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# Short Communication

# Sequential ethanol fermentation and anaerobic digestion increases bioenergy yields from duckweed

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duckweed, than if anaerobic digestion had been run alone.

# 1. Introduction

The economic and environmental disadvantages of fossil fuel consumption have increased the search for alternative resources to fulfill world's growing energy and chemical needs ([Jung et al., 2016](#page--1-0)). At the same time, conventional bioenergy crops have also been posing social, economic, and environmental challenges. Duckweed (Lemnaceae), a family of fast-growing, simple, floating aquatic plants, consisting of 38 species in five genera ([Les et al., 2002\)](#page--1-1), has been demonstrated to be a technically feasible alternative feedstock for bioethanol production due to several advantages: it can accumulate high amounts of starch (up to 46% of dry mass) under nutrient starvation [\(Zhao et al., 2015](#page--1-2)); has relatively little lignin content (1–3%); its small size (0.1–1 cm) eliminates the need for milling; and, because it floats, the harvesting process is relatively simple ([Cui and Cheng, 2015](#page--1-3)). Duckweeds are resilient to a broad range of nutrient concentrations; therefore, they can be grown on wastewater steams ([Cheng and Stomp, 2009\)](#page--1-4).

Due to its high and manipulatable starch content, duckweed is regarded as a promising bioethanol feedstock in the current literature. The studies conducted to date have focused on the utilization of the starch component only ([Xu et al., 2011; Yu et al., 2014\)](#page--1-5), or the fermentation of cell wall carbohydrates as well [\(Ge et al., 2012; Zhao](#page--1-6) [et al., 2014](#page--1-6)). The high level of variability in wastewater compositions, however, may cause uncertainties in starch and bioethanol potentials from wastewater-derived duckweed biomass. By comparison, a more resilient pathway for duckweed valorization could be anaerobic digestion, since this process converts not only sugars, but also proteins and lipids into biomethane. In addition, anaerobic digestion can be used to stabilize residual organics in the ethanol fermentation broth, and thereby help to compensate for the costs of ethanol production and distillation [\(Wu et al., 2015\)](#page--1-7). Indeed, the sequential process of ethanol fermentation and anaerobic digestion has been shown to increase the overall bioenergy yield of several other substrates such as food waste ([Wu et al., 2015\)](#page--1-7), oat straw [\(Dererie et al., 2011](#page--1-8)), and corn stalks (Vintilǎ [et al., 2013\)](#page--1-9). This combined approach may improve the sustainability of large-scale biorefineries.

organism ratio) of 1.0. This value was 51.2% higher than the biomethane yield of a replicate reactor with raw (non-fermented) duckweed. The combined bioethanol-biomethane process yielded 70.4% more bioenergy from

> Although some work has focused on ethanol production from duckweed, reports on its anaerobic digestibility are limited to a very few studies. An early study on anaerobic digestion of manganese-contaminated duckweed produced a maximum biogas yield of 176 ml/g with a methane content of 60% [\(Jain et al., 1992](#page--1-10)). Other work conducted on duckweed has focused on its co-digestion with other substrates, such as dairy manure [\(Triscari et al., 2009](#page--1-11)), to help balance the C/N ratio.

> To ensure that neither limitations nor inhibition will occur during anaerobic digestion due to substrate loading, the substrate-to-inoculum ratio (S/I) should be optimized [\(Chynoweth et al., 1993](#page--1-12)). The S/I not only affects total methane yield, but also its production rate ([Alzate](#page--1-13) [et al., 2012\)](#page--1-13). In the current study, the potential of increasing bioenergy yields obtained from duckweed grown in an ecological wastewater

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treatment system for nutrient removal was investigated using a sequential process: fermentation of duckweed and distillation of the resulting bioethanol, followed by anaerobic digestion of the residual fermented duckweed. In addition, the effects of S/I ratio on anaerobic digestion performance were evaluated through biochemical methane potential (BMP) assays.

