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# Q5 Impairment of the cortisol stress response mediated by the hypothalamus–pituitary–interrenal (HPI) axis in zebrafish (*Danio rerio*) exposed to monocrotophos pesticide

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## ABSTRACT

In teleosts, an important component of the stress response is coordinated by the hypothalamic–pituitary–interrenal (HPI) axis. Environmental contaminants might disrupt the stress axis and consequently affect the stress response in fish. To investigate the effect of monocrotophos (MCP) pesticide on the stress response of fish and its potential mechanisms, adult zebrafish (*Danio rerio*) were exposed to 0, 1, 10, and 100 µg/L of a 40% MCP-based pesticide for 21 d, after which time fish were subjected to a 3-min air-exposure stressor. Concentrations of the whole-body cortisol were measured by radioimmunoassay and abundances of transcripts of proteins involved in the HPI axis were determined using quantitative real-time PCR. Results showed that 100 µg/L of MCP pesticide decreased whole-body cortisol levels of female zebrafish in response to an acute stressor, but without any effect on the cortisol response in males. 100 µg/L MCP pesticide reduced POMC and GR expression in the brain, MC2R and P450<sub>11β</sub> expression in the head kidney, but enhanced 20β-HSD2 expression in the head kidney, suggesting that MCP damaged the HPI axis involving acting at pituitary regulatory levels, inhibiting cortisol synthesis and stimulating cortisol catabolism, or disturbing the negative feedback regulation. Additionally, MCP depressed liver GR transcription but did not affect phosphoenolpyruvate carboxykinase and tyrosine aminotransferase expression in zebrafish, suggesting a role for this pesticide in reducing target tissue responsiveness to cortisol. Considered together, the reduced ability to elevate cortisol levels in response to an acute stress may be an endocrine dysfunction occurring in zebrafish subchronically exposed to MCP pesticide.

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## 1. Introduction

The stress response is a highly conserved adaptation among vertebrates to maintain internal homeostasis in a variable environment (Ings et al., 2011; Tort, 2011). In teleosts, an important component of the stress response is regulated by the hypothalamic–pituitary–interrenal (HPI) axis, which controls the circulating levels of glucocorticoid cortisol (Wendelaar Bonga, 1997; Flik et al., 2006). In response to a stressor, corticosteroid-releasing factor (CRF) is released from the hypothalamus and stimulates the anterior pituitary to release the pro-opiomelanocortin (POMC)-derived adrenocorticotrophic hormone (ACTH) (Huising et al., 2004; Metz et al., 2004). ACTH binds to the

melanocortin 2 receptor (MC2R) on the steroidogenic interrenal cells of the head kidney, and activates the corticosteroid biosynthesis (Aluru and Vijayan, 2008). Several proteins/enzymes are involved in the conversion of cholesterol to cortisol in the interrenal cells including steroid acute regulatory protein (StAR) and 11β-hydroxylase (P450<sub>11β</sub>), while limiting the activity of cortisol involves the enzyme 11β-hydroxysteroid dehydrogenase 2 (11β-HSD2) and 20β-hydroxysteroid dehydrogenase 2 (20β-HSD2). Circulating cortisol levels are also tightly regulated by a negative feedback loop, including glucocorticoid receptor (GR) signaling in the brain, inhibiting the release of trophic hormones (CRF and/or ACTH) in response to elevated steroid levels (Wendelaar Bonga, 1997; Mommsen et al., 1999).

Any adverse effect on the functioning of the HPI axis would compromise the ability of fish to mount an adequate response to stressors (Hontela, 1998; Mommsen et al., 1999) and studies have shown that fish exposed to chronic chemical stressors display a decreased corticosteroid response to a second acute stressor. For instance, Benguira et al. (2002) reported that treatment with *o,p'*-dichlorodiphenyldichloroethane for 14 days reduced the normal capacity of rainbow trout (*Oncorhynchus mykiss*) to elevate plasma cortisol in response to an acute stress. Monocrotophos (MCP; CAS number, 6923–22–4), an organophosphorus pesticide, has been shown to exert endocrine-disrupting effects in teleost

**Abbreviations:** HPI, hypothalamic–pituitary–interrenal; CRF, corticosteroid-releasing factor; POMC, pro-opiomelanocortin; ACTH, adren-corticotrophic hormone; MC2R, melanocortin 2 receptor; StAR, steroid acute regulatory protein; P450<sub>11β</sub>, 11β-hydroxylase; 11β-HSD2, 11β-hydroxysteroid dehydrogenase 2; 20β-HSD2, 20β-hydroxysteroid dehydrogenase 2; GR, glucocorticoid receptor; MCP, monocrotophos; PEPCK, phosphoenolpyruvate carboxykinase; TAT, tyrosine aminotransferase; RIA, radioimmunoassay; 11β-HSD3, 11β-hydroxysteroid dehydrogenase 3.

