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Q5 Chemical characteristics of haze particles in Xi'an during 2 Chinese Spring Festival: Impact of fireworks burning

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A B S T R A C T

Fireworks burning releases massive fine particles and gaseous pollutants, significantly deteriorating air quality during Chinese Lunar New Year (LNY) period. To investigate the impact of the fireworks burning on the atmospheric aerosol chemistry, 1-hr time resolution of PM_{2.5} samples in Xi'an during the winter of 2016 including the LNY were collected and detected for inorganic ions, acidity and liquid water content (LWC) of the fine aerosols. PM_{2.5} during the LNY was 167 ± 87 μg/m³, two times higher than the China National Ambient Air Quality Standard (75 μg/m³). K⁺ (28 wt.% of the total ion mass) was the most abundant ion in the LNY period, followed by SO₄²⁻ (25 wt.%) and Cl⁻ (18 wt.%). In contrast, NO₃⁻ (34 wt.%) was the most abundant species in the haze periods (hourly PM_{2.5} > 75 μg/m³), followed by SO₄²⁻ (29.2 wt.%) and NH₄⁺ (16.3 wt.%), while SO₄²⁻ (35 wt.%) was the most abundant species in the clean periods (hourly PM_{2.5} < 75 μg/m³), followed by NO₃⁻ (23.1 wt.%) and NH₄⁺ (11 wt.%). Being different from the acidic nature in the non-LNY periods, aerosol in the LNY period presented an alkaline nature with a pH value of 7.8 ± 1.3. LWC during the LNY period showed a robust linear correlation with K₂SO₄ and KCl, suggesting that aerosol hygroscopicity was dominated by inorganic salts derived from fireworks burning. Analysis of correlations between the ratios of NO₃⁻/SO₄²⁻ and NH₄⁺/SO₄²⁻ indicated that heterogeneous reaction of HNO₃ with NH₃ was an important formation pathway of particulate nitrate and ammonium during the LNY period.

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52 Introduction

54 Air pollution caused by fine particulate matter (particles with an
55 aerodynamic diameter less than 2.5 μm, PM_{2.5}) has been a
56 persistent problem in China in the past decades due to fast

industrialization and urbanization without efficient emission 57
control. PM_{2.5} could penetrate deeply into the bronchi and lungs, 58
leading to respiratory and cardiovascular diseases (Bauer and 59
Koch, 2005; Chow et al., 2006). Recent studies have found that 60
the high level of PM_{2.5} in China has exerted a significant negative 61

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62 impact on human health (Sun et al., 2013; Ulbrich et al., 2009;
 63 Zhang et al., 2015). PM_{2.5} can be directly emitted from coal and
 64 biomass combustion and traffic exhaust (Feng et al., 2012; Lei
 65 et al., 2011; Wang et al., 2016b; Xu et al., 2016) and indirectly
 66 formed from photochemical oxidation of gaseous precursors
 67 and subsequent gas-to-particle phase partitioning. Fireworks
 68 burning, as a common human activity during festival events, is
 69 one of important sources of air pollution (Camilleri and Vella,
 70 2010; Drownick et al., 2006; Jiang et al., 2015; Jing et al., 2014).
 71 A number of studies have found that pyrotechnic displays on
 72 celebratory dates release massive SO₂, NO_x (Attri et al., 2001;
 73 Huang et al., 2012; Wang et al., 2007) and various fine particulate
 74 matters including organic and elemental carbon, ions (K⁺, Cl⁻
 75 and SO₄²⁻) and metals (e.g., Mg, K, Sr, Ba, Al, Cu and Pb) (Feng et
 76 al., 2016; Li et al., 2013; Moreno et al., 2007; Tsai et al., 2012).
 77 Those pollutants can significantly deteriorate air quality and
 78 raise health hazards especially for the respiratory-sensitive
 79 crowd (Becker et al., 2000; Gouder and Montefort, 2014). For
 80 example, it was reported that ten-minute exposure to SO₂
 81 at concentrations of 0.5 ppm could cause short-term asthma
 82 (Gong et al., 1995). In addition, studies (Godri et al., 2010;
 83 Hamad et al., 2016) also observed significant positive correlation
 84 between particulate oxidative burden and individual trace
 85 metals during fireworks burning periods.

86 Many observations based on filter samples have been
 87 conducted to investigate the impact of fireworks burning on
 88 air quality. However, such measurements were often per-
 89 formed with a time resolution from 12 to 24 hr, which is much
 90 longer than the duration time of fireworks burning activity
 91 (Cheng et al., 2014; Kong et al., 2015; Kumar et al., 2016), and thus
 92 could result in significant bias on understanding the physical-
 93 chemical properties of fine particles derived from fireworks
 94 burning due to the interferences of non-firework aerosols
 95 collected. Furthermore, those studies have mainly focused on
 96 chemical compositions of gaseous and particulate matters from
 97 fireworks burning (Huang et al., 2012; Lin et al., 2014; Moreno
 98 et al., 2007), and none of them has investigated the acidity and
 99 liquid water content (LWC) of aerosols derived from fireworks
 100 burning. Acidity is an important property of the aerosol, which
 101 could affect the formation of sulfate (Wang et al., 2016a), nitrate
 102 (Guo et al., 2016) and secondary organic aerosols (SOA) (Han
 103 et al., 2016; Jang et al., 2002). Aerosol LWC is also important for
 104 heterogeneous reactions and gas-to-particle phase partitioning
 105 of water-soluble compounds (Carlton and Turpin, 2013; Cheng
 106 et al., 2016; Wong et al., 2015). High LWC could significantly
 107 enhance secondary aerosol formation and increase aerosol
 108 light scattering efficiency (Chen et al., 2012; Malm and Day,
 109 2001). Therefore, characterization of physical and chemical
 110 properties of aerosols derived from fireworks burning including
 111 acidity and LWC with a high time resolution is necessary for
 112 a comprehensive understanding of their impact on the envi-
 113 ronment and human health.

