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A simple approach to fabricate granular adsorbent for adsorption of rare elements

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ABSTRACT

A kind of granular hybrid hydrogel was prepared under an ambient temperature in air atmosphere using Fenton reagent as the redox initiator, and its three-dimensional structured polymeric network can be formed by the grafting reaction of acrylic acid (AA) onto hydroxypropyl cellulose (HPC) with attapulgite (APT) as the inorganic component. The resulting granular hybrid hydrogel was marked as HPC-g-PAA/APT and used as the adsorbent to remove the rare earth elements, La(III) and Ce(III). The effects of pH and APT content on the adsorption capacity, as well as the adsorption isotherms and kinetics, were systematically investigated. Finally, the reusability of HPC-g-PAA/APT for La(III) and Ce(III) were evaluated. The results indicate that the adsorption process is pH-independent at pH \ge 4.0 and can be described using the pseudo-second-order kinetic model. The equilibrium isotherm matches well with the Langmuir model. The adsorbed La(III) and Ce(III). After five adsorption–desorption cycles, the adsorption capacity shows a slight decrease (about 15%), implying that the granular hybrid hydrogel can be used as an effective adsorbent for the removal and recovery of La(III) and Ce(III) from aqueous solution.

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1. Introduction

Rare earth elements (REEs), including the lanthanide series 23 from the atomic number 57 to 71 and the other two elements Sc 24 25 and Y, have a similar atomic structure and chemical and physical properties [1]. Rely upon their chemical, catalytic, electrical, 26 magnetic, and optical properties, REEs have been applied in many 27 high-tech fields, such as chemical engineering, information storage, 28 energy conservation and nuclear energy [2-8]. In recent years, with 29 the ever-increasing demand for high-purity rare earths and their 30 compounds, the separation and purification of REEs have gained 31 considerable attention. Up to now, many methods have been used 32 to separation, purification and preconcentration of REEs, such as co-33 precipitation [9], solvent extraction [10,11], and ion-exchange [12]. 34 However, these traditional methods have some disadvantages, such 35 as high consumption of reagent and energy, low selectivity, high 36 operational cost and generation of secondary metabolites [13]. In 37 comparison with conventional technologies, adsorption is one of 38 the most attractive methods and has been used for the concentra-39 tion and recovery of metal ion [14-17]. 40

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http://dx.doi.org/10.1016/j.ijbiomac.2014.08.039 0141-8130/© 2014 Published by Elsevier B.V. Recently, the adsorbents based on natural polymers have drawn wide attention, because of the features of rich material sources, low cost, non-toxicity and biodegradability, such as alginate [18], chitin [19], chitosan [20], β -cyclodextrin [21], cellulose [22], starch [23], and gum acacia [24]. Hydroxypropyl cellulose (HPC) is an ether of cellulose in which some of the hydroxyl groups in the repeating glucose units have been hydroxypropylated forming $-OCH_2CH(OH)CH_3$ groups using propylene oxide. It has been widely used in the tissue engineering [25] and drug release [26,27]. The HPC has ether bond (-C-O-C-) and hydroxyl (-OH) groups, which can serve as the coordination sites to form complexes with various heavy metal ions, so HPC can be used as a potential adsorbent material.

In this paper, HPC was used as the backbone to graft poly (acrylic acid) to form a granular hybrid hydrogel via a facile fabrication at room temperature using Fenton reagent as the redox initiator. In order to lower the cost of water treatment and improve the gel strength, attapulgite, a crystalline hydrated magnesium silicate with a fibrous morphology, large specific surface area and moderate cation exchange capacity, was introduced in the hydrogel matrix. The resulting hybrid hydrogel was evaluated the adsorption ability for two common rare earth elements La(III) and Ce(III), which advanced new materials such as super alloys, catalysts, special ceramics and organic synthesis [28,29]. The effects of pH and

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APT content on the adsorption capacity were studied. The adsorption isotherms and adsorption kinetics of HPC-g-PAA/APT for La(III)
and Ce(III) were investigated. Furthermore, the reusability of HPC-g-PAA/APT for adsorption of La(III) and Ce(III) were also evaluated.

9 2. Experimental

70 2.1. Materials

Acrylic acid (AA, chemically pure) was purchased from Shanghai 71 Shanpu Chemical Factory, China, and used without further treat-72 ment. Hydroxypropyl cellulose (HPC, the average molecular weight 73 is 1.0×10^4) was received from Amatek Chemical Technology 74 Co., Ltd., China. N,N-methylenebisacrylamide (MBA, chemically 75 pure) was received from Shanghai Yuanfan additives plant, China. 76 Ammonium iron (II) sulfate hexahydrate $((NH_4)_2Fe(SO_4)_2\cdot 6H_2O_1)$ 77 analytical grade), hydrogen peroxide solution (H₂O₂, analytical 78 grade), cerium (III) nitrate hexahydrate (Ce(NO₃)₃.6H₂O, analytical 79 grade) and lanthanum nitrate hexahydrate (La(NO₃)₃·6H₂O, ana-80 lytical grade) were all supplied by Sinopharm Chemical Reagent 81 Co., Ltd., China. Natural attapulgite (APT, industrial grade, the R&D 82 Center of Xuyi Attapulgite Applied Technology, Lanzhou Institute of 83 84 Chemical Physics, CAS) was milled and sieved through a 200-mesh screen prior to use. Other reagents used were all analytical grade 85 and all solutions were prepared with distilled water.

