

Available online at www.sciencedirect.com

ScienceDirect

www.elsevier.com/locate/jes

JES
 JOURNAL OF
 ENVIRONMENTAL
 SCIENCES
www.jesc.ac.cn

Q3 Phosphine production in anaerobic wastewater treatment 2 under tetracycline antibiotic pressure

Q5 Q4 Meiqing Lu^{1,2,3}, Xiaojun Niu^{1,2,4,5,6,*}, Weiyi Chen¹, Liang Zhu¹, Sheng You⁵, Xiaohong Gu²

4 1. School of Environment and Energy, South China University of Technology, Guangzhou 510006, China

5 2. State Key Laboratory of Lake Science and Environment, Nanjing 210008, China

6 3. School of Environmental Science and Engineering, Southern University of Science and Technology,

7 Shenzhen Key Laboratory of Soil and Groundwater Pollution Control, Shenzhen 518055, China

8 4. Guangdong Provincial Key Laboratory of Atmospheric Environment and Pollution Control, Guangzhou 510640, China

9 5. China Water Resources Pearl River Planning Surveying & Designing Co., Ltd., Guangzhou 510640, China

10 6. The Key Lab of Pollution Control and Ecosystem Restoration in Industry Clusters, Ministry of Education,

11 South China University of Technology, Guangzhou Higher Education Mega Centre, Guangzhou 510006, China

12

1 5 A R T I C L E I N F O

16 Article history:

17 Received 17 July 2017

18 Revised 27 October 2017

19 Accepted 30 October 2017

20 Available online xxxx

35 Keywords:

36 Tetracycline antibiotic

37 Anaerobic

38 Wastewater treatment

39 Phosphine production

40

A B S T R A C T

The influence of tetracycline (TC) antibiotics on phosphine (PH₃) production in the anaerobic 21 wastewater treatment was studied. A lab-scale anaerobic baffled reactor with three compart- 22 ments was employed to simulate this process. The reactor was operated in a TC-absence 23 wastewater and 250 µg/L TC-presence wastewater for three months after a start-up period, 24 respectively. The responses of pH, oxidation–reduction potential (ORP), chemical oxygen 25 demand (COD), total phosphorus (TP), enzymes activity (dehydrogenase and acid phosphatase), 26 and microbial community were investigated to reveal the effect of TC on PH₃ production. 27 Results suggested that the dehydrogenase (DH) activity, acid phosphatase (ACP) activity and 28 COD have positive relationship with PH₃ production, while pH, ORP level and the TP in liquid 29 phase have negative relationship with PH₃ production. With prolonged TC exposure, decrease 30 in pH and increase in DH activity are beneficial to PH₃ production, while decrease in COD and 31 ACP activity are not the limiting factors for PH₃ production. 32

© 2017 The Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences. 33

Published by Elsevier B.V. 34

46 Introduction

47 Phosphine (PH₃) is a kind of reactive and reduced phosphorus 48 compound which has been recognized as a gaseous carrier of 49 phosphorus in global biogeochemical cycles (Devai et al., 50 1999). It suggests a gaseous link to the phosphorus biogeo- 51 chemical cycle in the global environment, playing a favorable 52 part in alleviating the situation of phosphorus scarcity and 53 eutrophication (Devai et al., 1999). PH₃ exists in two different 54 forms: free gaseous phosphine (FGP) and matrix-bound phos- 55 phine (MBP) (Devai and Delaune, 1995), which have been found

in the environment especially in anaerobic areas, such as 56 sewage treatment plants, wetlands, lakes, offshore areas, 57 coasts, paddy fields, and even in the biosphere of Antarctic 58 (Devai and Delaune, 1995; J. Geng et al., 2005; J.J. Geng et al., 2005; Q6 59 Hong et al., 2010; Zhu et al., 2006, 2014; Niu et al., 2013). 60 Phosphorus in sewage was considered as a possible precursor of 61 PH₃ since it was reported that 30%–45% of the total phosphorus 62 changed into FGP and transported into the atmosphere during 63 wastewater treatment (Glindemann et al., 1996). 64

There are four main mechanisms of anaerobic biological 65 phosphine production: (1) thermodynamic process; (2) biological 66

* Corresponding author. E-mail: xjniu@scut.edu.cn (Xiaojun Niu).

