2O$ , while reactors with high recycle ratio emit low  $N_2O$  (Ahn et al., 2010; Foley et al., 2010). Several factors were generally accepted playing major roles in  $N_2O$  generation such as nitrite accumulation (Foley et al., 2010; Kampschreur et al., 2008; Zhou et al., 2008), dissolved oxygen limitation (Tallec et al., 2006; Lu and Chandran, 2010), insufficient organic carbon source (Hanaki et al., 1992; Lu and Chandran, 2010), short sludge retention time (SRT, Lotito et al., 2012), etc.

Oxidation ditches are widely employed in WWTPs due to their advantages in processing efficiency and operational cost (Liu et al., 2010). They have distinct features combining continuous stirring tank reactor (CSTR) and plug flow reactor (PFR), which have significant impact on magnitude and variation of  $N_2O$  emission. CSTR-like characteristic is demonstrated with the large internal recycle ratios of oxidation ditch (dozens to hundreds of times influent flow rate). Therefore, the high influent nitrogen concentration gets sufficient dilution and thereby the pressure on  $N_2O$  generation gets released (Foley et al., 2010; Liu et al., 2010). On the other hand, oxidation ditch demonstrates a PFR-like spatial characteristic during one cycle that the activated sludge experiences alternative aerobic and anoxic regions for nitrification and denitrification. Accordingly,  $N_2O$  emission presents a distinct spatial variation within the reactor, which also increases difficulties of precise quantification of the total emission amounts (Aboobakar et al., 2013).

To date, researches on  $N_2O$  emissions from the distinct wastewater treatment process oxidation ditch were scarcely reported. During this study, the magnitude and spatial variations of  $N_2O$  emissions were systematically monitored within a pilot-scale oxidation ditch. Furthermore,  $N_2O$  generation sources were explored with a series of batch experiments at various oxygenation levels and further interpreted by microbial community analysis based on high-throughput 16S rRNA gene sequencing. Finally, responses of the oxidation ditch to abnormal operations

like nitrogen overload shock and aeration failure shock were investigated.

## 2. Methods

### 2.1. Set-up and operation of the pilot-scale oxidation ditch

This study was conducted in a pilot-scale Carrousel oxidation ditch which was fabricated with plexiglass with working volume of  $1.4\text{ m}^3$  (Fig. 1). The reactor comprised four channels with  $1.15\text{ m}$  long,  $0.35\text{ m}$  wide and  $0.5\text{ m}$  working depth. Two surface impellers and four submerged stirrers drove the circulating flow in the channels at an average rate of  $0.01\text{ m s}^{-1}$ . A bioselector with working volume of  $50\text{ L}$  was setup to avoid sludge bulking. A clarifier of  $0.15\text{ m}^3$  was used for the separation of the activated sludge which was recycled back to the bioselector at 100% of the influent flow by a magnetic pump.

For the initial start-up, the reactor was seeded with activated sludge taken from a full-scale municipal WWTP in Langfang, China. Synthetic wastewater was pumped into the bioselector at a flow rate of  $0.1\text{ m}^3\text{ h}^{-1}$  to maintain a hydraulic retention time (HRT) of  $14\text{ h}$ . The synthetic wastewater was stored in a tank with a volume of  $1.8\text{ m}^3$  and the composition in  $1\text{ m}^3$  tap water was as follows:  $250.0\text{ g}$  sugar,  $191.1\text{ g}$   $NH_4Cl$ ,  $36.8\text{ g}$   $K_2HPO_4 \cdot 3H_2O$ ,  $12.0\text{ g}$   $MgSO_4 \cdot 7H_2O$ ,  $10.0\text{ g}$   $CaCl_2$ ,  $3.0\text{ g}$   $FeSO_4 \cdot 7H_2O$ , and  $50\text{ mL}$  trace elements solution (Zheng et al., 2014). This influent gave concentrations of chemical oxygen demand (COD) and  $NH_4^+-N$  as  $250.0\text{ mg L}^{-1}$  and  $50\text{ mg L}^{-1}$ , respectively.

Aeration was controlled at the rate of  $1.8\text{ m}^3\text{ h}^{-1}$  (an optimal value for efficient nitrogen removal) by a flowmeter and well distributed into the reactor with four  $1\text{ m}$  long microporous aerated pipes installed at the base of the second and third channels. SRT was maintained at approximately  $25\text{ d}$  by discharging  $60\text{ L}$  mixed liquor daily from the fourth channel. The temperature was maintained at  $22 \pm 2\text{ }^\circ\text{C}$  with help of heating rods and pH was around  $7.8$  without any extra control and adjustment. Influent and effluent samples were taken daily for analysis of COD,  $NH_4^+-N$ ,  $NO_2^- -N$  and  $NO_3^- -N$ .

After two-month steady-state operation, two kinds of stress tests were carried out including “ammonium overload shock” and “aeration failure shock” experiments. In the former experiment, synthetic wastewater containing  $NH_4^+-N$  concentration of  $250\text{ mg L}^{-1}$  (other compositions not changed) was pumped to the reactor at  $0\text{ h}$  and turned back to the normal influent after  $1\text{ h}$ . In the latter experiment, aeration rate was reduced to  $0.9\text{ m}^3\text{ h}^{-1}$  at time  $0\text{ h}$  and switched back to normal  $1.8\text{ m}^3\text{ h}^{-1}$  at  $2\text{ h}$ .

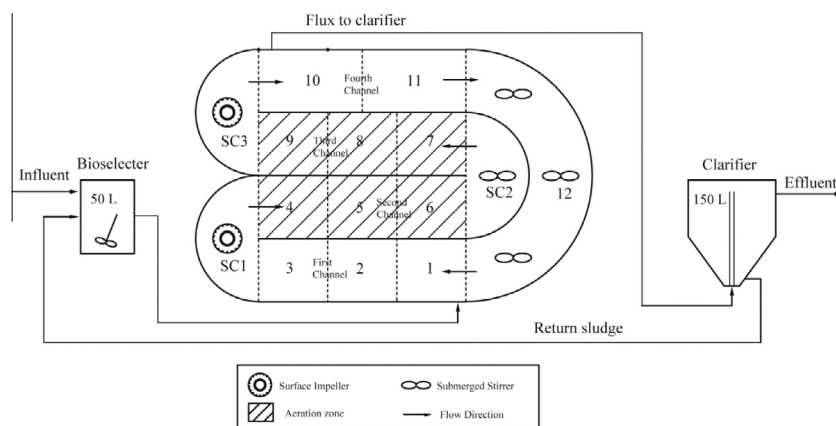


Fig. 1. The sketch of the pilot-scale Carrousel oxidation ditch and sampling zones.

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