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# Q1 Phytotoxicity and groundwater impacts of leaching from 2 thermal treatment residues in roadways

Q2 *Khamphe Phoungthong*<sup>1,2</sup>, *Li-Ming Shao*<sup>2,3</sup>, *Pin-Jing He*<sup>2,3</sup>, *Hua Zhang*<sup>1,2,\*</sup>

4 1. State Key Laboratory of Pollution Control & Resource Reuse, Tongji University, Shanghai 200092, China. E-mail: [khamphe@hotmail.com](mailto:khamphe@hotmail.com)

5 2. Institute of Waste Treatment and Reclamation, Tongji University, Shanghai 200092, China

6 3. Centre for the Technology Research and Training on Household Waste in Small Towns & Rural Area, Ministry of Housing and  
 7 Urban–Rural Development of PR China (MOHURD), Shanghai 200092, China

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## A B S T R A C T

The use of coal fly ash (CFA), municipal solid waste incinerator bottom ash (MSWIBA) and flue 17  
 gas desulfurization residue (FGDR) in road construction has become very common owing to its 18  
 economical advantages. However, these residues may contain toxic constituents that pose an 19  
 environmental risk if they leach out and flow through the soil, surface water and groundwater. 20  
 Therefore, it is necessary to assess the ecotoxicity and groundwater impact of these residues 21  
 before decisions can be made regarding their utilization for road construction. In this study, 22  
 the physico-chemical characteristics, leaching and phytotoxicity of these residues were 23  
 investigated. Specifically, multivariate analyses were used to evaluate the contributions 24  
 of the leaching constituents of the CFA, MSWIBA and FGDR leachates to the germination 25  
 index of wheat seeds. B, Ba, Cr, Cu, Fe and Pb were found to be more toxic to the wheat 26  
 seeds than the other heavy metals. Furthermore, the leached concentrations of the 27  
 constituents from the CFA, MSWIBA and FGDR were below the regulatory threshold limits of 28  
 the Chinese identification standard for hazardous wastes. Analyses conducted using a 29  
 numerical groundwater model (WiscLEACH) indicated that the predicted field concentrations 30  
 of metals from the CFA, MSWIBA and FGDR increased with time up to about 30 years at the 31  
 point of compliance, then decreased with time and distance. Overall, this study demonstrated 32  
 that the risks resulting from MSWIBA, CFA and FGDR leaching could be assessed before 33  
 its utilization for road construction, providing crucial information for the adoption of these 34  
 alternative materials. 35

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## 51 Introduction

52 Rapid urbanization has led to demands for construction  
 53 materials that are exceeding the availability of natural  
 54 materials. At present, several countries are attempting  
 55 to replace natural materials with residues produced by  
 56 thermal treatments (Ore et al., 2007) such as municipal  
 57 solid waste incinerator bottom ash (MSWIBA), coal fly ash

(CFA) and flue gas desulfurization residue (FGDR). More than 58  
 10 million tonnes of MSWIBA (National Bureau of Statistics of 59  
 China, 2015), 550 million tonnes of CFA (Yu et al., 2015), and 60  
 7.1 million tonnes of FGDR (Wang et al., 2013) are produced 61  
 annually in China, showing great potential for resource 62  
 reuse (Cheng et al., 2007; Córdoba, 2015). However, the types 63  
 and quantities of solid residues vary according to the waste 64  
 composition, thermal treatment process, flue gas cleaning 65

\* Corresponding author. E-mail: [zhanghua\\_tj@tongji.edu.cn](mailto:zhanghua_tj@tongji.edu.cn) (Hua Zhang).

technology and residue handling at various facilities (Seshadri et al., 2010).

CFA, MSWIBA and FGDR are usually considered non-hazardous wastes that can be utilized as alternative construction materials, especially in roadways. The mineral compositions and engineering characteristics of these residues are similar to those of natural sand or aggregates. In China, sustainable waste management policies are encouraging reutilization of these residues (Geng et al., 2009; Wang et al., 2010). The environmental benefits include lowering the amounts of waste being sent to landfills and replacing natural materials. Thus, reuse of the residues could save landfill capacity and reduce the environmental impacts arising from exploitation of natural materials as well as transport of these residues to landfill sites and of virgin aggregates to points of use.

However, when CFA, MSWIBA and FGDR are used in construction, hazardous constituents contained in the residues may be leached out by runoff, surface water or groundwater that may come in contact with the materials. Such leaching represents a potential threat to the environment (Cheng et al., 2008; Liyanage et al., 2013; Yu et al., 2013). The compositions of CFA, MSWIBA and FGDR have been described for many different incineration plants and countries (Shim et al., 2005; Rendek et al., 2007; Hua et al., 2010; Rocca et al., 2012; Chen et al., 2014; Córdoba, 2015; Phoungthong et al., 2016a). To assess the environmental impacts arising from use of these materials, information is needed not only on the total contents of constituents in the residue materials, but also on the amounts of these constituents that might reach the surrounding environment. Therefore, the leaching behavior of alternative materials under field site conditions and during standardized laboratory tests has been discussed by several researchers. It was reported that the impacts of leaching from MSWIBA and CFA used in road construction on the soil and groundwater were low (Schreurs et al., 2000; Bruder-Hubscher

et al., 2001). Badreddine and François (2009) assessed the fate of PCDD/Fs from municipal solid waste incineration residues that were used in four road construction sites (>10 years) as surface base, and sub-base courses, and found that the materials posed little harm to the quality of road soils. In another study (Lidelöw and Lagerkvist, 2007), however, higher concentrations of Cr and Cu were observed in MSWIBA than in crushed rock used in a construction site.

