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Evaluation of mercury biogeochemical cycling at the sediment-water interface in anthropogenically

4 modified lagoon environments

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ABSTRACT

The Marano and Grado Lagoon is well known for being contaminated by mercury (Hg) from 23 Q4 the Idrija mine (Slovenia) and the decommissioned chlor-alkali plant of Torviscosa (Italy). 24 Experimental activities were conducted in a local fish farm to understand Hg cycling at the 25 sediment-water interface. Both diffusive and benthic fluxes were estimated in terms of 26 chemical and physical features. Mercury concentration in sediments (up to $6.81 \mu g/g$) 27 showed a slight variability with depth, whereas the highest methylmercury (MeHg) values 28 (up to 10 ng/g) were detected in the first centimetres. MeHg seems to be produced and 29 stored in the 2–3 cm below the sediment–water interface, where sulphate reducing bacteria 30 activity occurs and hypoxic-anoxic conditions become persistent for days. DMeHg in 31 porewaters varied seasonally (from 0.1 and 17% of dissolved Hg (DHg)) with the highest 32 concentrations in summer. DHg diffusive effluxes higher (up to 444 $ng/(m^2 d)$ than those 33 reported in the open lagoon (~95 $ng/(m^2 d)$, whereas DMeHg showed influxes in the fish 34 farm (up to $-156 \text{ ng/(m}^2 \text{ d})$). The diurnal DHg and DMeHg benthic fluxes were found to be 35 higher than the highest summer values previously reported for the natural lagoon 36 environment. Bottom sediments, especially in anoxic conditions, seem to be a significant 37 source of MeHg in the water column where it eventually accumulates. However, net fluxes 38 considering the daily trend of DHg and DMeHg, indicated possible DMeHg degradation 39 processes. Enhancing water dynamics in the fish farm could mitigate environmental 40 conditions suitable for Hg methylation. 41

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55 Introduction

56 The Mediterranean coastal areas are essential ecological systems that constitute important resources of great recreational 57 58 and profitable value, in particular, fisheries and aquaculture play a crucial socio-economic role (Abdou et al., 2017). In the 59 60 Mediterranean, the aquaculture industry has grown rapidly 61 (FAO, 2016; Belias and Dassenakis, 2002), facilitated by the geography and chemical and physical conditions in the zone 62 63 (FAO, 2016). Over the last decades, aquaculture activities have 64 expanded rapidly throughout the world and the production 65 increased on an average rate of 23.5% from 2009 to 2014 (FAO, 2016). At present, approximately 44% of fish consumed are 66 farmed, and this percentage is predicted to reach 52% by 2025 67 68 (FAO, 2016).

69 One of the main environmental issues affecting the coastal 70 marine environments is the accumulation of contaminants in sediments and their possible mobility (Emili et al., 2016). The 71 sediment, in fact, is a sink for many chemical species 72 73 potentially toxic to the marine ecosystem, such as heavy metals (e.g., Chen et al., 2013; Kalantzi et al., 2016) or organic 74 compounds (e.g., Fent, 1996; Tas et al., 1996), but it is also the 75 76 site of early diagenesis biogeochemical processes that can mobilise contaminants, often transforming them into most 77 toxic forms which may subsequently accumulate through the 78 79 trophic chain.

80 In the coastal area of the Friuli Venezia Giulia region 81 (Northern Adriatic Sea, Italy), the Marano and Grado Lagoon is 82 well known for a high degree of anomaly in relation to 83 mercury concentration (Hg) in waters and especially sediments (Acquavita et al., 2012). This contamination is due to a 84 double source: the first is located within the Isonzo River 85 basin, the main river flowing into the Gulf of Trieste (Horvat 86 et al., 1999; Covelli et al., 2001, 2007), which is affected 87 88 by contamination from the Idrija mine (Western Slovenia) which was in operation for approximately 500 years. The 89 second source of contamination is located in the Aussa-Corno 90 River system, the result of chlor-alkali plant effluents 91 discharged into the Marano and Grado Lagoon (Covelli et al., 92 93 2009)