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Q8 fish (Tian et al., 2010; Zhang et al., 2013). Our recent study showed that  
 73 after 21 days of exposure to the MCP pesticide, concentrations of cortisol  
 74 were significantly decreased in plasma of female goldfish (*Carassius*  
 Q9 *auratus*) (Zhang et al., 2014). However, it is still unclear whether sub-  
 76 chronic exposure to MCP pesticide would affect the integrated stress  
 77 response to a secondary stressor.

78 During stress, elevated cortisol levels mediate crucial physiological  
 79 processes such as glucose plasma concentrations and various other  
 80 stress responses for coping with stress. In hepatocytes, glucocorticoid  
 81 hormones facilitate the process of gluconeogenesis by transcriptional  
 82 activation of gluconeogenesis enzymes including phosphoenolpyruvate  
 83 carboxykinase (PEPCK) and tyrosine aminotransferase (TAT). GR signal-  
 84 ing is thought to play a key role in mediating the stress effects of cortisol.  
 85 Cortisol up-regulates mRNA abundance of hepatic PEPCK mediated via  
 86 the GR, resulting in increased glucose production by hepatocytes  
 87 (Aluru and Vijayan, 2007). Joshi and Rajini (2012) demonstrated a 2-  
 88 hour MCP exposure induced increase in plasma corticosterone levels  
 89 with concomitant increase in blood glucose and activities of liver TAT in  
 90 rats (*Rattus norvegicus*). Accordingly, it is speculated certain stressor-  
 91 mediated liver metabolic capacity for coping with stress might also be  
 92 disturbed by the MCP pesticide.

93 The objective of this study was first to investigate the influence of a  
 94 21-d MCP exposure on the stress responsiveness of adult zebrafish by  
 95 analyzing whole-body cortisol levels as indicators of stress. By including  
 96 a second air exposure stressor after exposure of MCP, we could test if  
 97 MCP-exposure caused a dysfunctional response in the established  
 98 pathways, implying that fish would be incapable of responding to  
 99 an additional stressor when in a MCP-polluted environment. Second,  
 100 gene expression of CRF, POMC, and GR in the brain and MC2R, StAR,  
 101 P450<sub>11β</sub>, 11β-HSD2, and 20β-HSD2 in the interrenal tissue was  
 102 determined to investigate possible mode of actions of MCP on the  
 103 HPI axis. At last, liver GR, PEPCK, and TAT transcripts were used as  
 104 markers of target tissue cortisol responsiveness.

## 105 2. Materials and methods

### 106 2.1. Animals

107 Sexually mature Tu wild-type zebrafish (*Danio rerio*) of both sexes  
 108 (8–10 months-old;  $0.33 \pm 0.04$  g wet weight and  $2.73 \pm 0.07$  cm stan-  
 109 dard length for males,  $0.63 \pm 0.05$  g wet weight and  $2.68 \pm 0.09$  cm  
 110 standard length for females) were obtained from a local commercial  
 111 distributor and housed at a density of up to five animals per liter in  
 112 50 L tanks with dechlorinated tap water. Fish were continuously aerated  
 113 ( $7.0 \pm 1.0$  mg O<sub>2</sub>/L) at  $28 \pm 1$  °C, under a 14/10 h light/dark controlled  
 114 photoperiod. Fish were fed twice a day with freshly hatched brine  
 115 shrimp (*Artemia nauplii*) and were maintained healthy and free of any  
 116 signs of disease according to the National Institute of Health Guidelines  
 117 for Handling and Care of Experimental Animals. The animal utilization  
 118 protocol was approved by the Institutional Animal Care and Use Com-  
 119 mittee of the Ocean University of China.

### 120 2.2. Experimental design

#### 121 2.2.1. MCP pesticide exposure

122 MCP pesticide (3-hydroxyl-*N*-methyl-*cis*-crotonamidedimethyl  
 123 phosphate, 40% water-soluble preparation) was purchased from the  
 124 Qingdao pesticide factory in China. The concentration on the label was  
 125 40%, which was consistent with the actual concentration determined  
 126 by gas chromatography ( $40 \pm 0.1\%$ ) (Ru et al., 2003). The half-life of  
 127 MCP is approximately 66 days at pH 7.0 and 20 °C (Wang and Zhang,  
 128 1989).

