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Effects of biochars on the bioaccessibility of phenanthrene/pyrene/zinc/lead and microbial community structure in a soil under aerobic and anaerobic conditions

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Introduction

51 Co-contamination of organic and inorganic pollutants has 52 become a serious issue in agricultural soils located near 53 manufacturing regions because of industrial waste pro-54 duction, wastewater irrigation, and pesticide and chemical 55 leakage during agricultural production (Usman et al., 2016). The available fraction of these pollutants in soil can be taken 56 up by crops and transferred through the food chain, eventu-57 ally posing a threat to human health (Usman et al., 2016). 58 Therefore, immobilization of the co-contaminants in soil, 59 which seeks to reduce their transportation and bioavailability, 60 is an effective soil remediation strategy to reduce these envifor nonmental risks (Ahmad et al., 2014; Mohan et al., 2014). 62

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ABSTRACT

The immobilization of co-contaminants of organic and inorganic pollutants by biochar 19 is an efficient remediation strategy. However, the effect of biochar amendments on 20 the bioaccessibility of the co-contaminants in dry versus flooded soils has rarely been 21 compared. In batch experiments, bamboo-derived biochar (BB) had a higher sorption 22 capacity for phenanthrene (Phe)/pyrene (Pyr)/zinc (Zn) than corn straw-derived biochar 23 (CB), while CB had a higher sorption capacity for lead (Pb) than BB. After 150 days of 24 incubation, the amendments of 2% CB, 0.5% BB and 2% BB effectively suppressed the 25 dissipation and reduced the bioaccessibility of Phe/Pyr by 15.65%/18.02%, 17.07%/18.31% 26 and 25.43%/27.11%, respectively, in the aerobic soils. This effectiveness was more 27 significant than that in the anaerobic soils. The accessible Zn/Pb concentrations were also 28 significantly lower in the aerobic soils than in the anaerobic soils, regardless of treatments. 29 The Gram-negative bacterial biomass and the Shannon-Weaver index in the aerobic soil 30 amended with 2% CB were the highest. The soil microbial community structure was jointly 31 affected by changes in the bioaccessibility of the co-contaminants and the soil physiochemical 32 properties caused by biochar amendments under the two conditions. Therefore, dry land 33 farming may be more reliable than paddy soil cultivation at reducing the bioaccessibility of 34 Phe/Pyr/Zn/Pb and enhancing the soil microbial diversity in the short term. 35

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The use of biochar in environmental remediation has 63 been extensively investigated (Meyer et al., 2011). Biochars 64 can reduce the mobility and bioavailability of the organic 65 pollutants and heavy metals in soils through their large 66 surface area and rich functional groups via physical adsorp-67 tion, cation exchange, hydroxide formation and carbonate 68 69 precipitation (Sohi et al., 2010; Beesley et al., 2010; Park et al., 2011; Cao et al., 2011). Soil-use (e.g., dry land or paddy soil) 7071 can affect the environmental behavior of contaminants in soil and biochar-amended soil (Shankar, 2017). Studies 07 have reported that biochar can reduce the availability of the 73 co-contaminants of organic pollutants and heavy metals in 74 soils under aerobic conditions (Beesley et al., 2010; Cao et al., 75 2011; Wu et al., 2016). For example, Beesley et al. (2010) 76 reported that amendment with hardwood-derived biochar 77 decreased the concentration of accessible polycyclic aromatic 78 hydrocarbons (PAHs) by more than 50% and induced a 10-fold 79reduction in the available Cadmium (Cd) in the soil pore 80 water of aerobic soils. The extractability of CaCl₂ after 1 hr of 81 incubation was reduced by 25% for zinc (Zn) and 52% for lead 82 (Pb) in the presence of 5% Miscanthus straw-derived biochar 83 under aerobic conditions (Houben et al., 2013). Compared with 84 the studies on the immobilizing potential of biochar in aerobic 85 86 soils, those on the effect of biochar concerning the bioavailability of contaminants in anaerobic soils (e.g., paddy soil) 87 88 have been less reported. Chen et al. (2008) discovered that 89 the PAH sorption was significantly enhanced in paddy soil 90 amended with pine needles-derived biochar. The amendment of acidic paddy soil with sewage sludge biochar can reduce 91 the amounts of bioavailable arsenic, chromium, cobalt, nickel, 92 93 and Pb (Khan et al., 2013). Over five years, wheat strawderived biochar consistently decreased the bioavailable Cd 94 and Pb in a contaminated paddy field (Cui et al., 2016). 95 However, these studies all investigated either organic pollutants 96 or heavy metals. Whether biochar can simultaneously immobi-97 lize the co-contaminants of organic pollutants and heavy 98 metals in anaerobic (flooded) soil has rarely been reported. The 99 difference in the bioaccessibility of the co-contaminants in the 100 biochar-amended soils under aerobic versus anaerobic condi-101 tions also remains unclear. Selection of a reasonable cultivation 102method for remediated soil amended with biochar is consider-103 ably important to ensure safe agricultural production. 104