Most studies conducted to date have focused on the physico-chemical properties and constituents of residue leachates. Chemical analysis is used to quantify pollutant concentrations and cannot account for the interaction among the pollutants in complex mixtures (Fan et al., 2006), thus providing insufficient comprehensive risk information. On the contrary, ecotoxicity is the result of the combination of several factors, such as heavy metals, ammonia, salts and volatile fatty acids. Therefore, to reveal the feasibility of utilizing MSWIBA, CFA or FGDR as a construction material, an ecotoxicity risk assessment of leachates derived from these residues needs to be conducted in addition to traditional chemical analyses. A few studies have included ecotoxicological analysis of MSWIBA and CFA leachate using methods such as the Microtox® toxicity test, *Daphnia magna* immobility test, *Ceriodaphnia dubia* death test, worm mortality (Quilici et al., 2004; Tsiridis et al., 2006; Ribé et al., 2014; Phoungthong et al., 2016a), and plant assays (Radetski et al., 2004; Phoungthong et al., 2016b). Different levels of toxicity have been recorded from the various tests, and the toxicity was greatly influenced by the pH status of the solid samples, the types of the leachants, as well as the concentrations of heavy metals and carboxylic acids in the residues' leachates. Therefore, the residues need to be formally tested with ecotoxic and genotoxic sensitive tests before recycling. Little is currently known about the phytotoxicity of CFA and FGDR.

The ecotoxicity of the residues is highly dependent on the leachability of the constituents and the leaching environment

**Table 1 – Physico-chemical characteristics of the coal fly ash (CFA), municipal solid waste incinerator bottom ash (MSWIBA) and flue gas desulfurization residue (FGDR).**

| Properties                   | CFA1  | MSWIBA1 | MSWIBA2 | FGDR1 | FGDR2 | FGDR3 | FGDR4 | FGDR5 |
|------------------------------|-------|---------|---------|-------|-------|-------|-------|-------|
| LOI (%wt, at 600°C)          | 3.13  | 3.31    | 1.14    | 2.25  | 9.56  | 3.64  | 6.73  | 6.06  |
| Density (g/cm <sup>3</sup> ) | 1.66  | 2.12    | 1.83    | 1.82  | 1.66  | 1.54  | 1.53  | 1.33  |
| d <sub>50</sub>              | 30 µm | 3.10 mm | 3.80 mm | –     | –     | –     | –     | –     |
| Elemental content (mg/kg)    |       |         |         |       |       |       |       |       |
| As                           | 44.2  | 57.1    | 85.4    | 34.7  | 31.5  | 41.3  | 63.0  | 51.0  |
| B                            | ND    | ND      | ND      | 1407  | 574   | 1086  | 3090  | 774   |
| Ba                           | 359   | 1240    | 2090    | 2183  | 2799  | 3073  | 3013  | 3030  |
| Be                           | 6.42  | 1.78    | 2.68    | ND    | 0.13  | 0.17  | ND    | 0.75  |
| Bi                           | ND    | ND      | ND      | 52.6  | 48.9  | 51.3  | 52.3  | 55.0  |
| Cd                           | 1.01  | 4.91    | 7.52    | ND    | ND    | ND    | ND    | ND    |
| Co                           | ND    | ND      | 12.4    | ND    | ND    | ND    | ND    | ND    |
| Cr                           | 65.8  | 330     | 676     | ND    | 4.73  | ND    | ND    | 6.71  |
| Cu                           | 56.0  | 1670    | 1710    | ND    | 0.29  | 1.38  | 0.76  | 18.1  |
| Mn                           | 213   | 705     | 1080    | 8.01  | 342   | 370   | 136   | 429   |
| Ni                           | 26.8  | 131     | 2107    | 2.58  | 22.9  | 22.9  | 9.93  | 23.9  |
| Pb                           | 86.9  | 482     | 609     | 1.49  | 7.29  | 3.13  | 1.29  | 14.1  |
| Se                           | 16.8  | 1.60    | 3.13    | ND    | ND    | ND    | ND    | ND    |
| Sr                           | 544   | 219     | 366     | 256   | 1501  | 1706  | 886   | 1168  |
| V                            | 140   | 44.7    | 54.0    | 0.47  | 27.4  | 28.7  | 19.1  | 58.2  |
| Zn                           | 81.9  | 2100    | 2370    | 354   | 335   | 390   | 458   | 468   |

ND: not detectable.

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