Mercury is considered a priority contaminant due to its 94 toxicity (WHO, 1990), mobility and bioaccumulation potential 95 and exhibits complex biogeochemical cycling (Mason et al., 96 97 1993; Ullrich et al., 2001; O'Driscoll et al., 2005). Anoxic 98 sediments can become a significant source of methylmercury 99 (CH₃Hg or MeHg⁺), which is a potent neurotoxin (e.g., Grandjean et al., 2010) and endocrine-disrupting chemical (Tan et al., 2009), 100 highly toxic to humans and other organisms (Crespo-López 101 et al., 2009). This is the most toxic organic form that can 102 be highly biomagnified in the trophic chain, most significantly 103 in edible fish. Other microorganisms in addition to SRB are 104 involved in Hg methylation, such as iron-reducing (Fleming 105 106 et al., 2006; Kerin et al., 2006) and methanogenic (Gilmour et al., 107 2013; Hamelin et al., 2011; Podar et al., 2015) bacteria that are 108 capable of Hg methylation. It has been shown that sulphur cycling (Gagnon et al., 1996; Lambertsson and Nilssons, 2006) is 109 important in controlling the transformation of Hg species 110 (e.g., Han et al., 2007; Hollweg et al., 2009) such as the amount 111 and characteristics of dissolved organic matter (Ravichandran, 112

2004) and sorption/dissolution processes involving Fe/Mn 113 oxy-hydroxides (e.g., Gagnon et al., 1997; Schäfer et al., 2010). 114

Various research activities were carried out in order to more 115 fully understand the Hg biogeochemical cycle in the northern 116 Adriatic Sea in the last 20 years, from the more extended area of 117 the Gulf of Trieste (Covelli et al., 1999; Horvat et al., 1999) to the 118 Marano and Grado Lagoon (Covelli et al., 2008, 2011; Emili et al., 119 2012). Despite Hg issue, fishing, along with clam and mussel 120 collection and aquaculture practice are very widespread. All 121 these activities increase the chances of Hg uptake by living 122 aquatic organisms in the nearby area (Covelli et al., 2008). Few 123 studies have been performed in this area regarding Hg impact on 124 the trophic chain. Brambati (2001) considered Hg bioaccumula- 125 tion in seaweed, edible fish and birds. High Hg concentrations, 126 often one order of magnitude higher than the 0.5 mg/kg 127 European Community limit, were found along trophic chain 128 (especially in seabream and seabass). More recently, a multidis- 129 ciplinary research project called "MIRACLE" (Mercury Interdisci- 130 plinary Research for Appropriate Clam farming in a Lagoon 131 Environment) was conducted to identify which areas pose the 132 least risk of Hg bioaccumulation for commercial Manila clams 133 (Covelli, 2012). To this purpose, Hg and MeHg content in Tapes 134 philippinarum was monitored in natural and seeded populations 135 (Giani et al., 2012). Increased bioaccumulation of Hg but not 136 of MeHg with increasing size of wild clam populations was 137 observed at most sites depending on Hg concentrations in 138 bottom sediments but also on biogeochemical processes at the 139 sediment-water interface (SWI). 140

Fish farming is a historical activity for the local population, 141 covering 14% of the total lagoon area (Acquavita et al., 2015), 142 where mostly sea bass and sea bream are farmed. Fish farms 143 were built by isolating an area of the lagoon with an embank- 144 ment. They communicate with Lagoon waters through sluice 145 gates, which allow the water exchange between the external 146 lagoon environment and the fish farms. They can be self- 147 contaminated mainly by organic matter and heavy metals due 148 to inappropriate operational practices such as the cultivation of 149 high fish populations, excess of food given and use of antifouling 150 chemical substances (Belias and Dassenakis, 2002).

Since the bottom sediments in the fish farms are highly 152 contaminated by mercury, this study is a first step toward 153 understanding the role of the SWI in recycling mercury in this 154 anthropogenically modified lagoon environment. To this end, an 155 estimate of Hg and MeHg mobility and exchanges between 156 sediment and water column was performed at two different sites 157 located in this environment, in order to 1) highlight differences 158 with the open lagoon system and 2) to provide new insight for 159 future studies on Hg transformations (*e.g.*, methylation/demeth-160 ylation rates) and bioaccumulation in edible fish. The final aim 161 was to suggest good practices for improving the environmental 162 conditions of the fish farm in order to mitigate the impact of Hg 163 on this complex and highly anthropogenically modified lagoon 164 ecosystem.

1. Sampling strategy

1.1. Study area

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The wetland system of the Marano and Grado Lagoon 169 (Northern Adriatic Sea, Italy) is one of the best conserved 170

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