The activity and diversity of soil microorganisms that 105are important for the ecological functions of agricultural soils 106might be decreased by the combined pollution of heavy 107 metals and PAHs (Bourceret et al., 2016; Cao et al., 2008). 108 The physical properties of biochars (e.g., their pore structure, 109surface area and mineral content) support the functioning 110 of soil biota (Lehmann et al., 2011). Thus, the relationships 111 among co-contaminants, biochars and microorganisms should 112 113 be deeply understood. Biochar amendment might increase the microbial diversity of PAH-contaminated soil (Liu et al., 2015). 114 By contrast, Meynet et al. (2012) observed that the activity of 115PAH-degrading bacteria was more notable in unamended soil 116 117 than in the soil amended with activated carbon. However, reports comparing the effects of biochar on the soil microbial 118 119community structure under aerobic versus anaerobic conditions are also lacking. Studying the changes in microbial community 120 structure is helpful for selecting reasonable cultivation methods 121 for contaminated soils amended with biochar. 122

This study aimed to (1) compare the effects of biochar 123 on the bioaccessibility of the co-contaminants of organic 124 and inorganic pollutants in a field soil under aerobic versus 125 anaerobic conditions and (2) investigate the influence of 126 biochar amendments on the microbial community structure 127 under the two incubation conditions. Corn straw-derived 128 biochar (CB) and bamboo-derived biochar (BB) were amended 129 into an agricultural soil that was co-contaminated with 130 phenanthrene (Phe)/pyrene (Pyr)/Zn/Pb. The soils were incu- 131 bated under aerobic and anaerobic conditions to mimic upland 132 and paddy soils, respectively. A batch sorption experiment 133 was conducted to elucidate the Phe/Pyr/Zn/Pb sorption capac- 134 ities of the biochars. A phospholipid fatty acid (PLFA) assay was 135 conducted to analyze the soil microbial community structure. 136 To the best of our knowledge, this is the first study to compare 137 the effects of biochar amendments on the bioaccessibility of 138 organic and inorganic co-contaminants and on the microbial 139 community structure under aerobic versus anaerobic soils. 140

1. Materials and methods

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1.1. Soil and biochar preparation

An agricultural upland soil co-contaminated with Phe/Pyr/Zn/ 144 Pb for more than 40 years was sampled at a depth of 0–20 cm 145 in an area near a steel mill in the suburb of Nanjing, Jiangsu 146 Province. The soil was air-dried and sieved through a 2 mm 147 mesh. The soil had a pH of 7.36 and consisted of 27.60% 148 sand, 63.52% silt, and 8.88% clay. The dissolved organic carbon 149 (DOC) content of the soil was 309 mg/kg, and the total nitrogen 150

were 1.3, 0.57, and 17.4 g/kg, respectively.152Phe and Pyr were chosen as model compounds in this153study because they exist widely in PAH-contaminated soils,154and in relatively higher concentrations than other PAH155compounds (He et al., 2008). The Phe and Pyr concentrations156in the soil samples were (258.59 \pm 5.92) and (736.36 \pm 18.10)157µg/kg (dry weight), which exceeded the corresponding envi-158ronmental quality standards for soils in China (GB 15618-2008)159by one and two orders of magnitude. In addition, the Zn160and Pb concentrations in the soil were (116.93 \pm 14.36) and 161162(38.64 \pm 2.34) mg/kg (dry weight), respectively, which exceeded162the corresponding environmental quality standards for soils163in China (GB 15618-2008) by one order of magnitude.164

(N), phosphorus (P), and potassium (K) contents of the soil 151

CB was produced under oxygen-limiting conditions at a 165 final temperature of 300°C using a patented biochar reactor 166 (No. ZL2009 2 0232191.9) (Jia et al., 2013). BB, which is a 167 representative of commercial biochar and produced at 700°C, 168 was purchased from Shanghai Hainuo Charcoal Co., Ltd. The 169 surface area, pore volume, and pore size of these biochars were 170 measured via the Brunauer–Emmett–Teller (BET) method using 171 a V-Sorb 2800P analyzer (Gold APP Instruments Corporation, 172 Beijing, China). The elemental compositions were determined 173 using an elemental analyzer (ANA1500, Carlo Erba, Milano, 174 Italy) (Yuan et al., 2011). Table 1 details the physicochemical 175 properties of the biochars. 176

The concentrations of the \sum 16 PAHs in CB and BB were 177 778.72 and 413.17 µg/kg, respectively, which are both below 178 the recommended maximum limit set by the International 179

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