### 2. Materials and methods

# 2.1. Analytical methods

Total solids (TS), volatile solids (VS), total suspended solids (TSS), volatile suspended solids (VSS), and volatile dissolved solids (VDS) were determined according to Standard Methods No. 2540 [\(APHA/](#page--1-14) [AWWA/WEF, 2012](#page--1-14)). The suspended portion of samples was separated on glass fiber filters (AP40; Millipore, Billerica, MA, USA) using a vacuum filtration apparatus. Chemical Oxygen Demand (COD) was measured according to the closed reflux colorimetric method as described in Standard Methods, No. 5220 [\(APHA/AWWA/WEF, 2012](#page--1-14)).

Glucose and ethanol quantification were performed using a Waters high performance liquid chromatograph (HPLC) equipped with a refractive index detector (Waters, Milford, MA) and a Bio-Rad Aminex HPX-87H column (300 mm  $\times$  7.8 mm; Bio-Rad, Richmond, CA) with 0.8 ml/min of 0.012 N sulfuric acid as the mobile phase. The detector and column temperatures were constant at 35 °C and 65 °C, respectively. Prior to HPLC analysis, samples were centrifuged at 4 °C for 20 min at  $5200 \times g$  and the supernatant filtered through 0.2 μm nylon syringe filters. Theoretical maximum glucose concentration was calculated according to [Gulati et al. \(1996\).](#page--1-15)

Headspace gas volumes of anaerobic reactors were measured at 25 °C using a water displacement device filled with 0.01 M hydrochloric acid to prevent microbial growth. Volume readings were reported at standard temperature and pressure. Volumetric methane concentrations were determined by withdrawing headspace from the reactors using a 250 μL airtight syringe (Hamilton, Reno, NV, USA) and injecting into a gas chromatograph (model SRI310C, SRI Instruments, Torrance, CA, USA) equipped with a 6 foot molecular sieve column (Altech, 5605PC, MD) held at 80 ◦C.

#### 2.2. Plant material and cultivation

Duckweed used in this study was obtained on May 27, 2015, from two sources: 1) an open tank dedicated for growing duckweed in the Penn State Eco-Machine™ (EM), which is a pilot-scale ecological wastewater treatment system receiving on average ( $n = 4$ ) 3.6  $\pm$  1.1 mg/L phosphate,  $0.1 \pm 0.0$  mg/L ammonia, and  $11.1 \pm 3.0$  mg/L nitrate; and 2) an open pond within the effluent spray fields of the Penn State Wastewater Treatment Plant, a.k.a. the "Living-Filter" (LF), receiving on average (n = 3) 2.2  $\pm$  0.4 mg/L phosphate, 2.3  $\pm$  0.9 mg/L ammonia, and 7.8  $\pm$  0.8 mg/L nitrate. In both sources, duckweed was naturally present and had not been subjected to a frequent harvesting regime.

To identify the duckweed species present in each source, total DNA was extracted from duckweed tissue using a PowerPlant® Pro DNA isolation kit (QIAGEN, Hilden, Germany), and then amplified using a two-barcode PCR protocol (Borisjuk et al., 2014). After amplification, the DNA fragments were purified using a GeneJET PCR purification kit (ThermoFisher, Waltham, MA), and sent to the Genomics Core Facility (The Pennsylvania State University) for processing. Following a BLASTbased protocol for duckweed species identification ([Borisjuk et al.,](#page--1-16) [2015\)](#page--1-16), the EM duckweed was identified as a co-culture of Lemna japonica/minor (100% sequence identity to accession numbers KJ9211760.1 and DQ400350.1, respectively, in the NCBI database) and Wolffia columbiana (99.6% sequence identity to accession number GU454371.1); whereas the LF duckweed was identified as a monoculture of Lemna obscura (100% sequence identity to accession number GU454331.1).