129 Fish were acclimated for at least two weeks before the experiments,  
 130 then sexes were kept separate in 5 L glass beaker containing 2.5 L  
 131 dechlorinated tap water. Eight male (or female) fish were assigned to  
 132 each glass beaker and exposed to MCP pesticide (1, 10, and 100 µg/L)

dissolved in water for 21 d. Triplicate beakers were used for each treat- 133  
 ment group and two control groups with dechlorinated tap water (to 134  
 one of which no air-exposure stress would be administered) were 135  
 included in the exposure design. Fish exposure were conducted as a 136  
 semi-static toxicity test (all water/MCP pesticide of the beaker was 137  
 changed after each 24 h of exposure in order to keep the treatment 138  
 concentrations constant and water fresh) and other conditions were 139  
 the same as those described in Sect. 2.1. Fish were fed with freshly 140  
 hatched brine shrimp twice daily. A total of 240 fish (including 120 141  
 males and 120 females) were used for this study and no deaths were 142  
 observed in any of the treatment groups during experimentation. 143

#### 144 2.2.2. Air exposure

No food was provided 24 h before the end of the 21-d MCP exposure. 145  
 At 22nd day according to Ramsay et al. (2009), both the fish and the 146  
 water contained in the beaker were poured into a net, leaving the fish 147  
 suspended in the air for 3 min. Then all fish were immediately returned 148  
 to a prepared 50-L tank containing 30 L dechlorinated tap water with 149  
 the same temperature, allowed to swim freely in the tank, and sampled 150  
 at 3 min post-air exposure-stress. All sampling commenced at approxi- 151  
 mately the same time of day (9:00–10:00 h) in order to minimize the 152  
 possibility of fluctuations in cortisol due to natural circadian rhythms 153  
 (Schreck, 1981). 154

#### 155 2.2.3. Sampling procedures

Upon sampling, fish were euthanized with a lethal dose of MS-222 156  
 (500 mg/L), blotted on paper towels to remove excess water, and then 157  
 weighed. One subgroup of six fish per sex from each treatment group 158  
 was immediately frozen in liquid nitrogen, and stored at  $-80$  °C for 159  
 whole-body cortisol analysis. Tissues of the rest of the fish including 160  
 brain, liver, and head kidney were removed, 3 brains (livers or head 161  
 kidneys) were pooled to obtain 6 samples per treatment, placed in 162  
 RNase-free tubes, and stored at  $-80$  °C for RNA extraction. 163

#### 164 2.2.4. Whole-body cortisol extraction and measurement

Whole-body cortisol was extracted using a modification of a method 165  
 described by Sink et al. (2007). Individual whole, frozen zebrafish were 166  
 thawed, homogenized, and extracted twice with 2 mL diethyl ether. 167  
 After freezing the solution with liquid nitrogen immediately, the unfro- 168  
 zen portion (ethyl ether containing cortisol) was decanted. The ethyl 169  
 ether was transferred to a new tube and completely evaporated under 170  
 a gentle stream of nitrogen for 2 h, yielding a lipid extract containing 171  
 the cortisol. The extract was reconstituted with 0.3 mL phosphate- 172  
 buffered saline with 1% BSA, vortexed for 30 s, and stored at  $-20$  °C 173  
 until the radioimmunoassay (RIA) was conducted. 174

Cortisol in the whole-body extracts was measured using an RIA 175  
 using commercially available kits, following the protocols provided by 176  
 the manufacturer (Beijing North Institute of Biological Technology, 177  
 Beijing, China). The RIA kit for human cortisol was validated for use 178  
 with zebrafish samples by demonstrating parallelism between a series 179  
 of diluted and spiked samples in relation to the standard curve attached 180  
 to the assay kits. Standards and samples were added to the test tubes in 181  
 duplicate. The assay detection limits were 2 ng/mL and the inter- and 182  
 intra-assay coefficients of variation were <15% and <10%, respectively. 183

### 184 2.3. Gene expression analysis

Total RNA was isolated from the liver, brain, and head kidney from a 185  
 pool of three fish using the phenolic reagent TRIzol (Invitrogen, Carlsbad, 186  
 CA, USA) according to the manufacturer's protocol, and then was mea- 187  
 sured by spectrometry at OD<sub>260/280</sub>. Equal amounts of RNA (1 µg) were 188  
 reverse-transcribed into cDNA using the PrimerScript RT reagent Kit 189  
 with gDNA Eraser (Perfect Real Time) as described by the manufacturer 190  
 (Takara, Dalian, China). Separate samples were identically treated with- 191  
 out the addition of reverse transcriptase (PrimeScript RT Enzyme Mix 192

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