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Alkynyl carbon materials as novel and efficient sorbents for the adsorption of mercury(II) from wastewater

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ABSTRACT

For the first time, a series of alkynyl carbon materials (ACMs) were prepared via the 16 mechanochemical reaction of CaC₂ with six polyhalogenated precursors, namely CCl₄, 17 C₂Cl₆, C₂Cl₄, C₆Cl₆, C₆Br₆, and C₁₄H₄Br₁₀ (ACM-1, ACM-2, ACM-3, ACM-4, ACM-5, and ACM-6, 18 respectively) and used for the adsorptive removal of mercury from aqueous solutions. 19 Based on preliminary investigations, the adsorption of mercury on ACM-5 was studied in 20 depth. Specifically, the effect of pH on mercury adsorptivity, adsorption kinetics, 21 thermodynamics, isotherms, and recyclability was studied. The adsorptivity of mercury 22 on ACMs was found to be closely related to the hydrocarbon precursor, specific surface area 23 of sorbent, and the alkynyl content. ACM-5 showed the best performance and is among the 24 best raw carbonaceous sorbents reported so far, with a Langmuir saturated adsorption 25 capacity of 191.9 mg g⁻¹. The promising mercury adsorption performance mainly arises 26 from the strong Lewis soft acid-soft base interactions between the alkynyl groups and 27 mercury ions. The adsorption isotherms could be satisfactorily correlated with the 28 Langmuir equation. The results show that the ACMs can be used as efficient sorbents for 29 the removal of mercury and may also be useful for the adsorption of other heavy metals. 30 © 2017 The Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences. 31

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44 Introduction

46 Mercury ion (Hg(II)) pollution in the aquatic ecosystem has 47 caused widespread concern owing to the strong toxicity and 48 bioaccumulation of mercury (Boening, 2000; Fitzgerald and 49 Clarkson, 1991). Considering the serious hazards of mercury to the central nervous system, lungs, kidneys, and chromosomes 50(Clarkson, 1993; Renzoni et al., 1998), the World Health 51Organization (WHO) has specified the maximum allowable 52mercury ion concentration in drinking water as 1 µg/L. How-53 ever, enormous amounts of mercury are still being emitted into 54the environment through atmospheric deposition as well as 55

urban and industrial discharges (He et al., 2013; Li et al., 2009; 56 Zhang and Wong, 2007). Therefore, its removal from aquatic 57 systems has become an urgent problem. 58

For the complete removal of Hg(II) from aquatic ecosys- 59 tems, numerous techniques such as chemical precipitation, 60 ion exchange, coagulation, reduction, membrane filtration, 61 and adsorption have been explored (Hashim et al., 2011; Plaza 62 et al., 2011; Yu et al., 2008). Among these techniques, 63 adsorption is deemed to be the most practical and economical 64 approach (Fu and Wang, 2011). The identification of 65 efficient sorbents is the key to designing a good adsorption 66 process. In this context, carbon materials (CMs) have received 67

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considerable attention as sorbents owing to their widespread 68 availability, exceptional water filtration capabilities, and high 69 heavy metal adsorption capabilities (Mohmood et al., 2013; 70 Nasser et al., 2016). In particular, activated carbons (ACs), which 71 are cheap and environmentally benign sorbents, have been 72 extensively used for Hg(II) adsorption (Anirudhan and 73 Sreekumari, 2011; Bailey et al., 1999; Di Natale et al., 2006; Hadi 74 75et al., 2015; Kumar, 2006), although their adsorption capabilities 76 are often poorer than task-specific sorbents. Further, the Hg(II) 77 adsorptivity of ACs has been improved by modifying them with Lewis base atoms such as sulfur, oxygen, nitrogen, and 78 halogens (Anbia and Dehghan, 2014; Bhatnagar et al., 2013; Li 79 et al., 2005; Rivera-Utrilla et al., 2011; Yin et al., 2007). Among the 80 modified AC materials, S-modified ACs exhibit superior Hg(II) 81 adsorptivity, owing to the strong affinity between S and Hg(II) as 82 a result of soft base-soft acid interactions (Anbia and Dehghan, 83 2014; Beck et al., 2000; Bhatnagar et al., 2013; De Canck et al., 84 2010; Gomez-Serrano et al., 1998). However, the sulfur grafting 85 of ACs has low efficiency, which results in high wastage of 86 energy and resources during the modification process, thereby 87 limiting its commercial use (Bhatnagar et al., 2013; Hsi et al., 88 2001; Rivera-Utrilla et al., 2011). 89