67 reduction of phosphate; (3) metal corrosion and phosphate
68 reduction; (4) oxidation of reduced phosphorus compounds
69 (Roels and Verstraete, 2001). Though it is still unclear, the
70 “biological reduction of phosphate” and “metal corrosion and
71 phosphate reduction” mechanisms are more likely to be
72 acceptable. As for the mechanism of “biological reduction of
73 phosphate”, butyric acid bacteria was reported to be able to
74 generate gaseous phosphorus compounds as well as other eight
75 different bacterial strains: *Lactobacillus casei*, *Streptococcus lactis*,
76 a non-specified sulfate-reducing strain, *Aerobacter polymyxa*,
77 *Aerobacter macerans*, *Clostridium acetobutyricum*, *Clostridium*
78 *butyricum* and *Escherichia coli* (Tsubota, 1959). Other studies also
79 discovered that hydrogenase of *Desulfovibrio*, *E. coli*, *Salmonella*,
80 *Salmonella arizonae*, *Clostridium sporogenes*, *C. acetobutyricum*, and
81 *Clostridium cochliarium* had the ability to produce PH_3 (Iverson,
82 1968; Jenkins et al., 2000). In terms of “metal corrosion and
83 phosphate reduction” mechanism, Glindemann et al. (1998)
84 proved that bio-corrosion in iron was capable of forming PH_3 .

85 There are also studies suggested that some environmental
86 factors such as pH level ORP level, nutrient substances
87 (C-containing compounds and P-containing compounds),
88 enzymes (DH, ACP) activity have close relationship with PH_3
89 production. Glindemann et al. (1998) found that FGP produc-
90 tion could be promoted in a lower pH level in a corrosive
91 aquatic media affected by microbial metabolites. Ding et al.
92 (2005) found a lower level of pH generated more PH_3 in an
93 anaerobic process. Similar finding could be observed in the
94 study of lake sediments by J. Geng et al. (2005), J.J. Geng et al.
95 (2005), who reported that that MBP tended to be emitted from
96 soils with pH <6. Zhu et al. (2009) and Niu et al. (2013) also
97 discovered that slightly acid environment were more suitable
98 for PH_3 generation in Antarctic seabird guanos soils and paddy
99 fields. Oxidation reduction potential (ORP) is another factor
100 that affects the formation of PH_3 . H_2PO_4^- can be reduced to PH_3
101 by obtaining electrons from hydrogen and organic matter at
102 the ORP level lower than -0.32 V and -0.31 V (Roels and
103 Verstraete, 2001), and the studies by Li et al. (2010) and Feng
104 et al. (2008) indicated that a reducing environment with low
105 ORP is necessary for the production and preservation of MBP.
106 As for nutrient substances such as C-containing compounds
107 and P-containing compounds, it was reported that higher
108 organic carbon (OC) environment was more favorable to PH_3
109 generation in coastal sediments (Feng et al., 2008; Hou et al.,
110 2009), and both laboratory simulation and field observations
111 suggested that P-containing compounds had positive relation-
112 ship with PH_3 production as well (Devai et al., 1988; Han
113 et al., 2002, 2011; Song et al., 2011). Moreover, enzyme activity
114 also had intimate correlation with PH_3 production since DH
115 was a catalyst for formation of PH_3 (Niu et al., 2015), and ACP
116 was an indicator of microbes activities involving in the PH_3
117 production (Zhu et al., 2009, 2011).