For use in these experiments, harvested duckweed was rinsed with tap water and dried at 50  $\pm$  2°C to a constant weight over two days. The composition of the dried duckweed was determined by first grinding and sieving through mesh No. 20 (850 mm opening size), and then sending to Dairy One Wet Chemistry Laboratory (Ithaca, NY). The composition of EM duckweed was reported as 16.9% cellulose, 23.9% hemicellulose, 4.3% starch, 2.0% lignin, 26.0% crude protein, and 0.73 g VS per g TS. The composition of LF duckweed was reported as 17.0% cellulose, 18.1% hemicellulose, 15.9% starch, 1.1% lignin, 17.0% crude protein, and 0.81 g VS per g TS.

## 2.3. Inocula

#### 2.3.1. Yeast strain

For fermentation of duckweed, Saccharomyces cerevisiae (ATCC 24859) was enriched in culture medium with the following constituents (concentrations in parentheses are g/L): glucose (20); yeast extract (Difco, Sparks, MD) (6); CaCl<sub>2</sub>:2H<sub>2</sub>O (0.3); (NH<sub>4</sub>)<sub>2</sub>SO<sub>2</sub> (4); MgSO<sub>4</sub>:7H<sub>2</sub>O (1); and  $KH_2PO_4$  (1.5). The culture was grown at 30 °C for 24 h before being transferred to fermentation flasks as the inoculum.

#### 2.3.2. Anaerobic seed

Anaerobic seed was obtained from the Penn State Wastewater Treatment Plant secondary anaerobic digester. The inoculum was starved for two days prior to use in the BMP assays. The TS of the starved seed was 23.9  $\pm$  0.5 g/L, and the VS was 15.7  $\pm$  0.7 g/L, which is  $65.8 \pm 5.1\%$  of the TS.

#### 2.4. Fermentation experiments

Enzymatic saccharification of the duckweed was performed in 500 ml flasks with 200 ml distilled water and 10 g duckweed (dry weight). The pH was adjusted to 7.0  $\pm$  0.1 with 2 M hydrochloric acid prior to liquefaction by autoclaving at 95 °C under 103 kPa for 1 h. Flasks with EM and LF duckweed received 0.6 ml and 1.98 ml of α–amylase (Sigma Aldrich, A3403, USA) respectively, based on the starch content of each duckweed type, to achieve an amylase loading of 5000 units/g starch. Following liquefaction, the pH was adjusted to  $4.8 \pm 0.1$  with glacial acetic acid. After pH adjustment, 60 mg and 198 mg glucoamylase (Sigma Aldrich 10115, USA) were added to each flask containing EM and LF duckweed, respectively. In addition, all flasks received 2 ml cellulase (60 filter paper unit/g cellulose). Saccharification was then performed at 50 °C, while mixing at 120 rpm for 24 h in flasks sealed with cotton stoppers and parafilm. All experiments were conducted in triplicate under sterile conditions.

Following saccharification, the pH of each flask was increased to 7.0  $\pm$  0.1 by dosing with 2 M sodium hydroxide, and then 2 ml yeast culture was added. Flasks were incubated at 30 °C while mixing at 120 rpm for 48 h. Glucose and ethanol concentrations before and after fermentation were quantified. Fermented ethanol was then evaporated by vacuum extraction after the pH was increased to 7.8  $\pm$  0.1 by 2 M sodium hydroxide addition, in order to avoid escape of volatile fatty acids (VFAs) from the slurry. The triplicates for each duckweed type were then combined and subjected to BMP assays.

## 2.5. Biochemical methane potential (BMP) assays

The BMP assays with duckweed were carried out based on the protocol proposed for bioenergy crops and organic wastes ([Angelidaki](#page--1-17) [et al., 2009\)](#page--1-17) with slight modifications. Batch reactors (160 ml total volume, 120 ml working volume) were filled with 24 ml inoculum, and substrate (either raw EM or LF duckweed, or residual fermentation slurries, FEM or FLF), to provide an S/I of 0.5 or 1.0. To account for the effect of endogenous gas production by the anaerobic inoculum, control bottles were prepared with the same amount of anaerobic seed, but without substrate. Blank bottles were prepared with duckweed, but Download English Version:

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