Considering the difficulties in grafting soft bases onto CMs, 90 91 we have developed a facile approach for introducing alkynyl 92groups onto CMs via a one-step carbon synthesis process, 93 forming a new type of alkynyl carbon material (ACM). The ACMs 94 are expected to exhibit strong Hg(II) adsorption arising from the 95 strong interactions between the alkynyl group and Hg(II) (soft base-soft acid complexation), as manifested by the excellent 96 catalysis of HgCl₂ in the hydrochlorination of acetylene 97 (Hutchings, 1985). With this in mind, six ACMs were synthe-98 sized by reacting CaC₂ with a specific polyhalogenated hydro-99 carbon (PHHC) in a planetary ball mill at ambient temperature. 100 The Hg(II) adsorptivity of different ACMs was compared, and 101 the effect of pH, adsorption kinetics, adsorption isotherms, and 102recyclability were studied. 103

104 1. Experimental

106 **1.1. Materials and reagents**

107 Calcium carbide (75%-80%) was purchased from Alfa Aesar Chemical Company (China) and ground into 100 mesh parti-108 cles before use. Elemental mercury (99.9999%) was kindly 109 supplied by the Ministry of Environment Protection of China. 110 Carbon tetrachloride (AR), hexachloroethane (AR), tetrachlo-111 roethylene (AR), nitric acid (65%-68%), and hydrochloric acid 112(36%-38%) were obtained from Beijing Chemical Works 113 (China). Hexachlorobenzene (AR), hexabromobenzene (AR), 114 115decabromodiphenylethane (AR), and potassium hydroxide 116 (GR) were procured from Aladdin Industrial Corporation (China). Ultrapure water produced with a CP-500S water 117 purification system (CANPURE, China) was used for all the 118 experiments, unless specifically mentioned otherwise. 119

A stock solution of Hg(II) (100 mg/L) was prepared by the nitrolysis of specific amounts of elemental mercury, followed by dilution with water. It was stored in darkness, and the working solutions were prepared immediately before use by diluting the stock solution.

1.2. Synthesis and characterization of ACMs

A planetary ball mill (QXQM-2, Tianchuang, China) was used 126 for the mechanochemical synthesis of the ACMs. To a 127 stainless-steel pot (250 mL) containing 250 g of stainless- 128 steel balls of four different diameters (15, 12, 10, and 8 mm), 129 CaC_2 and PHHC were added in a specific mass ratio and the 130 pot was sealed air-tight. Ball milling was conducted at room 131 temperature and at a specific rotation speed under vacuum. 132 The resulting solid mixture was treated with dilute nitric acid 133 and washed three times with pure water. The ACMs were then 134 obtained by vacuum drying at 100°C for 5 hr. The yield of the 135 ACMs and the carbon content were determined by elemental 136 analysis (VarioELcube, Germany). The experimental condi-137 tions used, the ACM yields, and the results of elemental 138 analysis are shown in Table 1.

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1.3. Hg(II) adsorption by ACMs 140

All the adsorption experiments were conducted in 100 mL 141 conical flasks at specific temperatures. To the conical flask, 142 25 mg of sorbent and 50 g of 50 mg/L Hg(II) solution were 143 added and stirred magnetically at 200 r/min for 3 hr. The 144 liquid samples were withdrawn using a syringe with a 145 0.22 μ m membrane filtration head, and the Hg(II) concentra- 146 tions were measured by a spectrophotometric method using 147 rhodamine-6G in a LabTech UV–vis spectrophotometer at 148 575 nm (Ramakrishna et al., 1976). The Hg(II) standard curve is 149 shown in Appendix A Fig. S1. The equilibrium adsorbed 150 capacity of the sorbents was calculated using the following 151 equation:

$$q_{\rm e} = \frac{(C_0 - C_{\rm e})V}{W} \tag{1}$$

where, q_e (mg/g) is the equilibrium adsorption capacity of the **153** sorbent; C_0 (mg/L) and C_e (mg/L) are the Hg(II) concentrations 155 at initial and equilibrium states, respectively; V (L) is the 156 solution volume, and W (g) is the mass of sorbent. All the 157 samples were analyzed thrice and the average values are 158 reported. As shown in Appendix A Tables S2–S6, the average 159 absolute relative deviation (AARD) and the root mean square 160 deviation (RMSD) values of the experimental data are within 161 3.3% and 3.5%, respectively. Based on the initial experimental 162 results, ACM-5 was chosen as the sorbent for the subsequent 163 experiments.

1.4. Effect of solution pH 165

The effect of solution pH on Hg(II) adsorption was studied for 166 the ACM-5 sorbent in the pH range of 1.0–8.0. The pH values 167 were adjusted using 0.1 mol/L HNO₃ and 0.1 mol/L NaOH 168 solutions. A pH meter from Rex Electric Chemical (ID-2) was 169 used for the pH measurements. The pH meter was calibrated 170 using standard buffer solutions with pH values of 4.00, 6.86 171 and 9.18. 172

1.5. Adsorption kinetics and activation energy 173

The kinetics of Hg(II) adsorption on ACM-5 was studied at 20, $_{\rm 174}$ 40, and 60°C, at a pH of 5.8. Pseudo-first and second-order $_{\rm 175}$

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