87 2.2. Preparation of the granular adsorbent

A series of granular hydrogels was prepared according to fol-88 lowing procedure: 3.6 g AA, 0.15 g MBA and 45 mL distilled water 89 were added in a 250 mL three-neck flask. The mixture was stirred 90 at 500 rpm for 0.5 h. Afterward, 0.3 g HPC was added in the flask 91 and kept stirring for 4 h until HPC was completely dissolved. After-92 ward, a certain amount of APT (0.18g for 1:20 to AA, 0.36g for 97 1:10 to AA, 0.6 g for 1:6 to AA, 1.2 g for 1:3 to AA and 1.8 g for 1:2 0/ to AA) was added and stirred for 60 min. Later, a given mass of $(NH_4)_2$ Fe $(SO_4)_2$ ·6H₂O and H₂O₂ (3%) were all added. In just a few minutes, the reaction solution changed the color from colorless to 97 brown, hereafter, the granular hydrogel was observed. And then the reaction was kept for another 1 h at room temperature. When 100 the reaction was finished, the granular hydrogel was sieved and immersed in 2.0 mol/L NaOH solution to neutral pH (approximately 101 60% neutralization degree according to the amount of AA), then the 102 industrial alcohol was added to dehydrate. The resulting granular 103 104 hydrogel was dried at room temperature. The blank sample without APT was named as HPC-g-PAA/APT (AA/APT = 1/0), and the other 105 adsorbent was denoted as HPC-g-PAA/APT (AA/APT = a/1) (a = 20, 106 10, 6, 3, 2). All hydrogel samples used in the adsorption experiment 107 was ground and had a particle size in the range of 40-80 mesh. 108

109 2.3. Characterization

The conformation of granular hydrogel was analyzed by FTIR
spectra, which were determined by a Nicolet NEXUS FTIR spectrom eter in the wave number region of 4000–400 cm⁻¹ using KBr pellets.
The morphologies of the samples were observed using a JSM-6701F
Field Emission Scanning Electron Microscope (JEOL) after coating
the sample with gold film.

116 2.4. Batch adsorption studies

The stock La(III) and Ce(III) solution with a concentration of 1000 mg/L was prepared by dissolving an appropriate amount of La(NO₃)₃ or Ce(NO₃)₃ in distilled water. The different concentrations of La(III) and Ce(III) for adsorption study were all prepared by dilution of the stock solution. All the adsorption experiments were preformed with 25 mg of adsorbent and 25 mL La(III) or Ce(III) solution at desired concentration. The mixture was shaken with a constant speed (120 rpm) and constant temperature (30 °C) in a thermostatic orbital shaker (THZ-98A) for a given time, and then the adsorbents were filtrated. Afterward, the concentration of La(III) or Ce(III) in the solution was determined by UV-visible spectrophotometer (UV-3010, HITACHI), using Chlorophosphonazo mA (CPA-mA) as the complexing agent. The La(III) or Ce(III) uptake was calculated from the following equation:

$$Q_e = \frac{(C_0 - C_e) \times V}{m} \tag{1}$$

where Q_e is the removal amount of La(III) or Ce(III) at any time (mg/g), C_0 and C_e are the initial and final concentration of La(III) or Ce(III) (mg/L), *V* is the volume of solution (mL) and *m* is the weight of adsorbent (mg).

The influence of pH on the adsorption performance was studied in a wide pH range from 3 to 7. The initial pH of 300 mg/L La(III) and Ce(III) solution was adjusted by 0.1 mol/L NaOH or HCl solutions. The mixtures were all shaken for 2 h. The kinetics study was performed with 300 mg/L La(III) or Ce(III) solution by varying the adsorption time from 5 to 120 min. To evaluate the control factors of the adsorption process, the adsorption kinetics was evaluated based on the pseudo-first-order equation and pseudo-second-order equation. The adsorption isotherms for La(III) or Ce(III) were conducted with concentrations from 50 to 500 mg/L. The optimum pH for La(III) and Ce(III) adsorption was fixed at pH 5. To evaluate the control factors of the adsorption process, the Langmuir isotherm model, Freundlich isotherm model and Dubinin–Radusckevich (D-R) isotherm model were used.

2.5. Desorption and regeneration studies

In order to find the optimum condition using HCl solution as the desorbing reagent for La(III) or Ce(III)-loaded composite, different HCl concentration on the desorption efficiency was firstly studied. The HPC-g-PAA/APT (20/1) was immersed in La(III) or Ce(III) solution for reaching adsorption saturation. Then, the adsorbent loaded the rare earth was filtrate, put into 25 mL HCl solution with different concentrations from 0.10 to 2 mol/L and shaken for 2 h. After this adsorption–desorption process, the adsorption capacity and desorption capacity can be calculated, upon which the desorption ratios were expressed by the mass ratio of the eluted ions to the adsorbed ions.

The reusability of the adsorbent was evaluated according to the following procedures: 25 mg adsorbent was immersed in La(III) or Ce(III) solution until adsorption saturation was reached, then the adsorbent was separated from the solution by filtration and regenerated using 0.1 mol/L NaOH solution. The recovered adsorbent was washed with distilled water for three times and then used in next adsorption process. The consecutive adsorption–desorption process was performed for five times, and the adsorption capacity and recovery ratio of the adsorbent regenerated for different times were obtained. The solution concentration of adsorption process and desorption process were all determined and used to calculate the adsorption rate and the desorption rate.

3. Results and discussion

For the graft polymerization, persulfate, benzoyl peroxide and azoisobutyronitrile are the frequently used initiators. However, the grafting efficiency is low than redox initiator [30]. This is because that the hydroxyl radical (•OH) is the most reactive radical among the reactive oxygen species. In this study, the hydroxyl radical (•OH) was firstly produced through the decomposition of H_2O_2

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