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A comparison study of the start-up of a MnO_x filter for catalytic oxidative removal of ammonium from groundwater and surface water

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ABSTRACT

As an efficient method for ammonium (NH⁴) removal, contact catalytic oxidation technology 16 has drawn much attention recently, due to its good low temperature resistance and short 17 start-up period. Two identical filters were employed to compare the process for ammonium 18 removal during the start-up period for ammonium removal in groundwater (Filter-N) and 19 surface water (Filter-S) treatment. Two types of source water (groundwater and surface 20 water) were used as the feed waters for the filtration trials. Although the same initiating 21 method was used, Filter-N exhibited much better ammonium removal performance than 22 Filter-S. The differences in catalytic activity among these two filters were probed using X-ray 23 diffraction (XRD), scanning electron microscopy (SEM), X-ray photoelectron spectroscopy 24 (XPS), and compositional analysis. XRD results indicated that different manganese oxide 25 species were formed in Filter-N and Filter-S. Furthermore, the Mn3p XPS spectra taken on 26 the surface of the filter films revealed that the average manganese valence of the inactive 27 manganese oxide film collected from Filter-S (FS-MnOx) was higher than in the film collected 28 from Filter-N (FN-MnOx). Mn(IV) was identified as the predominant oxidation state in 29 FS-MnO_x and Mn(III) was identified as the predominant oxidation state in FN-MnO_x. The 30 results of compositional analyses suggested that polyaluminum ferric chloride (PAFC) used 31 during the surface water treatment was an important factor in the mineralogy and reactivity 32 of MnO_x. This study provides the theoretical basis for promoting the wide application of the 33 technology and has great practical significance. 34

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49 Introduction

Human activity has drastically impacted the global nitrogen cycle, leading to excessive amounts of ammonium (NH⁴₄) in groundwater and surface water (Qin et al., 2016). According to monitoring data from recent years, more than 40% of the surface water in China cannot meet the national standard for drinking water sources (<1.0 mg/L NH⁴₄-N, Water Quality Standard for Drinking Water Sources (CJ3020-93)), being mainly 57 contaminated by organic matter and ammonium (Feng et al., 58 2012). The presence of ammonium in water systems leads to 59 oxygen depletion, eutrophication of surface water and toxicity 60 for fish (Hasan et al., 2011; Li et al., 2011). Additionally, during 61 drinking water treatment, ammonium must be removed before 62 disinfection with chlorine because it consumes chlorine and 63 produces chloramines (Cai et al., 2014). The above-mentioned 64

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situation has threatened the availability of safe drinking waterand thus, human health.

To remove ammonium effectively and economically, a 67 great deal of work has been done on ammonium treatment 68 processes and removal pathways (Cheng et al., 2016; Du 69 et al., 2016; Huang et al., 2015). Generally, physicochemical 70 (chlorination, ion exchange, membrane filtration, etc.) (Feng 71 and Sun, 2015; Kurama et al., 2002; Leaković et al., 2000; 72 73 Sprynskyy et al., 2005) and biological methods (biological 74 filter, biological contact oxidation, alga, photosynthetic bacteria, etc.) (Cai et al., 2015; Han et al., 2013; Luo et al., 2016; Wen 75et al., 2016) are the most effective methods for ammonium 76 removal. However, there are both advantages and disadvan-77 tages for these methods. Excessive amounts of chlorine may 78 stimulate the formation of undesirable chlorinated by-products 79 in the drinking water (Choi and Valentine, 2002; Yu et al., 2007). 80 The application of ion exchange and membrane filtration 81 are limited due to their high capital and operational costs 82 (Bódalo et al., 2005; Cai et al., 2015). Low temperature is another 83 limiting factor in ammonium removal when using con-84 ventional biological approaches, which generally drastically 85 affects the processing rate of nitrifying bacteria (Huang 86 et al., 2013; Zhang et al., 2013). According to the literature 87 88 (Zhang et al., 2013), the optimal temperature for nitrification 89 in pure culture ranges from 25 to 35 °C. Below 15 °C, the 90 nitrification rate drops sharply, which could result in nitrite 91 accumulation.