118 However, an emerging contaminant, antibiotics have been
119 frequently detected in wastewater recently (Kümmerer, 2009).
120 The concentrations of antibiotics in wastewater treatment
121 plants range from ng/L to a few $\mu\text{g/L}$ (Kümmerer, 2009), a
122 higher level of mg/L can be seen in certain point sources such
123 as hospital or pharmaceutical industry effluents (Marathe
124 et al., 2013). The tetracycline (TC) group is one of the most
125 frequently detected antibiotics in wastewater (Watkinson
126 et al., 2007). It is a broad-spectrum active compound, and

127 has the ability to inhibit bacterial protein synthesis by binding
128 the 30S ribosomal subunit to prevent the association of the
129 aminoacyl-tRNA to the ribosomal acceptor-A site resulting in a
130 structural change of 16S rRNA (Cetecioglu et al., 2013). Another
131 significant feature of TC is that TC is highly absorbable to clay
132 materials, soil, sediments and its molecules existing in the
133 cation (TC^{+00}), zwitterion (TC^{+-0}), monoanion (TC^{--}) and
134 dianion (TC^{0--}) form in the corresponding condition accounted
135 for the pKa values was 3.3, 7.7, and 9.7, respectively (Figueroa
136 et al., 2004; Chen and Huang, 2009; Michael et al., 2013).

137 Since bacterial protein synthesis is vulnerable to TC, the
138 microbial activity and community composition are prone
139 to be affected (Maria et al., 2014; Zhang et al., 2013, 2016).
140 Matos et al. (2014) found that the fraction of *Flavobacterium*,
141 *Caulobacter* and *Zooglea*-like bacteria and of a member of the
142 *Sphingobacteriaceae* family decreased in the bacterial commu-
143 nity, while the fraction of *Sandarakinorhabdus*-like bacteria
144 increased under $50\ \mu\text{g/L}$ tetracycline exposure. Zhang et al.
145 (2013) discovered that trace TC could substantially change the
146 structure of the microbial community. The reason was that
147 TC-resistant species are capable of surviving and TC resis-
148 tance genes proliferated under TC exposure, while those do
149 not have the ability to resist TC tended to disappear. Zhang
150 et al. (2016) observed that the *Actinobacteria* percentages
151 and total antibiotic resistance genes abundances increased
152 along with TC addition, while *Proteobacteria* and *Bacteroidetes*
153 witnessed a decrease. Due to the change of microbes, the
154 degradation of nutrient substances and the activity of related
155 enzymes are easily affected. Loftin et al. (2005) discovered that
156 TC had a negative impact on COD removal in anaerobic batch
157 systems with a dose-response study of 1, 5, and 25 mg/L.
158 Cetecioglu et al. (2013) observed a noticeable effect of TC on
159 the overall COD removal at the dosing level 8.5 mg/L. As for
160 enzymes activity, DH is an intracellular enzyme, belonging to
161 the group of oxidoreductase that is able to oxidize a substrate
162 by a reduction reaction that transfers one or more hydrides
163 (H^-) to electron acceptors. It is an important catalyzer in
164 oxidation and reduction reactions. Previous study had sug-
165 gested that the inhibition of some competitive bacteria was
166 able to promote the population of DH-produced microorgan-
167 isms (Niu et al., 2015) ACP is an extracellular enzyme involved
168 in polyphosphate hydrolysis, which is sensitive to TC (Yang
169 et al., 2015). That indicated TC is easily to impact ACP activity,
170 and the study by Ghosh et al. (1999) revealed that the sludge-
171 absorbed cation (TC^{+00}) of TC were capable of inhibiting ACP
172 activity as well. Besides nutrient substances and enzymes
173 activity, other corresponding conditions such as pH, ORP in
174 the exposing system also tended to be affected by TC (Zhang
175 et al., 2013). It was reported that H^+ in the sludge was prone
176 to release (via exchange cation) to achieve sludge-solution
177 equilibrium upon TC^{+00} introduction (Figueroa et al., 2004),
178 which would lower the pH level in the system. In addition, the
179 transfer of H^+ would have a direct influence on ORP level
180 because H^+ was an electron carrier (Li et al., 2007). Overall,
181 microbial community and the corresponding conditions that
182 related to PH_3 production are vulnerable to the effects of TC.

183 However, to date, most studies on PH_3 are focusing on
184 the formation and distribution of PH_3 or the related environ-
185 mental factors on its emission, while studies of TC on PH_3
186 production have been fairly limited. In this study, we aim to

Download English Version:

<https://daneshyari.com/en/article/8865419>

Download Persian Version:

<https://daneshyari.com/article/8865419>

[Daneshyari.com](https://daneshyari.com)