114 In this study, atmospheric PM_{2.5} aerosols in Xi'an were
 115 collected during the Chinese Lunar New Year (LNY) of 2016 with
 116 a 1-hr time resolution and measured for inorganic ions, acidity
 117 and LWC. We first explored the difference in ion chemistry
 118 of the airborne fine particles in the city during the haze,
 119 clean and LNY periods, then investigated acidity and LWC of
 120 the aerosols in the three periods and finally discussed the
 121 formation mechanism of nitrate during the LNY episode.

1. Experimental section

122

1.1. Semi-continuous measurement of gas and aerosol phase pollutants

124
125

126 From 29 January to 16 February 2016, the field campaign was
 127 conducted at the site of the Institute of Earth Environment,
 128 Chinese Academy of Science (34.22°N, 108.88°E) (Fig. 1), in the
 129 northwest of Xi'an, inland China. During the sampling period
 130 the wind speeds were generally less than 2.5 m/sec with the
 131 lowest occurring in the LNY period (Fig. 2b). The concentrations
 132 of inorganic ions and associated gaseous precursors were
 133 measured with a 1-hr time resolution using in-situ Gas and
 134 Aerosol Compositions Monitor (IGAC). The IGAC configura-
 135 tion includes three units: (1) wet annular denuder (WAD) for
 136 capturing the precursor gas (HCl, HNO₂, HNO₃ and SO₂),
 137 (2) scrub and impact aerosol collector (SIC) to trap particle of
 138 fine fraction (PM_{2.5}), of which operating principle is similar
 139 to Steam Jet Aerosol Collector, (3) an ion chromatography (IC)
 140 (Dionex, ICS-1100) for the analysis of anions and cations
 141 (SO₄²⁻, NO₃⁻, Cl⁻, NO₂⁻, F⁻, NH₄⁺, Na⁺, K⁺, Mg²⁺ and Ca²⁺). Ambient
 142 aerosols from the inlet would flow through the WAD coated
 143 with absorption water solution containing H₂O₂ (4.2 nmol/L
 144 in water), which acts as the biocide and oxidant to promote the
 145 gaseous species to diffuse into the aqueous solution. After the
 146 WAD the particles enter the SIC, which creates a supersaturated
 147 environment for the growth of particles by deliquescence, and
 148 the grown particles are finally collected by inertial separation.
 149 Two pairs of syringe pumps separately transport the collected
 150 gas and particle samples to the IC system, which simulta-
 151 neously analyze the cations and anions of the sample. More
 152 details of a field measurement using IGAC were described
 153 elsewhere (Liu et al., 2017; Young et al., 2016).

154 To evaluate the measurement accuracy of IGAC, PM_{2.5}
 155 samples were also simultaneously collected using a high-
 156 volume filter sampler (1.13 m³/min) at the same site with a 3-hr
 157 time resolution. PM_{2.5} aerosols were collected onto pre-baked
 158 (450°C for 6 hr) quartz fiber filter, extracted with pure water and
 159 analyzed by the same ion chromatography (Wang et al., 2017b).
 160 As shown in Fig. S1, concentrations of major species such as SO₄²⁻,
 161 NO₃⁻ and NH₄⁺ measured by the IGAC were in a good agreement
 162 with those measured by the offline filter method ($y = 1.09x$ for
 163 SO₄²⁻, $y = 1.05x$ for NO₃⁻ and $y = 0.65x$ for NH₄⁺, respectively). Such a
 164 consistence was also reported by Liu et al. (2017), demonstrating
 165 the accuracy of the in-situ measurement using the IGAC monitor.

1.2. PM_{2.5} and meteorological data

166

167 The concentration of PM_{2.5} on the sampling days was
 168 downloaded from the website of Chinese air quality online
 169 monitoring analysis platform (<http://www.aqistudy.cn>). Meteorological
 170 data of Xi'an, including ambient temperature, relative
 171 humidity, wind speed and visibility, were obtained from the
 172 Shaanxi Meteorological Bureau and shown in Fig. 2a, b.

1.3. Acidity and liquid water content (LWC) of PM_{2.5}

173

1.3.1. Total hydronium ion

174

175 Total hydronium ions include in-situ H⁺, bisulfate ions (HSO₄⁻)
 176 and any other H⁺ in the solid phase of aerosol at nitrate

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