92More recently, a novel technology was proposed to remove ammonium from groundwater, which was called "contact 93 catalytic oxidation technology". Specifically, a manganese 94oxide filter was built to remove ammonium from groundwater 95 based on the abiotic transformation of ammonium by manga-96 nese oxide catalytic oxidation (Cheng et al., 2017). Enormous 97 98 effort has been expended on this research (Cheng et al., 2017; Guo et al., 2015, 2017). The technology was able to achieve 99 efficient ammonium removal from groundwater at low tem-100 perature (Guo et al., 2017). Additionally, the behavior of 101 ammonium catalytic oxidation by the manganese oxide 102 filter was investigated systematically, and the formation and 103 evolution of the manganese oxide films were characterized 104 extensively during the groundwater treatment (Cheng et al., 105106 2017). However, no studies have been done to explore the feasibility of applying this "contact catalytic oxidation 107 technology" to remove ammonium from surface water. This 108 definitely could have a significant impact in the field of 109 surface water purification, since a large number of water 110 treatment plants use surface water rather than groundwater 111 as a water source. Moreover, traditional drinking water 112treatment processes are usually ineffective in ammonium 113 removal, especially during the winter when the average 114 115temperature is always below 10 °C (Qin et al., 2016; Huang et al., 2013; Zhang et al., 2013). Accordingly, in this study 116 we attempted to apply this manganese oxide filter catalytic 117 oxidation technology to surface water treatment. Unlike 118 119 groundwater, surface water has some unique properties such as high turbidity, variable temperature and low hard-120 121 ness. All of these factors may affect the formation and/or the performance of manganese oxide films toward ammonium. 122 On the other hand, considering that coagulation is always a 123 124necessary step in surface water treatment plants, and some

coagulants (e.g., polyaluminum ferric chloride (PAFC)) could 125 remain in water after the coagulation/precipitation process, 126 these residual coagulants may also influence the removal of 127 ammonium by the manganese oxide filter. 128

In sum, this paper will compare the processes for ammo- 129 nium removal during the start-up period for ammonium 130 removal in groundwater and surface water treatment, and 131 carries out a preliminary study of the limiting factors for the 132 application of the technology in surface water treatment. This 133 study aims to provide the theoretical basis for promoting the 134 wide application of the technology. 135

Therefore, in this study, two types of source water 136 (groundwater and surface water) were used as the feed water 137 for the filtration trials. The ammonium removal performance 138 of these two filters was compared during the start-up period. 139 Meanwhile, the structure, composition and morphology of the 140 manganese oxides were also characterized. 141

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1. Materials and methods

1.1. The start-up and operation of filters

Two identical lab-scale gravity filters made of plexiglass were 145 adopted for the start-up phase of ammonium removal. 146 Schematic diagrams of the filter systems are shown in 147 Appendix A Fig. S1. The filter column used in this study had 148 an inner diameter of 100 mm and was filled with virgin quartz 149 sand of size fraction 0.75–1.20 mm, with a height of 1.0 m. The 150 detailed initiating processes of the filters are described in 151 Appendix A. Two types of source water were used as the feed 152 water for the filtration trials. Local raw surface water was used 153 as the influent of the filter (Filter-S) in our lab at the south 154 suburb of Xi'an, and local raw underground water as the 155 influent of the filter (Filter-N) in a lab at the north suburb 156 (Appendix A Fig. S2). The water quality of the raw ground- 157 water and surface water are shown in Table 1. During the 158 start-up period, potassium permanganate (MnO_{4}^{-}) and man- 159 ganese chloride (Mn(II)) with a mole ratio of 0.6 were added 160 to the feed water to form MnO_x, and then the suspension 161 was pumped into the filter. When the influent ammonium 162 could be removed almost completely, a filter was regarded 163 as being successfully started and the addition of potassium 164 permanganate and MnCl₂ solution was stopped. Backwashing 165 was performed when filters became clogged. Ammonium 166

Table 1 – Quality of the raw groundwater and surface water.			t1.1
Parameters	Groundwater	Surface water	$t_{1.4}^{1.3}$
рН	7–8	7–8	t1.5
Temperature (°C)	15-22	14–20	t1.6
DO (mg/L)	4–6	4–6	t1.7
TOC (mg/L)	1–3	1–5	t1.8
Total hardness (CaCO ₃)	70–80	190-200	t1.9
TDS (mg/L)	110-120	460-470	t1.10
Ca ²⁺ (mg/L)	35–40	25–30	t1.11
Mg ²⁺ (mg/L)	23–25	3–5	t1.12
HCO₃ (mmol/L)	6	1.35	t1.13

was continually added in order to maintain an ammonium 167

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